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Caption: Blue-spotted Mudskipper *Boleophthalmus boddarti* battle on land (surface behavior). © A. Kumaraguru



Genetic and reproductive characterization of distylous *Primula reinii* in the Hakone volcano, Japan: implications for conservation of the rare and endangered plant

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Abstract: Genetic and ecological evaluation are crucial in effective management of rare and endangered species, including those exhibiting complex breeding systems such as distyly. We studied a threatened distylous herb *Primula reinii* in the Hakone volcano, central Japan, to obtain baseline information of reproductive and genetic status towards conservation. In two representative populations inhabiting a central cone and somma of the volcano, population size, floral morph ratio, stigmatic pollen deposition, and fruit-set were measured. Using microsatellite markers, we evaluated genetic diversity, structure and differentiation of populations. Population bottlenecks and historical changes in population size were also estimated from genotype data. We found significant deviation from equal morph ratios in the central cone population, which also exhibited skewed mating success together with a high frequency of pollination within the same morph. These trends were not detected in the somma population. From genetic insights, the central cone population showed slightly lower genetic diversity, whereas no significant deviation from Hardy-Weinberg equilibrium was found in either population. The estimated moderate genetic differentiation and admixed genetic structure suggest recent lineage divergence and/or gene flow between populations. While robust evidence for a recent bottleneck was not obtained in our analyses, a clear signature of historical population contraction was detected in the central cone population. Our findings suggest that the skewed morph ratio strongly influenced the reproduction of small and isolated populations in the short-term, highlighting the vulnerability of distylous plant populations under ongoing anthropogenic pressure.

Keywords: Distyly, morph bias, reproduction, genetic diversity, volcanism

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Author contribution: KK conceived the project; MY, HS and KF collected the data; MY led the writing of the manuscript. HS and KK contributed critically to the drafts and give final approval for publication.

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INTRODUCTION

Worldwide, numerous plants are already threatened by human-caused stress (e.g., habitat destruction) and climate changes (Jackson & Kennedy 2009). Among these plants, species having a sophisticated entomophilous breeding system such as distyly (heterostyly) are likely to be the most vulnerable to the detrimental effects of isolation and unreliable pollination service due to anthropogenic environmental alteration (Washitani et al. 2005).

Distyly is a floral polymorphism, where populations have two floral morphs (a long- and short-styled morph; hereafter, referred to as the L- and S-morphs) that differ reciprocally in the heights of stigmas and anthers in flowers. Besides the morphological differences, distylous plants usually have a heteromorphic incompatibility system that prevents selfing and intra-morph mating (Barrett 2002); only cross-pollination (i.e., legitimate pollination) between L- and S-morphs results in seed setting. In theory, such morphologically and physiologically disassortative mating between floral morphs generally leads to an equilibrium with equal morph ratios, as a result of negative frequency-dependent selection and simple inheritance of distyly (Heuch 1979; Barret & Shore 2008). Accumulated evidence in distylous plants, however, has provided numerous examples of variation in population morph ratios (e.g., Kéry et al. 2003; Brys et al. 2008; Meeus et al. 2012). It has been advocated that floral morph bias can be governed by several factors, such as stochastic and deterministic events (Matsumura & Washitani 2000; Kery et al. 2003), maternal fitness differences between morphs (Hodgins & Barret 2006), and a combination of weak heteromorphic incompatibility and pollen limitation (Barret 1989; van Rossum et al. 2006; Brys et al. 2008).

Skewed morph ratios have often been found in small isolated distylous plant populations (e.g., Matsumura & Washitani 2000; Kery et al. 2003). Furthermore, it is well known that the deviation of morph frequencies from a 1:1 ratio can have negative reproductive and genetic consequences for populations. Indeed, skewed morph ratios result in the limited availability of compatible mates, which can contribute to reduced reproductive success (Kery et al. 2003; Wang et al. 2005; Pedersen et al. 2016) and increase the effects of genetic drift (Byers & Meagher 1992). Moreover, a combination of the loss of effective pollinators and the absence of a strict heteromorphic incompatibility system can increase inbreeding (selfing and biparental inbreeding) in morph-

biased populations (Barret 1989; Wilcock & Neiland 2002; van Rossum & Triest 2006). Another consequence of skewed morph ratios is an increase in genetic drift and inbreeding, which can lead to a loss of genetic diversity (van Rossum & Triest 2006; Meeus et al. 2012) and may eventually result in the breakdown of distyly (Washitani 1996). Therefore, conservation studies on threatened distylous species which integrate ecological and genetic approaches are indispensable for assessing current status and predicting future extinction probability, as well as for planning effective conservation and/or restoration strategies (Washitani et al. 2005).

The present paper investigated *Primula reinii* Franch. et Sav. (Primulaceae), an endemic primrose that inhabits mountainous regions of Japan. As in most primroses, the plant is a self-incompatible distylous species (Richards 2003). Because of its attractive and relatively large flowers with dwarf foliage, the primrose has suffered from anthropogenic activity (e.g. horticultural exploitation) in the wild. Based on the rarity and serious reductions in numbers and populations, the species is listed in the 'Vulnerable' category of the latest Japanese Red List (Ministry of the Environment 2019). Despite required effective management, especially for small populations exposed to ongoing human activities, little is known about their population status, reproductive success, or remnant genetic diversity. Simultaneously, this might provide an opportunity to study the immediate responses of a distylous plant population to demographic changes. In this study, we assessed the genetic and reproductive status of two neighborhood populations of *P. reinii* to provide baseline information pertinent to the conservation and preservation of this rare and endangered primrose. To discuss factors affecting reproductive success in natural habitats, we measured population size, morph frequency, stigmatic pollen deposition, and fruit-set within a population, and evaluated the genetic diversity and structure within and among populations.

MATERIALS AND METHODS

Plant species and study site

Primula reinii is a diploid ($2n = 24$) perennial herb occurring as a chasmophyte on wet shaded rocky cliffs in the mountains (Image 1). The natural populations are rare and usually small and isolated from each other because of their narrow edaphic niche and low dispersal ability arising from the nature of the species. A single ramet produces one to two pink flowers from

mid-April to early May. Although flower visitors are not well-known in this species, its narrow corolla tube and recent pollinator observations in their related species (Yamamoto et al. 2018) imply that long-tongued flying insects, such as bumblebees and bee flies can be effective pollinators for the species. Under cultivation conditions, the generation time (between seed germination and first flowering) of the species is estimated to be 2–3 years (Yamamoto et al. 2017).

Fieldwork was conducted from April 2013 to October 2014. We selected *P. reinii* populations from two sites in the Hakone volcano within a special protected zone of the Fuji-Hakone National Park in central Japan (Fig. 1a). These sites were severely isolated by a volcanic landform, i.e., caldera (distance: ca. 7 km) (Fig. 1b). One was on Mt. Kintoki (KIN population, 35.289°N, 139.004°E, 1,212m) and the other was on Mt. Komagatake (KOM population, 35.228°N, 139.021°E; 1,275m). In both sites, the primrose occurs on rock outcrops near the mountain top in which populations are composed of scattered patches within an area of approximately 20 × 20 m². Although a few smaller patches also located alongside the studied populations, these rim populations were sparse and separated geographically from the studied center population (>100 m). Thus, each studied population was considered to be one reproductive unit.

Formation and strong eruptive activities of the Hakone volcano initiated 650–350 ka and continued until 3ka (Nagai & Takahashi 2008). The mean annual precipitation and average air temperature at this area are 2,132mm and 8.8°C, respectively (<http://en.climate-data.org/location/769594/>, accessed 7 February 2019).



Image 1. Flowering ramet of *Primula reinii* on Mt. Komagatake. Photographed by Honami Sugawara in May 2013.

Currently, the Hakone volcano is an attractive destination to tourists. There is more than 10,000 people living within the caldera, and an approximate average of 60,000 tourists visit the area every day. Including *Primula reinii*, approximately 80% of endangered plants

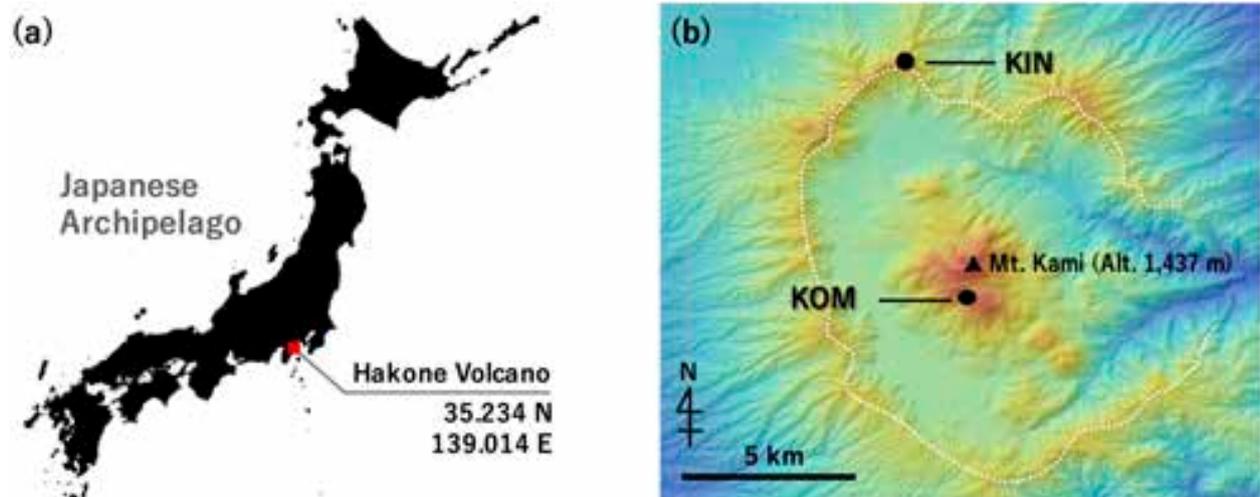


Figure 1. a—Location of Hakone volcano | b—Geographical map of Hakone volcano and location of the two populations of *Primula reinii* included in this study. KIN and KOM are population codes. The dashed line indicates the caldera rim of the Hakone volcano: Mt. Kami is the highest peak of the polygenetic volcanoes.

found in the volcano are presumed to have decreased due to habitat destruction and horticultural exploitation (Osawa & Inohara 2008).

Measurements of basic population traits, pollen deposition, and fruit-set

During the flowering season, the number of all flowering individuals and flowers within each population, along with flower morph (L- or S-morph), were recorded. Whether the two morphs were equally frequent within a population (i.e., deviations from a 1:1 morph ratio) was investigated with Chi-square goodness-of-fit tests.

To elucidate the role of pollen limitation and morph-ratio variation on female reproductive success, we measured the stigmatic pollen load. In the natural fields, 20 fully-opened flowers were collected from each population in the afternoon and carefully transported to the laboratory the same day. In the laboratory, flowers were dissected, and stigma removed and mounted on a microscope slide in aniline blue staining solution (0.1% aniline blue in 0.1 M K_3PO_4). Under a compound microscope (Olympus), we directly counted the number of legitimate (pollen from the opposite morph) and illegitimate (pollen from the same morph) pollen grains on stigmas of each floral morph based on pollen size differences (L-morph: $18.0 \pm 0.1 \mu\text{m}$; S-morph: $28.6 \pm 0.1 \mu\text{m}$; Y. Ojima, unpublished data).

At the beginning of the fruiting season (August), the population mean for fruit-set per flower was measured with the exception of some flowers that were used for the measurements of stigmatic pollen loads. Even if the fruit was set, the reproductive success will strongly depend on fruit predation (e.g., Matsumura & Washitani 2000; Yamamoto et al. 2013). Thus, we continued observations until October that was immediately before avulsion of the capsules.

All field surveys described above were conducted in 2013 and 2014 for KOM and KIN population, respectively.

Population sampling, DNA extraction, and microsatellite genotyping

In each population, 32 plants were sampled randomly (without distinction of floral morph types) from the entire area occupied by each population for genetic analysis. Leaf materials were collected and dried in silica gel. Before DNA extraction, leaves were homogenized with a disposable homogenizer (Biomasher 2; Nippi Co., Tokyo, Japan) to a fine powder. Total DNA was extracted from 40 to 80 mg silica-dried leaf tissue using the grass-fiber filter method (Takakura 2011). The extracted DNA was dissolved in a TE solution and stored at 4°C until use.

After the preliminary marker screening, the genotypes of each individual were characterized following seven microsatellite markers that were originally developed for *Primula sieboldii* E. Morren: ga0161, ga0218, ga0580, ga0691, ga1140, Pri0141, and 2ca135 (Ueno et al. 2003, 2006, 2009; Kitamoto et al. 2005). PCR amplifications and allele-size determination of fragment analysis were performed in accordance with the methods described by Yamamoto et al. (2017).

Population genetic analysis

For all seven microsatellite loci, the absence of linkage disequilibrium (LD) and the presence of null alleles were tested using Genepop v4.2 (Raymond & Rousset 1995). The LD test was verified using a Markov chain method with 1,000 dememorization steps, and 1,000 iterations per batch. Null allele frequencies were estimated by maximum-likelihood estimator based on the expectation-maximization algorithm (Dempster et al. 1977) with the default setting.

The following measures were calculated for each population: number of alleles (A), effective number of alleles (A_E), number of private alleles (A_p), expected heterozygosity (H_E), and inbreeding coefficient (F_{IS}). Deviations from the Hardy-Weinberg equilibrium were determined by the exact test and permutations. All measurements were calculated using GenoDive v2.0 (Meirmans & van Tienderen 2004). GenoDive was also used to compute the population's genetic differentiation pairwise F_{ST} and G'_{ST} indices (Hedrick 2005), and F_{ST} was tested for significance using 10,000 permutations.

To estimate genetic structure of *P. reinii* populations in Hakone volcano, we used the model-based clustering method STRUCTURE 2.3.4 (Pritchard et al. 2000) and non-model-based method principal component analysis (PCA). STRUCTURE analysis was conducted for all samples across the two populations. Under an admixture model with correlated allele frequency, 20 independent simulations were run for each K ($K = 1-5$) with 5×10^5 Markov chain Monte Carlo (MCMC) steps and a burn-in period of 10^5 interactions. The most likely value of K was determined by the ΔK method (Evanno et al. 2005) with STRUCTURE HARVESTER 0.6.94 (Earl & vonHoldt 2012). CLUMPAK (Kopelman et al. 2015) was used to average the outputs from multiple STRUCTURE runs and produce the graphical results. The F value, the amount of genetic drift between each cluster and a common ancestral population, was also calculated for each cluster. The PCA analysis was performed using the package *adegenet* 2.0.1 (Jombart 2008) in R 3.5.2 (R core Team 2018).

To detect a genetic imprint of past population

bottlenecks, we first used the heterozygosity excess method (Cornuet & Luikart 1996) implemented within the program BOTTLENECK v1.2 (Piry et al. 1999). This method is suitable to detect very recent and less severe bottlenecks, and has low false positive error rates (Williamson-Natesan 2005). All simulations were done with mutation-drift equilibrium conditions (2,000 replicates) under the stepwise mutation model (SMM), infinite allele model (IAM), and two-phase mutation model (TPM: 70% SMM and 30% IAM). A two-tailed Wilcoxon signed-rank test was used to determine a significant excess of heterozygosity.

We also calculated the *M*-ratio (Garza & Williamson 2001) for each population using Arlequin (Excoffier et al. 2005). The *M*-ratio test is considered to have a greater detection power for ancient and moderate-to-severe population declines in comparison with the former method (Williamson-Natesan 2005). *M*-ratio represents the number of alleles relative to the range in allele sizes. After a severe bottleneck, the number of alleles should reduce faster than the allelic size range, which results in a reduced *M*-ratio (Garza & Williamson 2001). Thus, the magnitude of the decrease reflects the severity and duration of the reduction in population size, and generally an *M*-ratio <0.68 is indicative of the presence of a bottleneck (Garza & Williamson 2001).

Finally, we conducted a Bayesian demographic analysis using the R package, Vareff (Nikolic & Chevalet 2014). In contrast to the first two moment-based methods, this coalescent-based approach can examine temporal changes in the effective population size (*N_e*). The function Vareff simulates prior changes in the effective population size from microsatellite data by resolving coalescent theory and using an approximate likelihood MCMC (Nikolic & Chevalet 2014). After a series of preliminary runs, we used the prior parameter settings for each population (Table S1), following recommendations from Nikolic & Chevalet (2014). We set the estimated mutation rate to 5×10^{-4} (Estoup et

al. 2002) for all loci, and ran each analysis under a two-phased mutation model with a proportion of 0.22 for multiple mutations (Peery et al. 2012), for 10^5 MCMC steps (NumberBatch = 1,000,000, LengthBatch = 10), sampling every 10 steps (SpaceBatch = 10) with an acceptance ratio of 0.25 (AccRate = 0.25), after burning of 10,000 steps. Estimations of sizes were searched for from sampling time to 5,000 and 500 generations ago.

RESULTS

Population traits and morph ratio

Each population trait is summarized in Table 1. A total of 72 flowering individuals and 99 flowers were found in the KIN population, whereas the KOM population had fewer (52 individuals and 69 flowers). The morph ratio in KIN did not deviate significantly from a 1:1 ratio (L-morph ratio = 0.54), even in the number of flowers (L-morph ratio = 0.52). In contrast, the number of flowering individuals of the L-morph was significantly higher than that of the S-morph in KOM (L-morph ratio = 0.65) and even higher for the number of flowers (L-morph ratio = 0.70).

Pollen deposition

Stigmas of both floral morphs received pollen grains in each population, but the numbers varied greatly between the individuals, ranging from zero to 321. In the KIN population, no differences in stigmatic pollen loads were detected between morphs (Fig. 2a). In addition, the proportion of deposited legitimate pollens was not significantly different between both morphs (Fig. 2c), while the proportions varied greatly among the L-morph stigmas in comparison with the S-morph. This is complemented by the result that the S-morph stigmas received significantly more legitimate pollen grains than the L-morph stigmas (Fig. 2b), implying that S-morph stigmas were pollinated more effectively than the opposite morph.

Table 1. Number of blooming plants during the flowering season in each population. Results of χ^2 goodness-of-fit tests for the similarity between the two morphs.

Number of	L-morph	S-morph	Total	χ^2	<i>P</i>
KIN pop.					
Flowering individuals	39	33	72	0.50	0.48
Flowers	51	48	99	0.09	0.76
KOM pop.					
Flowering individuals	34	18	52	4.92	0.03
Flowers	48	21	69	10.57	0.001

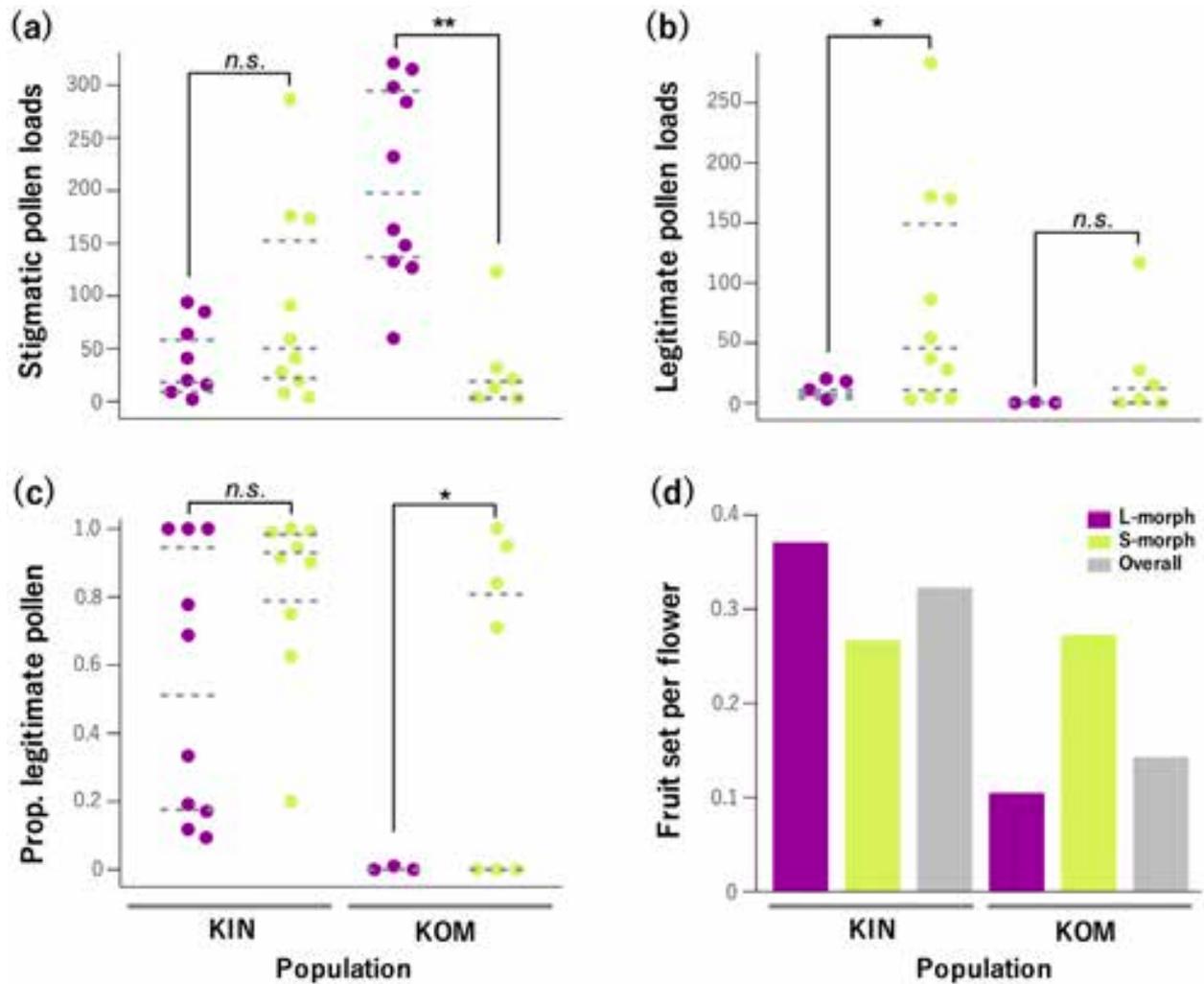


Figure 2. Pollen deposition on stigmas and the reproductive status of *Primula reinii* in the Hakone volcano. a—Stigmatic pollen loads | b—legitimate pollen loads | c—proportion of legitimate pollen within stigmatic pollen loads. a–c—are represented as beeswarm plots. Purple and green dots indicate L- and S-morphs, respectively. The dashed lines show the first quartile (lower line), median (center line), and third quartile (upper line). Asterisk denotes significant differences (Mann-Whitney U-test) between morphs: * $P < 0.05$, ** $P < 0.0001$. d—Mean population of fruit set per flower.

In contrast, in the KOM population, L-morph stigmas received a significantly greater number of pollen grains than the S-morph stigmas (Fig. 2a). After classifying pollen grains, however, we found no legitimate pollen grains loaded on the L-morph stigmas (Fig. 2b); that is, most L-morph stigmas were covered with a large quantity of illegitimate pollen. Although several S-morph stigmas were legitimately pollinated, similar to other populations (Fig. 2c), there was no significant difference in the number of legitimate deposited pollen grains between the two morphs (Fig. 2b).

Fruit-set

At the population level, fruit-set ratio was much higher in the KIN population than in the KOM population (32.3 %

and 14.3 %, respectively) (Fig. 2c). Within a population, both L- and S-morph scored comparable values in the KIN population (37.1 % and 26.7 %, respectively). In contrast, fruit-set of L-morph in the KOM population was less than half of the opposite morph (10.5 % and 27.3 %, respectively). We continued monitoring until October, but no evidence of fruit predation was found in either population, namely fruit-set was almost unchanged throughout the fruiting season.

Genetic diversity

LD between locus pairs was not significant. Although the frequencies of the majority of the null alleles were lower than 0.1, higher frequencies of null alleles were detected on 2ca135 and ga1140 loci in the KOM

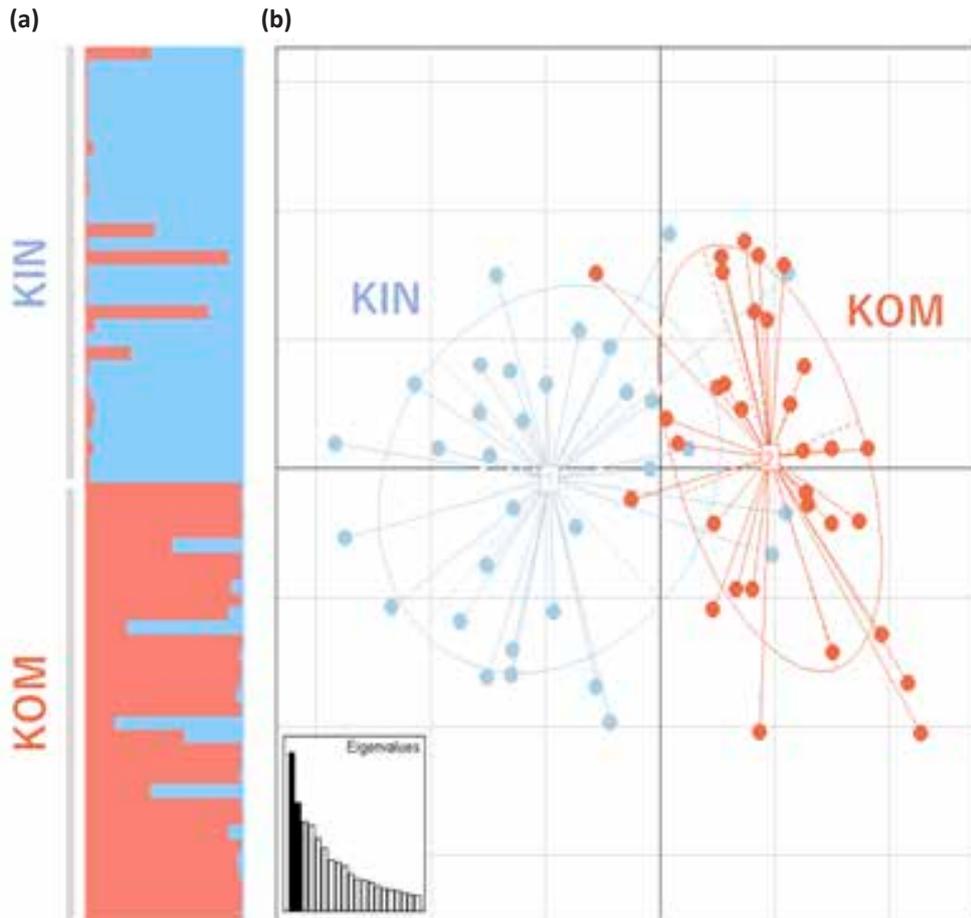


Figure 3. Genetic structure of 64 individuals of *Primula reinii* in the Hakone volcano. a—Assignment of individuals to clusters based on a STRUCTURE analysis for $K = 2$. b—Plots of individual scores on the principal component analysis (axis 1–2 are shown). Each plot is connected by a line to the 95% confidence ellipse centroid of the respective population. KIN and KOM are represented population codes.

Table 2. Genetic diversity and detection of a recent population bottleneck of the two *Primula reinii* populations.

Pop. code	Genetic diversity measurements					P values of Wilcoxon test			M-ratio	
	A	A_E	A_p	H_E	F_{IS}	IAM	SMM	TPM	Mean	SD
KIN	6.3	3.3	2.9	0.652	0.068	0.109	0.297	0.813	0.312	0.221
KOM	6.1	2.6	1.9	0.544	0.073	0.296	0.007	0.015	0.338	0.199
total	9.0	2.8	-	0.598	0.070	-	-	-	-	-

A, mean number of alleles; A_E , mean number of effective alleles; A_p , mean number of private alleles; H_E , expected heterozygosity; F_{IS} , coefficient of inbreeding; IAM, infinite allele model; SMM, stepwise mutation model; TPM, two-phase mutation model.

population (0.227 and 0.114, respectively). As the presence of null alleles may affect the estimation of genetic diversity or differentiation, we excluded the two loci and repeated several analyses to compare results between seven and five microsatellites. This trial revealed no clear difference in the results based on 5 vs. 7 loci (Table S2). Thus, seven loci were used in all

analyses described below.

Genetic diversity parameters for the two populations are presented in Table 2. In total, 63 alleles were amplified by seven microsatellite markers, with an average 9.0 alleles per locus. All diversity measurements were slightly higher in the KIN population ($A = 6.3$, $A_E = 3.3$, $A_p = 2.9$ and $H_E = 0.652$) than in the KOM population

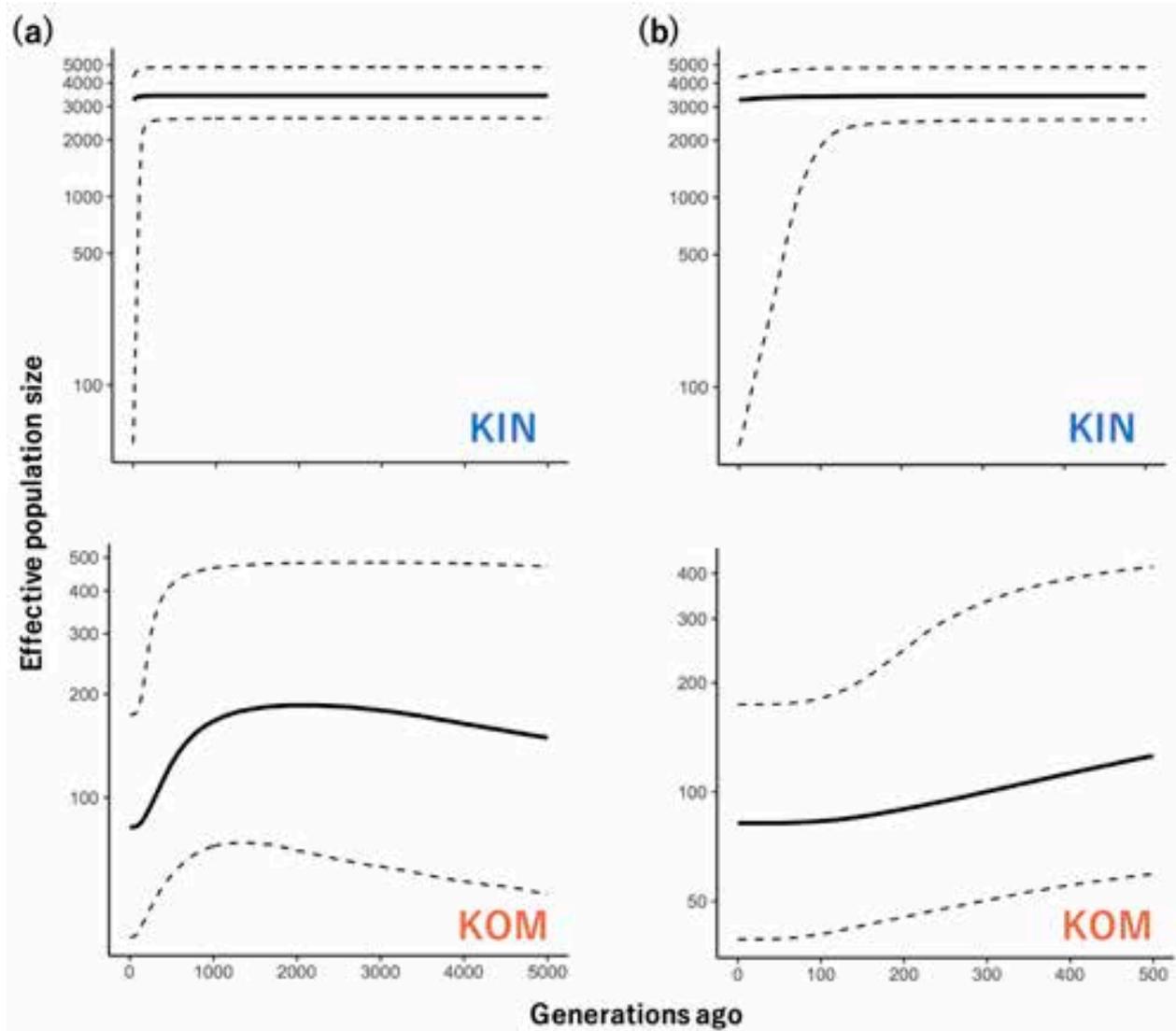


Figure 4. Estimates of effective population size of the two populations for a—the past 5,000 generations, and b—the past 500 generations. Bold lines indicate the median estimations, while the dashed lines above and below represent the 95% highest posterior density intervals.

($A = 6.1$, $A_E = 2.6$, $A_p = 1.9$ and $H_E = 0.544$). The inbreeding coefficient value (F_{IS}) was positive and comparable between the populations, but each value did not deviate significantly from zero.

Genetic structure and evidence a recent bottleneck

A moderate genetic differentiation was detected between populations ($F_{ST} = 0.115$, $P < 0.001$; $G'_{ST} = 0.286$). The STRUCTURE analysis based on the ΔK method indicated that ΔK was 462.5 for $K = 2$ and ΔK were < 3 for other values of K . Therefore, the optimal ΔK for $K = 2$ showed that the best-fit model for the 64 sampled individuals of *P. reinii* revealed two clusters (Fig. 3a). Although several admixed individuals were found in each population, all samples formed a clear genetic structure

between the two populations. The F values of clusters produced by STRUCTURE analysis were higher in the KOM population ($F = 0.186$) than in the KIN population ($F = 0.086$), indicating that the KOM population had undergone a larger genetic drift compared to that of the KIN population. In the PCA (Fig. 3b), the first two axes explained 17.0% and 10.7% of the variances in the experimental data, respectively. The results also distinguished the two populations, suggesting the existence of two genetic units corresponding to each population.

In BOTTLENECK analyses, the two-tailed Wilcoxon signed-rank test provided statistical support ($P < 0.025$) to the presence of a recent bottleneck in the KOM population under the SMM and TPM, whereas no

evidence was found in KIN (Table 2). On the contrary, the *M*-ratio test indicated that both populations experienced a reduction in population size. The *M*-ratio values were 0.312 and 0.338 for the KIN and KOM populations, respectively (Table 2). A clear signature of historical population contraction was detected only in KOM via the third method, the Bayesian population demographic analysis. The bottleneck began approximately 1,000 generations ago (Fig. 4a), whereas a gradual decline was settled at least 100 generations ago (Fig. 4b). In contrast, the KIN population seems to have historically had a large constant population size (Fig. 4a); however, recent changes were unclear due to the broad confidence levels (Fig. 4b).

DISCUSSION

Reproductive status and genetic diversity

The observed low reproduction in KOM is congruent with reports that morph-biased populations experience reduced reproduction (Byers & Meagher 1992; Kery et al. 2003; Wang et al. 2005; Pedersen et al. 2016). Given that almost all stigmas were covered with L-morph pollen grains (Fig. 2), it is plausible that frequent self- or intra-morph (i.e., illegitimate) pollination had occurred among the KOM L-morphs. Therefore, our ecological data indicate that the low fruiting success in KOM L-morphs was caused by stigmatic clogging (Yeo 1975) as a consequence of the skewed morph ratio. Because L-morph flowers generally produce greater amounts of pollen grains than S-morph flowers (Richards 2003), it is apparent that the total pollen pool within KOM was occupied by a large amount of L-morph pollen. Similar to our results, previous studies in distylous plants showed higher female reproductive success in the relatively less abundant morph than the dominant morph (e.g., Wyatt & Hellwig 1979; Thompson et al. 2003; Wang et al. 2005; García-Robledo & Mora 2007). Thus, these results may demonstrate negative frequency-dependent patterns of reproductive success in the distylous primrose.

The indices of genetic diversity were relatively high and comparable between the two populations (Table 2), despite the skewed morph ratio observed. In addition, each population exhibited low F_{IS} levels with no significant deviation from the Hardy–Weinberg equilibrium. These results allow for the conclusion that *Primula reinii* growing in the volcano had maintained sufficient genetic diversity as a result of outbreeding.

Overall, this study suggests the persistence of distylous self-incompatibility system in the *P. reinii*

populations. Nevertheless, determination of the exact causes of floral morph bias in KOM was not possible based on the limited ecological and genetic data currently available. Because skewed morph ratios are often explained by several biotic and abiotic factors as discussed in the Introduction, there is a need for future studies investigating the ability of selfing and intra-morph mating, maternal fitness differences between morphs, pollinator assemblage, and population demography.

Genetic differentiation and structure

Our molecular analysis showed that genetic differentiation was moderate between the two populations ($F_{ST} = 0.115$). Additionally, signs of genetic admixture between the populations were detected in PCA and STRUCTURE analyses (Fig. 3). There are at least two non-exclusive explanations for this: recent lineage divergence and gene flow. According to accumulated geographical surveys, Mt. Kintoki (locality of KIN pop.) and Mt. Komagatake (locality of KOM pop.) formed approximately ca 350–300 ka (Nagai & Takahashi 2008) and ca 27–20 ka (Kobayashi 1999; Nagai & Takahashi 2008), respectively. Formation of the central cone (i.e., Mt. Komagatake) clearly corresponded to the period of the last glacial maximum (LGM; ca 25–15 ka), suggesting that the KOM population was established at least after the last glacial period. The observed high *F* value (STRUCTURE analysis) and low private alleles in KOM may support a migration scenario that the population experienced a founder effect arising from a post-glacial refugial isolation and subsequent migration from the lowland of the caldera to the high-altitude areas of the central cone during the late Pleistocene and Holocene. Hence, it is plausible that the detected genetic admixture between populations suggests incomplete lineage sorting (i.e., sharing ancestral polymorphism between populations) due to recent lineage divergence.

Given the geographically close relationship between the populations (Fig. 1b), the presence of contemporary gene flow will also be taken into consideration. Because the two populations are severely isolated by a volcanic landform, gene flow mediated by pollen would be a plausible hypothesis. Moreover, in the flowering season we found claw marks, a useful indicator for the pollination services provided by bumblebees (Washitani et al. 1994), on the petals of each population. This may suggest that the bumblebees have a key role in pollination within the *Primula reinii* populations. Although bumblebees are known as strong-flying insects (e.g., Rao & Strange 2012), previous observations in other *Primula* species have demonstrated that pollen transfer by bumblebees

generally occurs within short distances (e.g., Ishihama et al. 2006). Therefore, we determined that the pollen flow between the populations might occur contemporarily but on very rare occasions. Nevertheless, deciding among the possible explanations for the genetic composition of the primrose in the Hakone volcano is difficult due to the weak evidence based on an insufficient number of loci.

Recent and historical demography

The two tests for a recent bottleneck yielded mixed results (Table 2). Based on the BOTTLENECK analysis, only the KOM population exhibited excess heterozygosity. In contrast, the *M*-ratio test supported a recent population size reduction in both populations. As mentioned above, however, because these inconsistent results might be attributed to the low statistical power of our sample size (e.g., number of loci or individuals), our results should be interpreted with caution. Nevertheless, such conflicting results often indicate the severity or timing of the reduction in population size (Williamson-Natesan 2005; Marshall et al. 2009; Padilla et al. 2015; Tóth et al. 2019), and were expected due to the differences in power detecting a bottleneck (Peery et al. 2012).

Considering the robust results in KOM, it is likely that the morph-biased population may have undergone more recent and severe bottlenecks in comparison with another population. In theory, the BOTTLENECK analysis can demonstrate population bottlenecks over a period of 0.2–4.0 N_e generations (Cornuet & Luikart 1996). Assuming for KOM population of $N_e = 100$ (Fig. 4) and a generation time of 2–3 years, it translates into approximately 50–1000 years before the present. On the other hand, a clear sign of recent (within 100 generations) population bottleneck was not found in the Bayesian demographic analysis (Fig. 4). Therefore, based on results from a series of demographic analyses, it is difficult to draw a definitive conclusion on whether recent bottlenecks occur or not, and thus, we defer a final conclusion until more genetic data are available in the future.

Contrary to this, the Bayesian demographic analysis provided strong evidence in support of a historical population bottleneck in KOM inhabiting the central volcanic cone. The first signs of population decline would have occurred 2–3 ka (assuming a generation time of 2–3 years). This timeframe post-dates a climatic warming, known as the Jomon optimum transgression, that occurred approximately 6ka, implying that historical population bottlenecks were likely due to volcanic activities as opposed to climatic events. According to geological records, the last major eruption of the volcano

was from the central cone in 2.7–2.9 ka (Kobayashi 1997; Kobayashi et al. 2006), and intermittent phreatic eruptions continued until present-day. Although speculative, these evidences may support the idea that the historical population declines experienced by the KOM could have been associated with repeated eruptive activities in the central cone. Perhaps, the detected recent bottlenecks in KOM are caused by eruptive activities rather than human activities.

On the other hand, the estimated effective population size in the KIN population inhabiting the somma mountains was large and constant in the long term, suggesting that the population has been maintained without suffering from volcanic eruptions occurring in the central cone. Further studies for the lineage divergence and demographic history of *P. reinii* in this region, using more informative datasets (e.g., single nucleotide polymorphisms), will be valuable because volcanism is one of the key abiotic factors in the plant's diversification and distribution in Japan (e.g., Yoichi et al. 2017; Nagasawa et al. 2020), located in the Pacific Ring of Fire.

Implication for conservation

Our study suggests that morph imbalances are striking effects on the reproduction of *P. reinii* population in the short-term. Accordingly, a measure of morph ratio should be given top priority in conservation management of the species, and enhancement of habitat monitoring should be considered as in situ managements to protect remnant individuals and to maintain optimum morph frequencies from horticultural exploitation. Considering the observed negative frequency-dependent patterns of reproductive success, if heteromorphic self-incompatibility is totally strict in *P. reinii*, the skewed morph ratio in KOM may be improved in the future when regeneration is successful. However, the exact breeding system of the species remains poorly understood. Therefore, in addition to other examinations (e.g., the germination requirements and the effect of storage time of seeds) towards a future ex situ conservation strategy, the levels of within morph fertility and selfing ability should be resolved immediately to evaluate the medium- to long-term risk of extinction in the remnant populations across species distribution ranges.

The two surveyed populations in the Hakone volcano were distinguished by two genetic clusters, suggesting that each population should be divided into a different management unit to maintain evolutionary distinctiveness and ecological viability (Moritz 1994; Frankham et al. 2002). The moderate genetic



differentiation and the presence of large amounts of private alleles between the populations highlight this suggestion; thus, artificial inter-population crossing should be avoided in this case. Nevertheless, the lack of samples from other parts of the volcano will influence the estimated genetic structure. Thus, an exhaustive population sampling, including other remnant small population, is required to elucidate the genetic structure and demographic history of *P. reinii* occurring in the Hakone volcano as is also needed for planning conservation strategies.

To our knowledge, this is the first conservation genetics study on threatened plants in the Hakone volcano, which harbors approximately 1,800 plant species (Tanaka 2008). Thus, the results discussed here will be useful for designing both in situ and ex situ conservation strategies for *P. reinii* as well as other plants inhabiting the volcano and shed light on the instability of plant populations due to the impacts of volcanism and human activities. Our study highlights the importance of studies in conservation, integrating ecological and genetic approaches to accurately assess the population status of endangered species and draw up effective conservation strategies.

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Table S1. Prior parameter settings for each population using VarEff software.

Parameters	KIN	KOM	Description
JMAX	4	4	Number of when the effective size has changed
DMAX	10	7	the maximal distance between alleles
NBAR	100	100	prior value for the effective population size
RHOCORN	0	0	coefficient of correlation between effective population size in successive intervals
VARP1	3	3	variance of prior log-distribution of effective population size
VARP2	3	3	variance of prior log-distribution of time intervals
GBAR	10000	5000	number of generations since the assumed origin of the population
Diagonale	0.5	0.5	a smoothing parameter

Table S2. Genetic diversity measurements and population differentiation between the two populations based on the selected five loci (*ga0161*, *ga0218*, *ga0580*, *ga0691* and *Pri0141*).

Pop. code	A	A_E	H_E	F_{IS}	F_{ST}	G'_{ST}
KIN	6.0	3.3	0.637	0.020	0.111	0.249
KOM	5.0	2.1	0.472	0.029		





A review about fish walking on land

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Abstract: Mudskippers are amphibious species inhabiting semi-terrestrial ecosystems like mudflats, mangroves, marshy swamps, intertidal regions, and estuaries. Around 34 diversified species are found across the globe. Mudskipper belongs to the Oxudercidae family and the subfamily is Oxudercinae. The occurrence of species is vastly found across the Indo-West Pacific region, the tropical western coast of Africa and in the Indian Ocean. Mudskippers are known for being the biological indicator and also an indicator of estuarine safety monitoring. They are used by people for prey-catching baits. This review paper explains the ecological indicators, taxonomy, species diversity, habitat, behavioural pattern, respiration & kinematics, feeding ecology, reproduction, nutrition content & its medicinal value, and threats to mudskippers.

Keywords: Amphibious fish, distribution, ecological indicator, mudskippers, species diversity.

பரவெட்டிமீன்(மட்ஸ்கிப்பர்கள்) (Mudskipper) நிலநீர் வாழிகள் வகுப்பைச்சார்ந்த, சேற்று திட்டிகள், சதுப்புநிலங்கள், அலைஏற்ற பகுதிகள், ஆற்று முகத்துவாரம் மற்றும் கழிநிலங்கள் போன்ற பகுதிகளை வாழ்விடமாக கொண்டவை. ஏறக்குறைய 34 பன்முகப்பட்ட பரவெட்டிமீன் இனங்கள் உலகில் காணப்படுகின்றன. பரவெட்டிமீன்/மட்ஸ்கிப்பர்கள் ஆக்ஸுடெர்சிடே கோபிடே (Oxudercidae Gobiidae) குடும்பத்தை சேர்ந்தவை. இந்தோ-மேற்கு பசிபிக் பகுதி, ஆப்பிரிக்காவின் வெப்பமண்டல மேற்கு கடற்கரை மற்றும் இந்தியப் பெருங்கடலில் இந்த உயிரினங்களின் நிகழ்வு பரவலாகக் காணப்படுகிறது. பரவெட்டிமீன் உயிரியல் குறிகாட்டியாகவும், ஆற்று முகத்துவாரங்களின் பாதுகாப்பு கண்காணிப்பின் குறிகாட்டியாகவும் அறியப்படுகின்றன. மனிதர்களின் பயன்பாட்டிற்கு பெரிதும் உபயோகபடாததால், இவை பெரும்பாலும் இரையை பிடிக்கும் தூண்டலில் பயன்படுத்தப்படுகின்றன. இந்த ஆய்வுக் கட்டுரை, சுற்றுச்சூழல் குறிகாட்டிகள், வகைபிரித்தல், இனங்களின் பன்முகத்தன்மை, வாழ்விடம், நடத்தை முறை, சுவாசம் மற்றும் இயக்கவியல், உணவு சூழலியல், இனப்பெருக்கம், ஊட்டச்சத்து தரவுகள், அதன் மருத்துவ மதிப்பு மற்றும் அபாயங்கள் ஆகியவற்றை விளக்குகிறது.

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INTRODUCTION

Osche (1962) suggested that mudskippers would be useful for interpreting the transition of species from water to land. There is some evidence about the transition of fishes to tetrapods, which occurred more than 360 million years ago (Ulrich & Elliott 2013). There are about 40 species classified under 10 genera. Mudskippers predominantly inhabit the mangrove forest and mudflats of the Indo-West Pacific region, the tropical western coast of Africa, and Indian Ocean coastlines. Mudskippers are known to be burrow dwellers preferring the swampy marshes, estuaries, and the intertidal regions for their living (Murdy 1989). The feeding ecology of the Mudskipper *Periophthalmus barbarous* was found to be 'opportunistic feeders' meant to be satisfying the food needs with the available resources (Chukwu & Deekae 2013). Generally, mudskippers feed on algae, detritus, diatoms, nematode, polychaetes and eggs of fishes, along with mud and sand particles (Ravi 2013).

Since long ago, researchers have explored mudskipper's distribution pattern, species diversity, behavioural patterns and their locomotory activities. In 1989, Murdy studied the morphological features of mudskippers, which paved the way for taxonomic classification of mudskippers. Also, he hypothesized the relationship of *Oxudercinae* gobies with other genera and groups. Mudskipper's early developmental age & growth, respiratory & circulatory adaptations, and feeding ecology were also studied by many researchers across the globe (Jaafar & Murdy 2017).

Due to the medicinal properties found in mudskippers, it gains commercial value either through food consumption or as traditional medicine by humans (Kanejiya et al. 2017a). Sometimes they are even used as bait for catching prey (Gadhavi et al. 2017). A few of the vernacular names used in different places for referring to mudskippers are *Periophthalmus darwini*, *Periophthalmus* sp., and *Periophthalmus novaeguineensis* as sakomo, *Periophthalmus* as nebesokera, *Periophthalmus weberi* as paraguamo, *Periophthalmodon freycineti* as genora, *Boleophthalmus caeruleomaculatus* as ebanea, *Boleophthalmus* sp. as poti, *Scarpellos histophorus* as seekakowea, *Oxuderces wirzi* as canipo (Polgar & Lim 2011). Mudskippers are known as 'vaetti uluvai' in Tamil (Ragunathan 2010).

Ecological Indicator

Monitoring the intertidal and estuarine health depending on their abundance mudskippers have direct

affection towards salinity and temperature variations (Kanejiya et al. 2017c). They also contribute towards the growth of mangrove trees, when the population of mudskippers is abundant, consequently providing appropriate nutrients supporting the growth of mangroves (Shenoy et al. 2012).

Polgar (2009) attempted the studies of Malayan mudskippers as a biomonitor in mangroves concerning species area relationship (SAR) and revealed that the destruction of habitats resulted from reduced species abundance. Mudskippers bioaccumulate pollutants through ingestion as they tend to be prey for many mudflat species. Mudskippers ensure the coastal, intertidal, and estuary's health (Polgar & Lim 2011). Studies have shown that mudskippers act as a biological indicator of addressing the severity of oil pollution in the coastal areas of the Persian Gulf. Biomarker responses of *Periophthalmus waltoni* were recorded and discovered that it acts as the ideal candidate for bioindicator in the coastal regions and mudflats. Correspondingly, it can be utilized for achieving sustainable development (Shirani et al. 2012).

Similarly, at Kuwait Bay, *Periophthalmus waltoni* acts as a bioindicator in identifying the bioaccumulation of metals (Bu-Olayan & Thomas 2008). It tends to be a feeble indicator at the Persian Gulf marine ecosystem, a biomonitor candidate-*Boleophthalmus dussumieri* used in the monitoring of polycyclic aromatic compounds (PAH). It might be various other environmental factors which are acting in that ecosystem leading *B. dussumieri* to be helpless (Sinaei & Mashinchian 2014). Metal toxicity and accumulation of metal toxicants like Zn, Cd, Pb, Cu have suggested that mudskippers are an ideal biomonitor. It was observed that their fins and liver are the biomonitors and accumulate metal toxins. Due to the contaminants of various metal pollutant, it is not recommended for human consumptions (Ikram et al. 2010).

Taxonomy

Mudskippers belong to the family Oxudercidae and the subfamily *Oxudercinae*. Some of the species are tabulated in Table 1 and Image 1.

Species Diversity

Globally, 34 diverse species of mudskippers were recognized, and among them, nine were found in Indian coastlines (Murdy et al. 1989). Nearly five diversified mudskipper species have been effectively documented from Gujarat coastlines (Devendra et al. 2016). In the southeastern coast of India, mudskippers were recorded

Table 1. Taxonomic rank, genus, and species of mudskippers around the world with reference to various authors belonging to different periods.

	Tribe (taxonomic rank)	Genus	Species	References
1	<i>Oxudercini</i>	Parapocryptes	<i>Parapocryptes rictuosus</i> , <i>Parapocryptes serperaster</i>	Valenciennes 1846
2	<i>Oxudercini</i>	Apocryptodon	<i>Apocryptodon madurensis</i> , <i>Apocryptodon punctatus</i>	Tomiyama 1934
4	<i>Oxudercini</i>	Oxuderces	<i>Oxuderces dentatus</i> , <i>Oxuderces wirzi</i>	Koumans 1938
5	<i>Periophthalmini</i>	Apocryptes	<i>Apocryptes bato</i>	Hamilton 1822
6	<i>Periophthalmini</i>	Pseudapocryptes	<i>Pseudapocryptes borneensis</i> , <i>Pseudapocryptes lanceolatus</i>	Bloch & Schneider 1801
7	<i>Periophthalmini</i>	Zappa	<i>Zappa confluentus</i>	Roberts 1978
8	<i>Periophthalmini</i>	Scartelaos	<i>Scartelaos cantoris</i> , <i>Scartelaos gigas</i> , <i>Scartelaos histophorus</i> , <i>Scartelaos tenuis</i>	Chu & Wu 1963
9	<i>Periophthalmini</i>	Boleophthalmus	<i>Boleophthalmus hirdsongi</i> , <i>Boleophthalmus boddarti</i> , <i>Boleophthalmus caeruleomaculatus</i> , <i>Boleophthalmus dussumieri</i> , <i>Boleophthalmus pectinirostris</i> ,	Pallas 1772
10	<i>Periophthalmini</i>	Periophthalmodon	<i>Periophthalmodon freycineti</i> , <i>Periophthalmodon schlosseri</i> , <i>Periophthalmodon septemradiatus</i> ,	Valenciennes 1846; Pallas 1772 Hamilton 1822
11	<i>Periophthalmini</i>	Periophthalmus	<i>Periophthalmus argenteolineatus</i> , <i>Periophthalmus barbarus</i> , <i>Periophthalmus chrysospilos</i> , <i>Periophthalmus gracilis</i> , <i>Periophthalmus kalolo</i> , <i>Periophthalmus malaccensis</i> , <i>Periophthalmus minutus</i> , <i>Periophthalmus modestus</i> , <i>Periophthalmus novaeguineensis</i> , <i>Periophthalmus novemradiatus</i> , <i>Periophthalmus waltoni</i> , <i>Periophthalmus weberi</i>	Valenciennes 1846 Hamilton 1822

most during the post-monsoon season, and their preferred habitats are estuarine lands and mangrove areas. A group of eight species from the Malay Peninsula and 12 species from Sumatra is present along the Straits of Malacca (Takita & Ali 1999). Earlier, there were nine species of mudskippers recorded in peninsular Malaysia (Polgar 2009). Recent studies have updated the total count to 17 in peninsular Malaysia (Khaironizam & Rashid 2005). Nineteen species have been recorded recently in the Ramsar site, Johor, Malaysia (Hui et al. 2019). Some of the mudskipper species found in Merauke District, Indonesia are *Boleophthalmus boddarti*, *B. pectinirostris*, *P. takita*, *P. argenteolineatus*, *Scartelaos histophorus*, and *Oxuderces dentatus* (Elviana et al. 2019). *Periophthalmus waltoni* is maximally distributed along the Persian Gulf though there are many threats to its population density (Sharifian et al. 2018). Though there is a diverse population of *Periophthalmus barbarus* around southeastern Nigeria, it is affected by overexploitation (Abiaobo & Udo 2017). People do not consume *Periophthalmus novemradiatus* as a result of which their growth rate is rapidly increasing along the

Bakkhali River Estuary, Bangladesh (Rahman et al. 2015). Complete phylogeographic studies of *Periophthalmus* distributed along Indo-Pacific region helped to understand its evolutionary history (Polgar et al. 2014). *Periophthalmus spilotus*, a new species of mudskipper, was identified from Sumatra, Indonesia (Murdy & Takita 1999). Similarly, a new species *Parapocryptes serperaster* has been recorded in peninsular Malaysia (Khaironizam & Rashid 2000). *Periophthalmus walailakae* has been recorded in southeastern India (Mahadevan et al. 2019a). There were about 24 newly recorded species during the recent studies in Indonesian waters (Pormansyah et al. 2019).

Living Habitat

Mudskippers inhabit riparian areas with soft and muddy plains. Also, they inhabit where the salinity level is found to be low and the place rich with benthic invertebrates (Baek et al. 2008). Numerous species were living on the rocky coastline as much, as they occupy mudflats, mangroves, and sand flats (Gordon et al. 1968). Rehabilitating the coastal region with mangrove



Image 1. Blue-spotted Mudskipper *Boleophthalmus boddarti*. © A. Kumaraguru

saplings gives a better habitat for various species, and it was observed that large-sized mudskippers were found in the breakwater (sheltered area for mangroves) (Hashim et al. 2010). *Pseudapocryptes elongatus* is able to tolerate the salinity, and hence it survives in the open sea, coastal mangroves and inland habitat during different stages of their growth (Bucholtz et al. 2009). Mudskippers alter the environmental conditions improving the growth of young mangroves as they mix the soil with detritus (Ravi et al. 2013). Studies related to the microhabitat selection of Chinese mudskippers identified their preferences, such as salinity levels, land or water; water and air temperatures; light or dark; and various combinations among them were conveyed (Gordon et al. 1985). The major threats to the mudflats are soil erosion, macro algae, terrestrialization, and lack of estuarine water, human interference, and discharge of effluents (Ravi 2012). *Periophthalmodon septemradiatus* species is found to be the first species inhabiting and breeding from a saline environment to a

completely fresh water region of the Mekong River (Mai et al. 2019). *Boleophthalmus pectinirostris* has shown behavioural preferences in choosing their microhabitat at their early juvenile stage (Chen et al. 2008). To determine the habitat selection, genomic studies were performed, and also the comparative analysis among different species were conducted (Cai 1996). Different mudskippers inhabit different microhabitats and have different burrow construction methods (Clayton 1993). Habitat selection is dependent on its ecological interactions (Polgar & Crosa 2009). Some species such as *Periophthalmodon septemradiatus* is found to survive in habitats with low salinity and far from the sea, whereas *B. boddarti* survives close to the sea with high salinity (Khaironizam & Rashid 2003). Mudskippers preferred thick mudflat areas for carrying out their burrowing activity effortlessly (Kanejiya et al. 2017c). Burrows constructed by *Boleophthalmus boddarti* are classified as follows: burrows with single apertures represent newly constructed one, and the other single and double

openings ones currently exist. In contrast, the burrows with multiple apertures represent collapsed ones due to human interventions (Ravi et al. 2004).

Behavioral pattern of mudskipper

Population density is inversely proportional to the growth rate as the availability of food decreases due to an increased population. *Boleophthalmus pectinirostris* (Blue-spotted Mudskipper) is found to have the longest lifespan; seven and six years in males and females, respectively (Nanami & Takegaki 2005). A study on the terrestrial life of mudskippers with *Periophthalmus sobrinus* showed their survival capacity out of water is one and a half days. Surprisingly, there was no affection in metabolic and heart rates as well as the lactic acid concentration in blood during their living out of the water (Gordon et al. 1969). Whereas Chinese Mudskipper *Periophthalmus cantonensis* survives for two and a half days out of the water and observed medium sensitiveness of metabolism towards temperature. Starvation for 9.5 days did not affect the excretion of ammonia but affected excretion of urea (Gordon et al. 1978).

Boleophthalmus dussumieri has separate exit and entry for males and females. Juveniles pierce deep into mud during high tides. Adults neglect to build chimneys surrounding their holes as the consistency of the soil is between sand and clay (Rathod et al. 2019). Mudskippers growth rate is affected when being exposed to pollution at the embryonic stage (Kruitwagen et al. 2006). There is a significant role played by aquaporins (integral membrane proteins) for adapting themselves to the terrestrial lifestyle. The selective changes, like pore formation and substrate selection, have a substantial contribution to their adaptation to an amphibious lifestyle (Lorente-Martinez et al. 2018). The species *Periophthalmodon schlosseri* is found to be an ideal species for aquaculture and more mudskipper studies (Quang 2016). Mudskippers like *Boleophthalmus boddarti* constructs mud-walls, for territorial exclusion or spacing, territory, and reduces hostility. This is based on their abundance. These regional behavioural patterns give better knowledge about the elastic disc concept of territories (Clayton 1987). Mudflats are vital for the survival of mudskippers but global warming is a serious threat as the mudflats are greatly affected due to high temperature. Spawning season in *Scartelaos gigas* is from May to July. The growth rate is dependent on diet, water temperature, and mudflat exposure (Park et al. 2002).

Some mudskipper species tend to be more

comfortable with a terrestrial lifestyle rather than an aquatic lifestyle. One among them is *Periophthalmodon schlosseri*, which has gill arrangements which are highly adaptable for air-breathing, and they spend less time in marine habitats (Takeda et al. 1999). *Boleophthalmus boddarti* builds mud walls for two significant reasons: to avoid hostility between neighbours and as assistance for feeding. Diatoms are the most preferred food for mudskippers. They prefer feeding on mud slopes as a measure of preventing intervention by their neighbours (Clayton & Wright 1989). *Periophthalmus sobrinus* prefers to live unaccompanied and rarely lives within closed groups. There were large spacings between nests and dark places were preferred for foraging (Gordon et al. 1968). The growth rates and life duration of both the sexes of *Pseudapocryptes elongatus* obtained from Sundarbans, India is four-plus years for both the sexes as the maximum age. Their growth index (Φ) is 4.394 (males) and 4.503 (females) possessing larger caudal fins (Mahadevan et al. 2019c).

The reason for aggressiveness in *Periophthalmus modestus* is the hypothalamic hormone, arginine-vasotocin (VT) (Nao et al. 2013). The foraging behaviours in *Periophthalmus waltoni* are not influenced by environmental factors. They hunt in the same area as their prey stays inside the burrow for a longer period of time (Clayton & Snowden 2000). The growth of *Boleophthalmus boddarti* has been recorded high during their juvenile period, decreasing in successive years because of maturation and spending their energy in spawning (Ravi & Rajagopal 2007).

Respiration and kinematics

Anatomical characteristics of mudskipper gills decide on adaptations of their habitat and *B. boddarti* is one that shows excellent adaptation to aquatic lifestyle. In comparison, terrestrial adaptations were favoured more at odds with their amphibious lifestyle (Low et al. 1988). Since mudskippers appear to change the way their skin breathes, their epidermis and skin layers have been studied in detail (Beon et al. 2012). They adapt towards terrestrial lifestyle by secreting mucus, and their head containing dense capillary network assists cutaneous respiration (Jie et al. 2003). Studies related to gaseous exchange and their demand for oxygen were done in the intertidal regions (Karen 1993). Ammonia excretion happens in mudskippers (*Periophthalmodon schlosseri*) through their head. Ammonia gets collected in their burrow through the acidification process, preventing them from reverse fluctuation (Randall et al. 2004). Mudskippers maintain the air phases according

to the tide. They can breathe both aquatically and aerially when there is a high tide, while at low tides they transfer air into their burrow for breathing (Lee et al. 2005). Comprehensive research was conducted in *Periophthalmus magnuspinnatus* on cutaneous respiration and its relationship with skin layers (Park 2002). Several mudskippers reported evaporative water loss along with their behavioural adaptations (Dabruzzi et al. 2011). Significant characteristics of gills found in *Periophthalmodon schlosseri*, contribute to their ability to live inland for a longer period of time (Wilson et al. 1999). The modifications and transformations that occur in the gill respiratory vasculatures and the mudskippers of the bucco-opercular cavities were examined using the technique of corrosion casting. This has helped to define their adaptation to an amphibious lifestyle (Gonzales et al. 2011). Histological studies using the paraffin method have studied the ambiguity in the structure of the gills and simultaneously compared their aquatic and terrestrial lifestyles (Supriyati et al. 2019). For their survival, condition-specific biochemical adaptations occurred during hypoxia, where the glycolysis process is modified to provide energy during muscle movements, and lactate is accumulated (Chew & Ip 1992).

Contradictory circumstances occur concerning their adaptations between *Periophthalmodon schlosseri* and *Boleophthalmus boddarti*. Around the same time, *P. schlosseri* tends to be adaptable to a terrestrial lifestyle and is an excellent candidate for aquaculture studies related to air-breathing fish. *B. boddarti*, however, appear as opposed to the earlier one (Kok et al. 1998). Mudskipper is an anomaly that stores pre-entry air inside their burrow and has no metabolism affection. They regularly maintain their air stages and adapt them to their amphibious lifestyle (Ishimatsu et al. 1998). The essential feature of their adaptation to terrestrial life is the presence of dermal bulges, thick middle cell layer and a vascularized epidermis (Zhang et al. 2000). In low tide conditions, the intertidal fishes, like mudskippers, tend to have many options to choose from. Mudskippers either agree to be an aquatic or temporary terrestrial living being (Karen 1995). Mudskipper's locomotion has been experimentally tested using water on gelatin and glass along with the assistance of a system for digital image processing. Wang et al. (2013) discovered through their study that mudskipper uses both body and pectoral fins for movement in the water and on gelatin, whereas they use only pectoral fins for land action. Escapism is a critical behaviour that is required for their survival. A few studies have shown that, during aquatic and terrestrial lifestyles, mudskippers alter their position

as escapists (Swanson & Gibb 2004).

Feeding ecology

The mudskipper *Pseudapocryptes dentatus* follows herbivorous feeding patterns. Their main order is Diatoms-Bacillariophyceae, green algae and blue-green algae. Levels in size and metabolism are inversely proportional (Sarker et al. 1980). Both plants and animals were documented while studying the *Periophthalmus barbarous* stomach. This included mainly crabs, fish scales, and insects. We can recognize from this that mud-skippers are opportunistic feeders and are an ideal aquaculture choice (Chukwu & Deekae 2013). As described, their key food items include -diatoms, nematode, polychaetes, fish eggs, algae, detritus, along with particles of mud and sand (Ravi 2013). It has been shown, according to the study conducted with *Periophthalmodon schlosseri*, that there are differential preferences in the selection of food products between male and female. Females preferred small-scale fish, namely *Oryzias* sp., and males over small-scale fiddler crabs because they were highly involved in land activities (Zulkifli et al. 2012).

The *periophthalmus sobrinus* feeds on small animals. Their food sources are polychaetes, polydora, terebellid, nematodes, crustaceans, copepods, tanaids, prawns, schizopod larvae, alpheid shrimp juveniles, *Uca chlorophthalmus*, and tiny sand crab (Stebbins & Kalk 1961). The discerned food items of *Periophthalmus waltoni* are crustaceans (high occurrence), snails (slightly lower occurrence), 3.4% insects (lower appearance), and fishes (least) (Mhaisen & Al-maliki 2013). *Boleophthalmus pectinirostris* selects their meal based on the abundance, availability, size of the diatoms and temperature. The size of the species reflects over its feeding apparatus (Yang et al. 2003).

Boleophthalmus boddarti likewise feeds on Bacillariophyta (Quang 2015). The anatomical characteristics in *Periophthalmus kuelreuteri* are examined by light and X-ray cinematography. It pushes forward with its pelvic fins when the prey gets near, and brings the jaws near to the target. Opercular bones, aid swallowing into the pharynx to position the prey. They found it easy to capture the prey in the land by biting and open mouth. It reaches the stomach via an esophagus after passing the pharyngeal jaws (Sponder & Launder 1981).

Reproduction

Studies were conducted in Nigeria's lagoon swamps, which concentrated primarily on sex ratios, egg

diameters, gonadosomatic index and levels of maturation (Lawson 2010). Reports on the growth of eggs with artificial fertilization and larval reports in the species *Periophthalmus cantonensis* were carried out (Tshako et al. 2003). The mudskipper's eggs were laid deep within the burrows where there is a hypoxic environment, and male mudskippers supply the oxygen by depositing oxygen through water. Once the development of the eggs is complete, they are released from the burrows and hatched by tides, making them prepare themselves during severity (Ishimatsu et al. 2007). A research on *Apocryptes bato's* (Gobiidae) reproductive biology in the Payra River, southern Bangladesh, helped to understand the basics of reproduction in mudskippers (Ferdous et al. 2018). Studies related to reproductive biology, fertilization, maximum sizes achieved along with spawning will allow local fishermen to know the exact time of catching them, rather than disturb them during spawning seasons. The concept was prominent from the studies of sex maturation of *Boleophthalmus boddarti* (Quang et al. 2015). The species *Periophthalmodon septemradiatus* lays eggs year-round. Observing their complete duration during the maturation stages helped to understand reproductive biology and to learn about the methods of conservation (Dinh et al. 2018). Studies at intertidal swamps of the Imo River estuary dealt about the reproductive biology of *Periophthalmus barbarous* along with growth, mortality, recruitment pattern, gonadosomatic index, and spawning season (Etim et al. 2002).

Studies on nutrition content and medicinal value

Nutrient content of three species *Periophthalmus waltoni*, *Boleophthalmus dussumieri*, and *Scartelaos histophorus* were measured, such as starch, protein, and lipid. *Boleophthalmus dussumieri* is rich in nutrients and it comes from the liver. Because of this adventitious effect, people consume them in the Bhavnagar coast, Gujarat (Kanejiya et al. 2017a). Mudskippers are known to be very rich in proteins and other nutrients. They are either used in traditional medicines in countries like Malaysia or as bait and also utilized for consumption. During winter, mudskippers are caught and sold at the market by fishermen using net trap methods (Kanejiya et al. 2017b). Likewise, there are growing demands for mudskippers at Narmada estuary, Gujarat, and they are favoured by local people. As a result of this, the ecosystem can get highly exploited and can impact biodiversity (Bhakta et al. 2018). Though mudskipper meat possess nutritional benefits there are researchers stating about metal toxicity and bioaccumulation when

consumed.

The research by Looi et al. (2016) concluded that the bioaccumulation of mercury in *Periophthalmodon schlosseri* did not cause any serious effects when consumed. From the evaluation of the nutrient content of *Pseudapocryptes elongatus*, it has been shown to contain sufficient protein, carbohydrate, lipid, important and non-essential amino acids, polyunsaturated fatty acid (PFA) with a greater amount than saturated fatty acid (SFA). For consumption it is highly recommended because of its nutritive value (Mahadevan et al. 2019b). Mudskippers are widely available in Nigeria and their prices are comparatively low, and they are consumed by the elderly (Edun et al. 2010). Exposure of mudskipper *Boleophthalmus boddarti* to natural radionuclides (^{238}U , ^{226}Ra and ^{210}Pb and ^{210}Po), the radionuclide concentration was found to be below the limit and therefore, did not have a profound impact in Bombay Harbour and coastal zone (Bangera & Patel 1984).

Threats to mudskippers

The effect on mudskipper density was recorded in the mudflats of Hathab coast, Gujarat, considering salinity and temperature as independent variables. Accordingly, mudskipper abundance and distribution have become a vital indicator for determining intertidal region health (Kanejiya et al. 2017c). Post-tsunami studies at Mudasolodai, Tamil Nadu, revealed that rapid changes in soil morphology led to changes that directly affected the mudskippers' livelihood. For construction, they usually prefer clay-rich soil, rather than sandy soil (Ravi 2005). Although metal uptakes were increasingly high during lower salinity, the levels of salinity in the intertidal mudskipper *Periophthalmus cantonensis* did not affect the metal (Cd, Se, & Zn) concentration factors (CF) (Ni et al. 2006). Mudskippers face other menaces due to metal toxicity. To illustrate, when exposed to Cr (VI), *Boleophthalmus dentatus* causes affection in the activity of Na^+ , K^+ and ATPase, and further affects membrane activity, also causing metabolic stress (Kundu et al. 1995). Similarly, the same has also been deduced in other studies pertaining to *Boleophthalmus dentatus* obtained from the Gulf of Katch (Lakshmi et al. 1991). The shrimp effluents collected at the mudskipper habitat in the northern Persian Gulf have proved to support mudskipper enrichment, which helps them increase their species density and length. Despite severe changes and reduced dissolved oxygen, shrimp effluents create favourable physical conditions for their growth; it survived amid these adversities (Kohan et al. 2018). Mudskippers are known for their versatile



behaviour and amphibious character. They have less understanding of the public and have lost interest. The most striking characteristic of them is their survival ability and adaptation to different environmental changes. Surprisingly, mudskippers feed on the available foodstuffs, rather than relying on specific foods. Further research can be done with them as they define themselves as an ideal candidate for aquaculture. Gills' anatomical features are specifically built to suit both the terrestrial and aquatic lifestyles. They got adapted from the initial stages of egg production to adjust to severities such as high tide, and airflow maintenance.

CONCLUSION

Mudskippers are known for their significant behaviour of adapting themselves to different environment and amphibious nature. Mudskippers are less popular among people and aren't noticed much. Their most amazing feature is their capacity to survive and their adaptation to various changes in the environment. Surprisingly, mudskippers feed on the foodstuffs available, rather than rely on particular foodstuffs. Further research can be carried out as they may be an ideal candidate for aquaculture. The anatomical characteristics of gills are designed primarily to accommodate both the terrestrial and aquatic lifestyles. From the initial stages of egg production, they are taught to adjust themselves to severities such as high tide and maintenance of the airflow. The most striking aspect about them is their position in coastal areas as a biological indicator, pollution monitor, and estuarine health monitor.

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Diversity, distribution and conservation status of the Adder's-tongue ferns in Goa, India

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Abstract: The cosmopolitan fern genus *Ophioglossum* (Ophioglossaceae) is distributed from low to high altitude plateaux, coastal plains and forest floors of India. This genus has received special attention from pteridologists worldwide since the discovery that *Ophioglossum reticulatum* possesses the largest number of chromosomes. There are, however, no reported studies of *Ophioglossum* in Goa, hence the present investigation was undertaken to study the diversity, distribution and conservation status of *Ophioglossum* in that state. A total of six species were collected from different localities, of which four (*O. nudicaule*, *O. lusitanicum*, *O. parvifolium*, and *O. reticulatum*) are reported as new distributional records for Goa State. A detailed morpho-taxonomy, illustration and photographs of all collected species are given, along with a key to the species.

Keywords: Illustration, morpho-taxonomy, new record, Western Ghats.

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INTRODUCTION

The state of Goa is located on the western coast of India between Maharashtra and Karnataka states, with a geographical area of 3,702km². It lies on the coastal plains of the Western Ghats between 14.899–15.799 °N and 3.681–74.336 °E. The major rivers are Mandovi and Zuari (Meteorological Centre, Goa 2020), and the climatic features fall under the tropical monsoon climate, making the region generally warm and humid throughout the year. The average rainfall ranges between 2,500–4,500 mm/annum, and highest rainfall is observed during June–August. The average temperature ranges between 16.2°C–36.7°C, with maximum temperature during the month of April–May. The state has a national park (Molem) and six wildlife sanctuaries which covers an area of 755km² (Hiremath 2003; Jadhav & Patil 2012).

The genus *Ophioglossum* L. belongs to the primitive family Ophioglossaceae. About 50 species are accepted internationally (Hassler & Schmitt 2020), and 19 species are documented from India (Patil & Dongare 2014; Patil et al. 2018 & Kachhiyapatel et al. 2018), of which the status of a few were unresolved (Hassler & Schmitt 2020). All are terrestrial forms except the epiphytic *O. pendulum* L. The genus was studied in India for the first time by Beddome (1883). After his monumental contribution on the Indian ferns, the genus *Ophioglossum* L., received further attention from Blatter & d'Almedia (1922), Chakravarty (1951), Mahabale (1962), Panigrahi & Dixit (1969), Khandelwal (1987), Khullar (1994), and Goswami (2007), in their respective works. It is characterized by simple trophophyll with a spike (rarely bi or trifurcate) bearing two rows of sporangia. In all species the trophophyll is pale green, green, dark or light green at maturity, and produces a paler spike (Khullar 1994; Goswami 2007; Goswami et al. 2008). *Ophioglossum gomezianum* Welw. ex A.Braun, *O. indicum* B.L.Yadav & Goswami, *O. lusitanicum* L., and *O. rubellum* Welw. ex A.Braun, however, produce yellow, pink, reddish to brown tinge of trophophyll, respectively (Goswami 2007). Patil & Dongare (2014) studied the diversity and distribution of *Ophioglossum* from the Western Ghats and reported 06 species. Recently, Fraser-Jenkins et al. (2018) accepted only 12 species for India.

The diversity of pteridophytes in Goa is less understood, with few reports. Dalgado 1898; Blatter & d'Almedia 1922; Vartak 1966; Rao 1985–1986; Naithani et al. 1997, and Irudayaraj & Bir 1997. Manickam et al. (2004) studied the pteridophytes of the Western Ghats of Goa and recorded 51 species. Further, Kerkar & Shetkar (2009) studied the order Pteridales and recorded 16

species in Goa belonging to six genera from five families. In 2010, Datar & Lakshminarasimhan (2010) studied the pteridophyte flora of the Western Ghats of Goa and documented 47 species. These studies reported two species of *Ophioglossum*: *O. costatum* R.Br., and *O. gramineum* Willd., from the Western Ghats of Goa. While studying the pteridophytes of the Western Ghats, however, the authors visited different places in Goa and observed that the plateaux and coastal plains are favourable for *Ophioglossum* and earlier workers may have missed some species. Thus the present investigation was undertaken to study the morpho-taxonomy, species composition, distribution and conservation status of *Ophioglossum* from Goa.

MATERIALS AND METHODS

Field visits were carried out during 2014–2018 in different areas of Goa State. The collected specimens of *Ophioglossum* were processed in the laboratory. Pressed specimens were fixed with 4% formalin and affixed to herbarium sheets using synthetic gum (Fevicol). Specimens were identified using literature, including Blatter & d'Almedia (1922); Panigrahi & Dixit (1969); Fraser-Jenkins et al. (2017), and Patil & Dongare (2014). The voucher specimens are deposited in the herbarium of the Department of Botany, Maharaja Sayajirao University of Baroda, Vadodara, Gujarat (BARO).

RESULTS

During the present investigation authors collected six species of *Ophioglossum*: *O. costatum* R.Br., *O. gramineum* Willd., *O. lusitanicum* L., *O. nudicaule* L.f., *O. parvifolium* Grev. & Hook., and *O. reticulatum* L. The detailed morphology, diversity, distribution, illustrations, photographs, phenology, ecology, and conservation status of each species is given. An identification key for the taxa recorded from Goa State is given below.

Key to the species for Goa state

- 1a. Trophophylls having costa or yellow band at centre *O. costatum*
- 1b. Trophophylls without costa or yellow band at centre 2
- 2a. Trophophylls base cordate, strobili having > 20 pairs of sporangia *O. reticulatum*
- 2b. Trophophylls base not cordate, strobili having

- < 20 pairs of sporangia 3
- 3a. Trophophylls attached to substratum *O. parvifolium*
- 3b. Trophophylls above the ground 4
- 4a. Trophophylls ovate-lanceolate, 1–2 cm above the ground *O. nudicaule*
- 4b. Trophophylls linear-lanceolate or spatulate not flat on ground 5
- 5a. Trophophylls linear-lanceolate or grass like *O. gramineum*
- 5b. Trophophylls spatulate-lanceolate not grass like *O. lusitanicum*

Ophioglossum costatum R. Br., Prod. Fl. Nov. Holl. 163. 1810. Panigrahi & Dixit, Proc. Nat. Inst. Sci. India 35: 249. 1969; Patil & Dongare, Indian Fern J., 31: 17–24. 2014.

Lectotype: from Australia, Queensland, Arnhem North Bay, R. Brown 118. 14.2.1803.

Ophioglossum pedunculatum Desv., Mag. Nat. Fr. Berlin 5: 306. 1811.

O. brevipes Bedd., Ferns. Southern India 23. t. 72. 1863.

O. bulbosum Bedd., Ferns. Brit. India Suppl. t. 28. 1876.

O. fibrosum Schum., Bedd., Handb. 465. t. 289. 1883.

Plant terrestrial, 12–25 cm in height, pale-green; rhizomorph 0.5–1 cm, subterranean, disc like or globose, bearing numerous yellow-brown, fleshy, unbranched, roots; common stalk 2–3 cm, subterranean-terrestrial; trophophylls 2–6 x 0.5–1 cm, 1–3 simple, elliptic-lanceolate, apex acute-apiculate or obtuse-round, base cuneate, margin entire, green-pale green, glabrous on both sides, coriaceous, costa present, prominent, yellow; texture coriaceous, thick; veins indistinct, simple reticulate, anastomosing; fertile segment 9–19 cm, unbranched, inserted on adaxial position of leaf; strobili 2–5 cm, linear-lanceolate, apex blunt-lanceolate, 20–55 pairs of sporangia, pale yellow, unbranched; spores 20–40 µm in diameter, trilete, foveolate.

Phenology: sterile phase – June–July; fertile phase: July–September (rarely in October)

Distribution: India (Andhra Pradesh, Assam, Chhattisgarh, Goa (Canacona, Dharbandora, Mapusa, Pernem, Phonda, Quepem, & Sattari), Gujarat, Himachal Pradesh, Jharkhand, Karnataka, Kerala, Madhya Pradesh, Maharashtra, Odisha, Rajasthan, Tamil Nadu, Uttar Pradesh, and West Bengal), Bangladesh, Indonesia, Malaysia, Philippines, Sri Lanka, Thailand, Africa, and Australia.

Ecology: Very common, collected from grassy plateau. The population size varies according to the water content of the soils. The population size is larger when water availability is more and vice-versa.

Conservation status: It is collected from throughout the Goa except seashore areas. Population comprises about 300–400 individuals per km² and the area of occupancy (AOO) is 50–60 km². Therefore, as per IUCN categories and criteria (IUCN red list of Threatened Species ver. 2017-1), it is assessed as Least Concerned (LR) species for Goa State.

Ophioglossum gramineum Willd. Nov. Act. Acad. Erfurt. 2: 18. t.f.1. 1802; Beddome, Handb. Suppl. Ferns Brit. India 108. 1892; Balakrishnan et al., Bull. Bot. Surv. India 2: 337. 1960; Panigrahi & Dixit, Proc. Nat. Inst. Sci. India 35: 250. 1969; Patil & Dongare, Indian Fern J. 31: 17–24. 2014.

Type: from West Africa, St. Thomae, De Friedland B. *Ophioglossum dietrichiae* Prantl, Ber. Deut. Bot. Ges. 1: 352. 1883.

O. gregarium Christ, Nova Guinea Bot. 8: 164. 1909;

O. inconspicuum (Racib.) Alderw., Bull. Dépt. Agric. Ind. Néerl. 21: 9. 1908.

O. prantlii C.Chr., Ind. Fil. 2: 471. 1906.

O. inconspicuum forma *majus* Alderw., Bull. Dépt. Agric. Ind. Néerl. 21: 9. 1908.

Ophioglossum gramineum var. *majus* (Alderw.) Wieff., Blumea 12(2): 324. 1964. *Ophioglossum gregarium* Christ, Nova Guinea, Bot., 8: 164. 1909.

Ophioglossum gracile Pocock ex J.E.Burrows, Bothalia 25(1): 61. 1995.

O. vulgatum var. *gramineum* (Willd.) Hook. f., Fl. Nov. Zel. 2: 50. 1854.

Plant terrestrial, 4–8 cm in height, green-pale green; rhizomorph sub-globose-tuberous, subterranean, bearing numerous, fleshy, fibrous roots; common stalk 0.5–2 cm, subterranean-terrestrial, flat; trophophylls 1–2 cm, 1–2, linear grass-like, apex acuminate, margin entire, soft, green-pale green; texture coriaceous, thin; veins parallel, anastomosing, forming parallel areoles; fertile segment 3.5–6 cm, unbranched, pale green, flat-round; strobili 0.5–1.5 cm, linear-lanceolate, apex pointed, 6–10 (rarely >10) pairs of sporangia, arranged in two alternate rows, pale green-yellow; spores 25–40 µm dia., trilete, exine reticulate.

Phenology: Sterile phase: Jun–July; fertile phase: July–August (rarely September–October)

Distribution: India (Andhra Pradesh, Chhattisgarh, Goa (Canacona, Dharbandora, Mapusa, Pernem, Phonda, Quepem, & Sattari), Gujarat, Karnataka, Kerala, Madhya

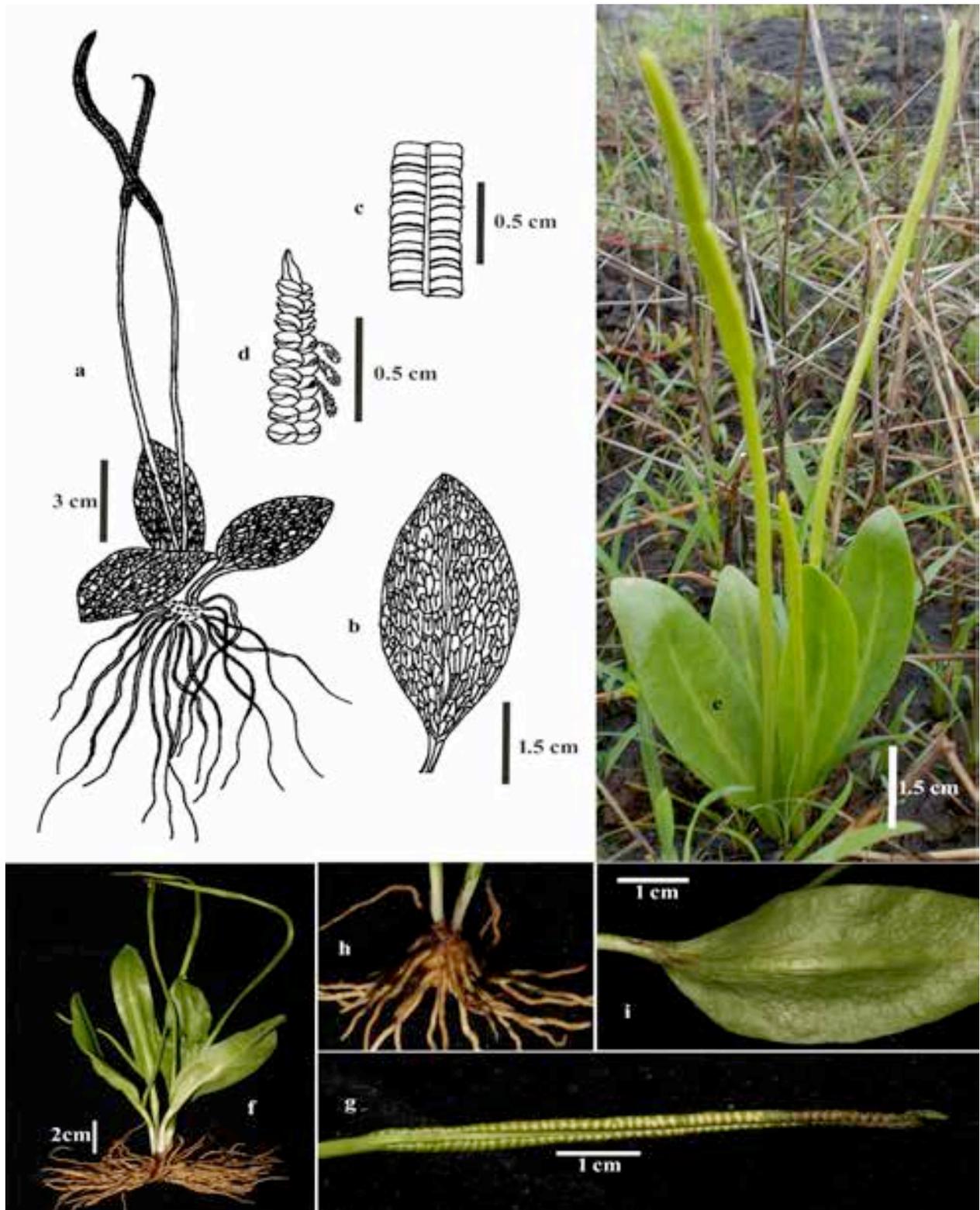


Image 1. *Ophioglossum costatum*: a, e-f—habit | b—enlarged trophophyll showing venation | c-d, g—enlarged strobilus | h—globous rhizomorph | i—trophophyll showing costa. © Sachin M. Patil

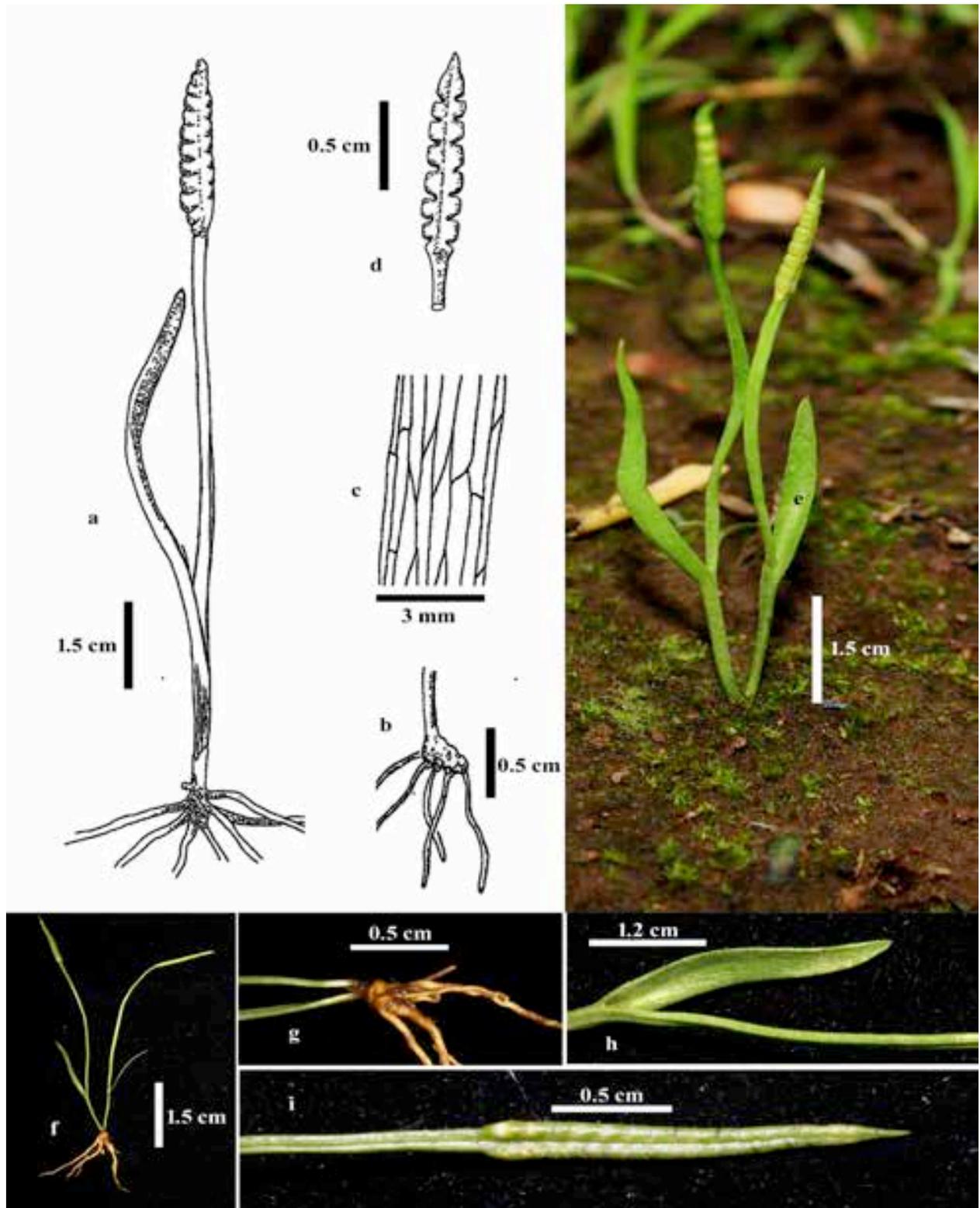


Image 2. *Ophioglossum gramineum*: a, e-f—Habit | b, g—enlarged rhizomorph | c—enlarged view of venation pattern | d, i—enlarged strobilus | h—enlarged trophophyll. © Sachin M. Patil

Pradesh, Maharashtra, Meghalaya, Rajasthan, Tamil Nadu, Tripura, Uttarakhand, Uttar Pradesh, & West Bengal), Indonesia, Malaysia, Myanmar, Philippines, East Borneo, East Java, New Guinea, Vietnam, Africa, and Sri Lanka.

Ecology: Common fern collected from grassy plateaux associated with *O. nudicaule*, *O. parvifolium* and *O. costatum*.

Conservation status: It is collected from plateaux situated in Goa State. Population comprises about 300–400 individuals per km² and the area of occupancy (AOO) is 50–60 km². Therefore, as per IUCN categories and criteria (IUCN red list of Threatened Species ver. 2017-1), it is assessed as Least Concerned (LR) species for Goa State.

Ophioglossum lusitanicum L., Sp. Pl., 2: 1063. 1753; Clausen, Mem. Torrey Bot. Club, 19 (2): 159. 1938; Mahable, Bull. Bot. Surv. India, 4: 71. 1962; Panigrahi & Dixit, Proc. Nat. Inst. Sci. India 35: 251. 1969; Patil & Dongare, Indian Fern J., 31: 17-24. 2014.

Type: from Portugal, "Habitat in Lusitania", not designated.

Ophioglossum loureirianum C.Presl, Suppl. Tent. Pterid. 55. 1845.

O. braunii Prantl, Ber. Deutsch. Bot. Ges. 1: 351. 1883.

Plant terrestrial, 3–10 cm in height, small; rhizomorph 0.5–1 cm, subterranean, sub-globose-tuberous with or without stoloniferous fleshy roots; common stalk subterranean, white; trophophylls 1–2, erect, red-brown-green, spatulate, linear-lanceolate or elliptic-lanceolate, acute-acuminate apex, cuneate-attenuate base, entire margin; texture coriaceous, thin; veins indistinct, anastomosing, forming parallel areoles; fertile segment round, unbranched, green-yellow brown; strobili 1–2 cm, linear-lanceolate, 8–10 (rarely >10) sporangia in two rows, green-yellow; spores 20–25 µm dia., trilete, exine reticulate.

Phenology: sterile phase: June–July; fertile phase: August–September

Distribution: India (Andhra Pradesh, Assam, Bihar, Goa (Mapusa, Pernem, & Phonda), Jammu & Kashmir, Karnataka, Kerala, Madhya Pradesh, Maharashtra, Odisha, Punjab, Tamil Nadu, & Uttar Pradesh), Algeria, Morocco, Portugal, Spain, Tanzania, Tunisia, and Uganda.

Ecology: Common species, collected from grassy plateaux of Goa, associated with *O. costatum*, *O. nudicaule*, *O. parvifolium* and *O. reticulatum*.

Conservation status: It is collected from plateaux situated at Mapusa, Pernem, and Phonda. A population of about 100–200 individuals was found. The area of

occupancy (AOO) is 10–20 km² per locality and considered as Data Deficient (DD) because the explorations in the state are not completed and there is a possibility of more locations of occurrence.

Ophioglossum nudicaule L.f., Suppl. Pl. Syst. 443. 1781; Beddome, Handb. Ferns Br. India, 464, t. 228. 1883; Panigrahi & Dixit, Proc. Nat. Inst. Sci. India 35: 252. 1969; Manickam & Irudayaraj, Pterid. Fl. West Ghats 48-49. t. 27. 1992; Patil & Dongare, Indian Fern J. 31: 17-24. 2014.

Type: South Africa: Cape of Good Hope.

Ophioglossum capense Sw., Schard. Journ. 1801(2): 308. 1803.

Ophioglossum capense Schlech. var. *nudicaule* (L.) Schlech., Fil. Prom. Bonae Sp.: 9. 1825.

Ophioglossum ellipticum Hook. & Grev., Icon. Filic. t. 40 A. 1828.

Ophioglossum lineare Schlechter & Brause, Bot. Jerb., 49: 59, fig. 3F. 1912.

Ophioglossum luersseni Prantl, Ber. Deut. Bot. Ger. 1: 352. 1883.

Plant 4–12 cm height, green, terrestrial herb; rhizomorphs subterranean, sub-globose (at young)-tuberous (at maturity), bearing many soft, pale brown, unbranched, fleshy, fibrous, stoloniferous roots; common stalk 1–3 cm, subterranean-terrestrial, white (subterranean), green (terrestrial); trophophylls 0.5–2 x 1–2 cm, 1–2 (rarely 3), green, tuft ovate-elliptic, apex acute-obtuse, base cuneate, margin entire, glabrous; veins indistinct, simple reticulate, anastomosing, with or without included veinlets; fertile segment 2–8 cm, unbranched (rarely branched), green at young, yellow at maturity; strobili 1.5–2.5 cm long, with 10–20 sporangia per strobilus, arranged in two alternate rows, linear-lanceolate, apex pointed; spores 30–40 µm dia., trilete, with reticulate ornamentation.

Distribution: India (Andhra Pradesh, Himachal Pradesh, Jammu & Kashmir, Goa (Phonda & Canacona), Gujarat, Jharkand, Karnataka, Kerala, Maharashtra, Sikkim, Tamil Nadu, & West Bengal), China, Indonesia, Malaysia, and Thailand.

Phenology: sterile phase: July–August; fertile phase: August–September.

Ecology: The species is growing in patches on fully exposed plateaux or open grassland, associated with *O. costatum*, *O. gramineum* and *O. parvifolium*.

Conservation status: It is collected from open grasslands on plateaux situated in Phonda and Canacora. The area of occupancy (AOO) is 10–20 km² per locality and considered as Data Deficient (DD) because the explorations in the state are not completed and there is



Image 3. *Ophioglossum lusitanicum*: a, e–g—Habit | b, i—enlarged rhizomorph | c—enlarged trophophyll showing venation | d, j—enlarged strobilus | h—enlarged trophophyll. © Sachin M. Patil

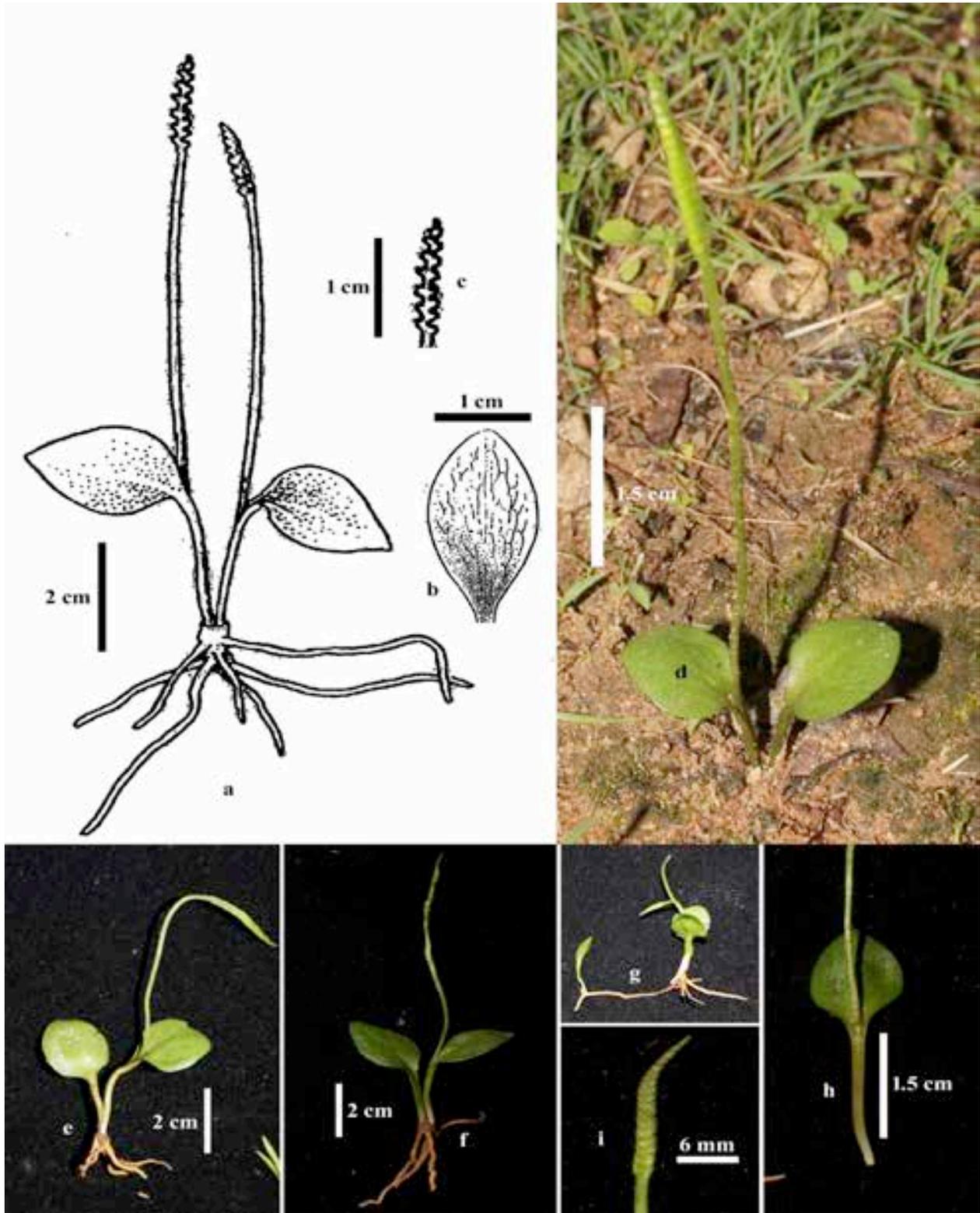


Image 4. *Ophioglossum nudicaule*: a, d-f—Habit | b, h—enlarged trophophylls | c—enlarged strobilus | g—stoloniferous roots. © Sachin M. Patil

possibility of more locations of occurrence.

Ophioglossum parvifolium Grev. & Hook., Bot. Misc. 3: 218. 1833; Patil & Dongare, Indian Fern J. 31: 17–24. 2014.

O. macrorrhizum Kunze Analecta Pteridogr.: 2. 1837.

O. schmidii Kunze in Linnaea 24: 246. 1851.

O. luerssenii Prantl Ber. Deutsch. Bot. Ges. 1: 352. 1883.

O. pumilum (Racib.) Alderw. Malayan Ferns: 774. 1909.

Type: India, Gujarat, Valsad District, Wilson Hills, 24 Aug 2014, Patil & Kachhiyapatel 38 (BARO No. 1234500700). *typ. cons. prop.* (Mazumdar et al. 2018)

Plant 4–8 cm, terrestrial herbs; rhizomorph subterranean, sub-globose-tuberous, pale brown, bearing many soft, fleshy, unbranched, fibrous, stoloniferous roots; common stalk subterranean, white, round, glabrous; trophophylls 0.4–1 cm, 1–2 (rarely 3–4), pale green-green, ovate-lanceolate, apex acute-apiculate, margin entire, base cordate; texture coriaceous, thin; *veins* indistinct, simple reticulate, four or five veins passing up through the stalk of the blade; fertile segment 2–6 cm long, unbranched, green-pale green; strobili 0.5–1 cm long, 5–7 (rarely >10) pairs of sporangia, arranged in two alternate rows; *spores* 25–45 µm dia., trilete, exine reticulate.

Distribution: India (Madhya Pradesh, Goa (Canacona, Dharbandora, Mapusa, Pernem, Phonda, Quepem, & Sattari), Gujarat, Maharashtra, Karnataka, Rajasthan, Kerala, & Tamil Nadu), China, South America, Sumatra, Malaysia, and Thailand.

Ecology: Common species grows in patches on fully exposed areas from low land to high land areas.

Conservation status: It is collected from different plateaux of Goa state. Population comprises about 300–400 individuals per km². The area of occupancy (AOO) is 50–60 km². Therefore, as per IUCN categories and criteria (IUCN ver. 2017-1), it is assessed as Least Concerned (LC) species for Goa state.

Ophioglossum reticulatum L. Sp. Pl. 2: 1063. 1753; Beddome, Ferns. Southern India 23. t. 70. 1863; Beddome, Handb. Ferns. Bri. India, 465. t. 290. 1883; Panigrahi & Dixit, Proc. Nat. Inst. Sci. India 35. 257. 1969; R.D. Dixit, Cens. Ind. Pterid. 24. 1984; Manickam & Irudayaraj, Pterid. Fl. West Ghats: 51. t. 29. 1992; Patil & Dongare, Indian Fern J. 31: 17–24. 2014.

Lectotype: (Tardieu Blot, in Aubreville, Flor. Gabon 8: 30. 1964): from C. America, "*Ophioglossum cordatum* et *reticulatum*" in Plumier, Traité Foug. Amér., 141, t. 164,

1705.

Ophioglossum peruvianum Presl, Suppl. Tent. Pterid 52. 1845.

Ophioglossum petiolatum sensu Wieffering, Blumea, 12: 327. 1964.

Ophioglossum cordifolium Roxb., Hort. Bengal. 75. 1814;

Plant 15–25 cm, terrestrial herb; rhizomorph 3–7 mm long, subterranean, tuberous, bearing few long, thick fleshy unbranched, pale brown roots, with 1–2 trophophylls; common stalk 2–5 cm, terranean, white at base, green above; trophophylls 1, 2–5 x 1–3 cm, cordate-broadly ovate, rarely elliptic-oblong, apex rounded or acute, base cordate; veins reticulate, areoles with or without included free veinlets; fertile segment 12–18 cm long, round; strobili 1–5 cm, 20–45 pairs of sporangia, apex pointed, sporangia arranged in two alternate compact rows; spores 30–45 µm dia., spherical, trilete, exine hemispherical.

Distribution: India (Andhra Pradesh, Arunachal Pradesh, Assam, Bihar, Chhattisgarh, Goa (Sattari & Canacona), Gujarat, Himachal Pradesh, Jammu & Kashmir, Jharkhand, Karnataka, Kerala, Madhya Pradesh, Maharashtra, Meghalaya, Mizoram, Odisha, Rajasthan, Sikkim, Tamil Nadu, Tripura, Uttarakhand, Uttar Pradesh, & West Bengal), Malay Peninsula, tropical America, Africa, Philippines, and Sri Lanka.

Phenology: sterile phase: July–August; fertile phase: August–October

Ecology: Very common, found beneath trees in the forest along with *Ophioglossum costatum*, *O. gramineum* and *O. parvifolium*.

Conservation Status: It is collected from forest regions at Sattari and Canacona. The area of occupancy (AOO) is 10–20 km² per locality and considered as Data Deficient (DD) because the explorations in the state are not completed and there is possibility of more locations of occurrence.

DISCUSSION

Earlier only two species, *Ophioglossum costatum* and *O. gramineum* were reported from Goa (Manikam et al. 2004; Datar & Lakshminarsimhan 2010). During the present investigation six species were collected, of which *O. lusitanicum*, *O. nudicaule*, *O. parvifolium*, and *O. reticulatum* have been added to the flora of Goa. *O. costatum*, *O. gramineum*, and *O. parvifolium* are considered least concerned (LC) species, while *O. lusitanicum*, *O. nudicaule*, and *O. reticulatum* are

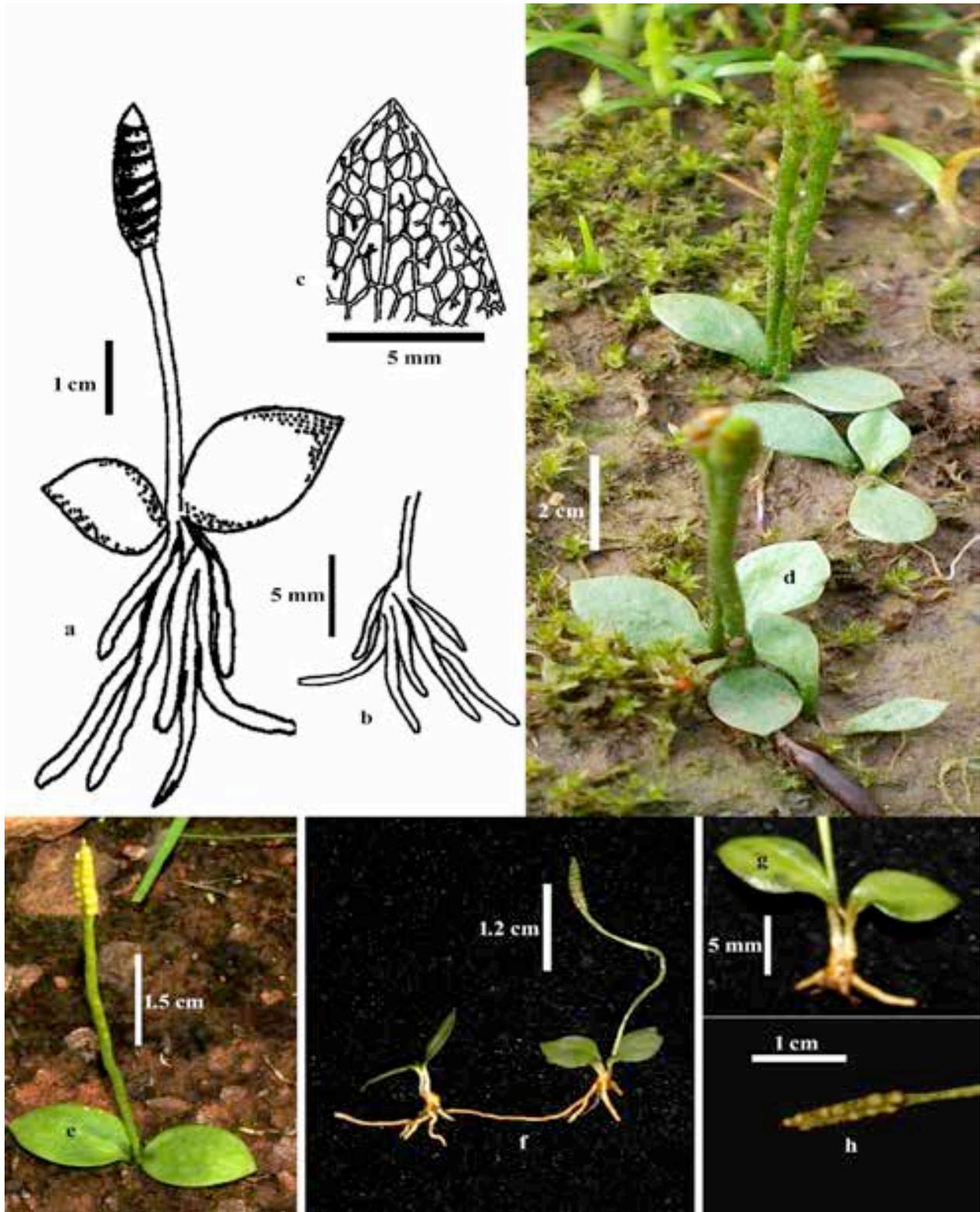


Image 5. *Ophioglossum parvifolium*: a, d–e—Habit | b, g—enlarged rhizomorph | c—enlarged trophophylls showing venation | f—stoloniferous roots | h—enlarged strobilus. © Sachin M. Patil

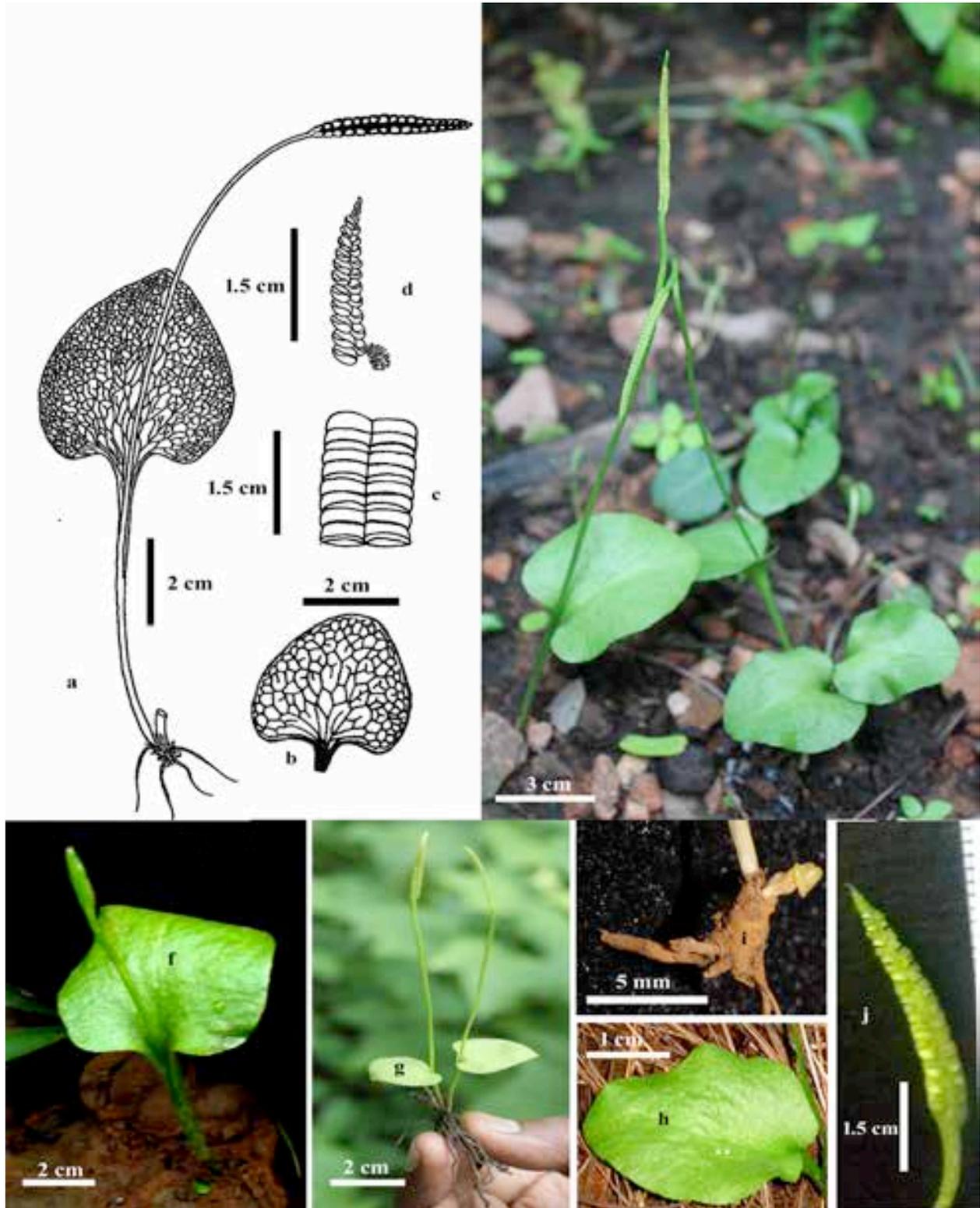


Image 6. *Ophioglossum reticulatum*: a, e-g—Habit | b—enlarged trophophyll showing venation | c-d, j—enlarged strobilus | h—enlarged trophophyll | i—rhizomorph. © Sachin M. Patil

considered data deficient (DD) species. The maximum diversity was observed in open grassy habitat and plateaux, where we collected *O. costatum*, *O. gramineum*, *O. lusitanicum*, *O. nudicaule*, and *O. parvifolium*. Coastal plains contained *O. costatum* and *O. gramineum*, and on forest floors we collected *O. nudicaule* and *O. reticulatum*.

In *Ophioglossum costatum* the trophophylls (leaf-like segment) have a central yellow band (i.e. costa), hence the specific epithet is *O. costatum*. Amongst the taxa reported from India, *O. costatum* is only one with a costa (Image 1), thus it can be easily differentiated from other Indian residents. In the field it is difficult to locate *O. gramineum* because this species is found on grassy plateaux and its appearance is similar to a grass (hence specific epithet *O. gramineum*). Generally the trophophylls are linear-lanceolate like grass, and are part underground and part above ground with a common stalk (Image 2). A species similar to *O. gramineum* is *O. lusitanicum*, however, the latter has green-brown, spatulate-lanceolate trophophylls and a subterranean common stalk (Image 3). *Ophioglossum nudicaule* and *O. parvifolium* are closely allied and commonly confused species. *O. nudicaule* is smaller in size, hence the specific epithet is *parvifolium* (*parvum* = small and *folium* = trophophylls). *Ophioglossum nudicaule* has trophophylls that are ovate-obovate, 1–2 cm above the ground, common stalk subterranean-terrestrial, whereas in *O. parvifolium* the trophophylls are ovate-broadly lanceolate, attached or flat on the ground, common stalk subterranean only (Image 4–5). *Ophioglossum reticulatum* is well-known and popular in biological world because it has the highest number of chromosomes ($n=740$). It is allied and confused with *O. petiolatum*, however, *O. reticulatum* has cordate trophophylls, 2–5 cm above ground, petiolate, common stalk subterranean-terrestrial, having maximum number of sporangia (Image 6).

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An inventory of the native flowering plants in East Siang District of Arunachal Pradesh, India

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Abstract: The present study is an outcome of floristic surveys of East Siang District of Arunachal Pradesh, carried out during 2016–2019, and also a compilation of earlier published reports. Vegetation analysis of this area along with a checklist of 508 taxa is presented. A total of 503 species, one subspecies and four varieties of native flowering plants belonging to 348 genera and 102 families are reported. Among these, 11 taxa are endemic to India, two Critically Endangered, one Vulnerable, one Near Threatened, two Data Deficient, and others either Least Concern or Not Evaluated as per IUCN criteria. The study also documents two new distributional records for the flora of Arunachal Pradesh, and range extension of six lesser-known endemic species. The most dominant families were found to be Poaceae (27 species), followed by Lamiaceae (23 species), Gesneriaceae (22 species), and Rubiaceae (20 species). The number of new taxa described from the region, endemism, and the Red Listed plants strongly reflect the floristic importance of the region, which is in dire need of conservation.

Keywords: Checklist, conservation, endemism, Himalayan flora, northeastern India, taxonomy.

Abstract (in Adi language): Ager Sim East Siang District Arunachal Pradesh lo 2016 – 2019 Ditag delo Ito. East Siang lo deddine nesi - neyang em pado, ajokon ee Tani gidangso India petom lo Kapanekom kado delokke akon akon ee Arunachal Pradesh lok East Siang Goralok nyomrang kider petom lo panekom kado. Ditag anyi aum solo East Siang lok Nesi-Neyang Researchers kider ee deddine ani-ani neyang em report delokke new species discoveries em itung. Deddine angu angu kangki kangki manam nesi- neyang em padoaai idola ajo kon nesi neyang kider si nyoknam lo adung aipe kajun tatjun la bulum ijun mamil apena ditag kider lo bulu tani among holok nyoknam lo kadung. Nesi Neyang si Tani ngolum delokke Simon-sili, Takom-taruk lope ager abido. Donam tiinam lokke ila dungkeng - dakkeng, kusureng dadi lope ager amangko kamangdo. Bulu mai nguluk turkeng ngakeng em bilenne ila bulum ngulukom ayang pe ido delokke atel kokom tani among sok nyokmo mape aido. Ngolu east siang monam lo magola ditag aum 508 Nesi- Neyang ko kalen malen la atlung. Akokom nyomrang kider em gairupe mamil deddine malen penam ee kadungaa.

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Author details: MOMANG TARANG, a PhD scholar, has authored more than 20 research articles and has described 10 taxa new to science. She is involved in two major projects 'Biocultural studies of the Adi tribe in Arunachal Pradesh' and 'Revision of the family Gesneriaceae in Arunachal Pradesh'. DIPANKAR BORAH has authored several research articles and has described 11 taxa new to science. He is interested in native and endemic vascular flora of northeastern India as well as ethnobotany of the region. His current research project is 'Enumeration of the biodiversity of Behali Reserve Forest, Biswanath, Assam.' HUI TAG has been involved in biocultural studies of the ethnic communities of Arunachal Pradesh. He is also working in the field of nutraceutical and drug discovery targeting some medicinal plants. RITESH KUMAR CHOUDHARY has worked on the flora of all four biodiversity hotspots of India, and also in Vietnam and Korea. He has described 12 new plant species, authored three books and more than 70 research papers. He is actively engaged in resolving the taxonomy and phylogeny of plants using molecular data. He has been recently nominated for Prof. V.V. Sivarajan Gold Medal by Indian Association of Plant Taxonomy for his contribution to the field of plant taxonomy.

Author contribution: MT, DB, HT, and RKC conceptualized the research. MT and DB carried out the field work. All authors contributed in data compilation, analysis and writing of the manuscript.

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INTRODUCTION

Northeastern India is situated in the transition zone between the Himalayan and Indo-Burma global biodiversity hotspots (Olson & Dinerstein 1998). The region has a long international boundary of about 2,000km with China and Bhutan in the north, Myanmar in the east, Nepal in the west, and Bangladesh in the south-west; it harbours more than one-third of the country's total biodiversity and 50% of the floral wealth (Takhtajan 1969). Arunachal Pradesh is the largest among the seven administrative states of northeastern India, covering an area of about 83,700km², of which 82% is under forest cover. It hosts a rich floral and faunal diversity due to its physiographic variation of elevations ranging 150–6,500 m and unique climatic conditions (Chakravarty et al. 2012; Taram et al. 2018). The vegetation ranges from tropical mixed and broadleaf evergreen forest to alpine meadow and scrubs (Singh & Singh 1991), and comprises many hill ranges such as Mishmee Hills, Abor Hills, Patkai Hills, Dafla Hills, and Aka Hills. The study area of East Siang is an important administrative district of Arunachal Pradesh and part of the erstwhile Siang frontier division. It also falls under the Abor Hills of the eastern Himalaya. In 1980, Siang frontier division was bifurcated into two districts namely West Siang and East Siang, headquartered at Along and Pasighat, respectively. In 1994, East Siang District was further divided into Upper Siang and East Siang, and in 2015 and 2017 Siang and Lower Siang districts were separated from East Siang. East Siang District is a mountainous area with river plains as well as rocky slopes and has a total area of 3,603km², lying approximately between 27.71 and 28.85 North latitudes and 94.70 and 95.58 East longitudes (Figure 1).

Griffith (1836), carried out the first floristic work in the Abor and Mishmi Hills, and published the 'Flora of Mishmi Hills'. With the advent of the 20th Century, many more plant explorations were carried out resulting in significant publications such as 'On the botany of Abor Expedition' by Burkill (1925); 'Botanical exploration in the Mishmi Hill' by Ward (1929), 'Observation of the Flora of Siang Frontier Division' by Rao & Joseph (1965), 'Floristic diversity assessment and vegetation analysis of Upper Siang District of Eastern Himalaya' by Choudhary (2008a,b), and Choudhary et al. (2012). Moreover, several studies on the taxonomy and ethnobotany of the plants of the region were also published (Tag et al. 2008; Singh et al. 2011; Yumnam et al. 2011; Yumnam & Tripathi 2013; Baruah et al. 2013; Boko & Narsimhan 2014; Mibang & Das 2017) which documented more

than 300 ethnobotanically useful plant species present in the Siang belt. Considering its floristic richness, a large chunk of this district was demarcated as the erstwhile Lali Reserve Forest, which was notified as Lali Wildlife Sanctuary in 1978 and subsequently renamed Daying Ering (D'Ering) Memorial Wildlife Sanctuary in 1986.

Although the general floristic account of the region was documented in several scattered publications, a comprehensive checklist for this important area highlighting the endemic and native species was not available. The recent discovery and documentation of several new species and new records (Jeyaprakash et al. 2014; Gogoi & Borah 2015; Tatum & Das 2016; Nangkar et al. 2017; Borah & Das 2018; Borah et al. 2018, 2020a,b; Taram & Borah 2020; Taram et al. 2020a,b,c,d,e) from the East Siang District and other districts of Arunachal Pradesh and its neighbouring regions, highlights the rich biodiversity of the area and warrants more exploration and research. The present study is, therefore, aimed to provide an updated checklist of the native flowering plants of East Siang District. It also provides information on the endemic and Red list plant species (IUCN 2020) distributed in the area, based on the field surveys and scrutiny of herbarium collections and published literature.

METHODS

Several field surveys of different localities of East Siang District were carried out from April 2016 to December 2019, covering most vegetation types and different elevations. The plant specimens collected were photographed in the field, press dried and mounted on herbarium sheets following standard methods (Jain & Rao 1977). Nomenclature was followed using online databases (POWO 2019; Tropicos 2020; WFO 2020) and classified as per Angiosperm Phylogeny Group system, APG IV (Chase et al. 2016). Identification was done using relevant literature (Hooker 1872; Kanjilal et al. 1934; Hajra et al. 1996; Chauhan et al. 1996; Giri et al. 2008; Chowdhery et al. 2009; Ambrish 2013) and scrutiny of the herbarium specimens housed in ASSAM, ARUN, HAU, North East Institute of Folk Medicine (NEIFM), Pasighat and several virtual herbaria like K, E, PE, etc. Information on the conservation status was noted following the International Union for Conservation of Nature (IUCN 2020) Red List. The voucher specimens were deposited in HAU (Herbarium of Rajiv Gandhi University, Arunachal Pradesh). The field number of the specimens scrutinized at the herbarium of North East Institute of Folk Medicine



Figure 1. East Siang District.

(NEIFM), Pasighat, Arunachal Pradesh, are prefixed with NEIFM in the Appendix 1.

RESULTS AND DISCUSSION

The structure and composition characteristics of the flora of East Siang District can be classified into two major climatic zones and five vegetation zones following the earlier classification pattern proposed by Champion & Seth (1968) and Kaul & Haridasan (1987) (Image 1; Table 1). These are discussed below:

Tropical forest

This type of vegetation can be mainly seen in the area of lower elevation such as Pasighat, Ruksin, Balek, Ledum, Magnang, Sille, Rani, Bilat, Mebo, Ngopok, Kiyit up to an altitude of 900m. This forest is characterized by tall trees with close canopy and receives heavy rainfall during monsoon season. The lower elevation areas of the district are occupied by tree species like *Gynocardia odorata*, *Liquidambar excelsa*, *Trevesia palmata*, *Garcinia pedunculata*, *Terminalia myriocarpa*, *Dillenia indica*, *Actinodaphne obovata*, *Cinnamomum bejolghota*, *Litsea glutinosa*, *Litsea monopetala*, *Duabanga grandifolia*, *Magnolia hodgsonii* etc. A large chunk of this forest is being invaded by exotic weeds like *Ageratum houstonianum*, *Chromolaena odorata*, *Cuscuta cassyoides*, *Mikania*

micrantha, and *Paederia foetida* (not included in the present checklist). Many common species of bamboos and orchids can also be found throughout these forests. The tropical forest of the district can be further divided into three subtypes:

A. Grasslands: The grasslands are found in the alluvial soils of the Siang River basin. Daying Ering Memorial Wildlife Sanctuary of this district occupies mostly this vegetation. Several scattered trees are also found in the areas, though the diversity is very less. *Carex baccans*, *Oplismenus burmanni*, *Erioscirpus comosus*, *Themeda villosa*, *Saccharum spontaneum*, *Saccharum arundinaceum*, *Phragmites karka*, and *Thysanolaena latifolia* are the most commonly found grasses and sedges as well as a few grassland orchids such as *Pachystoma pubescens* are also found.

B. Tropical semi-evergreen forest: This type of forest can be encountered in Sirki, Bodak, Ponging, and adjacent areas, which are dominated by trees like *Cordia dichotoma*, *Duabanga grandiflora*, *Ficus auriculata*, *F. crassiramea*, and *Toxicodendron hookerii*. In the next storey, *Maesa indica*, *Abroma augustum*, *Leea indica*, *Mussaenda glabra*, *Buddleja asiatica*, *Coffea bengalensis*, *Saurauia sinohirsuta*, *Sabia lanceolata*, and several species of wild *Citrus* can be commonly found. The ground storey is composed of herbs like *Viola betonicifolia*, *Lobelia nummularia*, *Persicaria capitata*, and *P. hydropiper*.



Image 1. Different habitat types of East Siang District: A—perennial waterfall at Sirki | B—evergreen forest along the Siang river basin | C—dense tropical forests at Pasighat | D—open tropical forests at Ruksin. © Dipankar Borah

Table 1. Major climatic and vegetation zones of East Siang District.

Climatic zone	Vegetation zone	Champion & Seth (1968)	Altitudinal range (in m)
1. Tropical	Grasslands	Unclassified	alluvial plains
	Tropical semi-evergreen	2/B/C1/1a, 2/B/C1b/IS1	near alluvial plains
	Tropical evergreen	1/B/C1, 1/B/C2	up to 600
	Tropical wet evergreen	8/B/C1	up to 900
2. Subtropical	Subtropical evergreen	3C3/Bb, 3C/IS2	900–1500

C. Tropical evergreen forest: This type of forest can be seen in most of the areas of the district such as Renging, Ledum, Mikong, and Pasighat. Due to heavy rainfall in the area, luxuriant growth of the tropical flora can be seen. The common tree species of this forest are *Actinodaphne obovata*, *Alstonia scholaris*, *Artocarpus lacucha*, *Callicarpa arborea*, *Canarium strictum*, *Litsea monopetala*, *Wallichia oblongifolia*, *Trevesia palmata*, *Rhus chinensis*, *Liquidambar excelsa*, *Morus macroura*, *Ficus semicordata*, *F. tinctoria*, *F. variegata*, and *F. virens*. The second storey, however, comprises of *Saurauia punduana* and *Litsea cubeba*. Epiphytic plants like *Aeschynanthus micranthus*, *A. acuminatus*, *A. superbus*,

Thunbergia coccinea, and *Dischidia bengalensis* are also common. The ground storey comprises of *Brachystemma calycinum*, *Hellenia speciosa*, *Phrynium pubinerve*, *Alpinia nigra*, *Curculigo capitulata*, and several others.

D. Tropical wet evergreen forest: These forests receive comparatively high rainfall (ca. 2,000mm or more) and the temperature ranges from 10–30 °C. These forests harbour the most diverse flora in the district comprising of three storeys. Ruksin, Rani, Sile, and Magnang host such forests where the elevation is very low compared to the other areas of the district. Being mostly plains, the regions are most prone to deforestation for agriculture, and hence large chunks of

such forests are under threat. A lot of tall tree species with close canopy can be seen growing luxuriantly in these areas. The commonest tree species of this storey are *Castanopsis indica*, *Chisocheton cumingianus*, *Toona hexandra*, *Aesculus assamica*, *Garcinia pedunculata*, *Balakata baccata*, *Gmelina arborea*, and *Bauhinia variegata*. Whereas the second storey comprises of small trees, lianas, and shrubs like *Saurauia napaulensis*, *S. armata*, *Fissistigma polyanthum*, *F. bicolor*, *Entada phaseoloides*, *Dalhousiea bracteata*, *Phlogacanthus curviflorus*, *Aralia armata*, *Caryota urens*, and *Calamus erectus*. The rich epiphytic flora can be seen holding the branches of the tree species in the area. Some of the common orchids of this forests are *Cymbidium aloifolium*, *Dendrobium aphyllum*, *D. nobile*, and *D. lasiopetalum*. The herbaceous flora of this area includes *Bonnaya antipoda*, *Lobelia zeylanica*, *Cynoglossum wallichii*, and several Begoniaceae, Balsaminaceae, & Zingiberaceae members.

Subtropical forest

This type of forest can be occasionally seen in the northern and eastern parts of the district, on high hilltops. Most of these forests lie at elevations of 900m and above. The elevational gradient plays a significant role in vegetation composition of the forest and a subtle change in the floristic composition can be observed in the areas. The ground storey is composed of various species of Urticaceae (e.g., *Elatostema*, *Pilea*, and *Boehmeria*), *Paris polyphylla*, *Wallichia triandra*, *Argostemma verticiliatum*, *Phlogacanthus vitellinus*, *Zeuxine flava*, *Cheirostylis parvifolia*, even *Ficus hederacea*, and *Henckelia mishmiensis* are seen growing luxuriantly on the rocky walls. The top storey comprises of sporadic distribution of *Phoenix rupicola*, *Pandanus furcatus*, and *Pterospermum lanceifolium*.

The present study documents a wide range of flowering plant diversity of East Siang District and altogether reports 503 species, one subspecies and four varieties, representing 348 genera belonging to 102 families (Appendix 1). The most species-rich families are Poaceae (27 species), Lamiaceae (23 species), Gesneriaceae (22 species), Rubiaceae (20 species), Fabaceae, Acanthaceae, Orchidaceae (19 species each), and Moraceae (18 species) (Appendix 1).

New records for the flora of Arunachal Pradesh

The present study documents two interesting species which were never recorded from Arunachal Pradesh. A brief taxonomic description of these species is given below to facilitate their easy identification:

Table 2. Ten dominant families of East Siang District.

	Family	No. of genera	No. of species
1	Poaceae	21	27
2	Lamiaceae	15	23
3	Gesneriaceae	08	22
4	Rubiaceae	17	20
5	Fabaceae	18	19
6	Orchidaceae	17	19
7	Acanthaceae	10	19
8	Moraceae	05	18
9	Urticaceae	09	15
10	Malvaceae	10	13

Mycetia mukerjiana Deb & Ratna Dutta (Image 2A–D)

Mycetia mukerjiana can be distinguished from its closely allied species *M. fangii* K.J.Yan & Z.Q.Song by its eglandulose calyx (vs. glandulose calyx), calyx lobes subequal to the corolla (vs. calyx lobes much shorter than the corolla), and longer bracts and bracteoles (Yan et al. 2016). *M. mukerjiana* was known so far from Assam, Mizoram, and Nagaland states of India and Bangladesh (Das & Rahman 2010; Chaturvedi et al. 2011; Barbhuiya et al. 2014), but never from Arunachal Pradesh. It was collected from Sirki area of East Siang District of Arunachal Pradesh during our floristic survey.

Specimens examined: (MT2075) (HAU), 06.vii.2018, 28.103N & 95.267E; 500m, India, Arunachal Pradesh, East Siang, Pasighat, Sirki.

Citrus indica Yu. Tanaka (Image 2E–I)

Citrus indica can be recognized from other *Citrus* species growing in the region by its unifoliate leaves, 5–7 pairs of prominent secondary veins, globose to obovoid fruits, depressed-obtuse at apex, deep orange to scarlet red when ripe, thin pericarp, as well as soft and thin mesocarp, 8–11 endocarp segments, polygonal pulps, sticky and yellow as well as 5–7 flattened, ovoid seeds per fruit. It was so far known from Assam, Manipur, Meghalaya and Nagaland (Borah et al. 2018b) but never from Arunachal Pradesh. During the present study, it was collected from Pasighat area of East Siang District.

Specimens examined: (MT2074) (HAU), 28.063N & E 95.324E; 180m, 24.iv.2019, India, Arunachal Pradesh, East Siang, Pasighat.

Endemism and range extension

Within the investigated regions, 11 species endemic to India were recorded. Out of which, five species are narrowly endemic to East Siang District of Arunachal



Image 2. New records for the flora of Arunachal Pradesh. (A–D)—*Mycetia mukerjiana* Deb & Ratna Dutta: A—leaf | B—inflorescence | C—flower top view | D—open flower showing stamens. (E–I)—*Citrus indica* Yu. Tanaka: E—leaf | F & G—flower | H—fruit | I—seed. © Dipankar Borah

Pradesh, and six endemic to the eastern Himalaya (Appendix 1). Among these, new distribution localities were recorded for two endemic species. Their details are as follows:

(1) *Hornstedtia arunachalensis* S. Tripathi & V. Prakash (Zingiberaceae) described from Papum Pare District of Arunachal Pradesh (Tripathi & Prakash 1999) was recorded during the present investigation from Sirki, East Siang District, extending its known range by 200km (Image 3E).

(2) *Henckelia mishmiensis* (Debb. ex Biswas) D.J. Middleton & Mich. Möller, earlier known only from Mishmi Hills of Arunachal Pradesh (Sinha & Dutta 2016), was recorded during the present investigation from Pasighat and Bodak of East Siang District, extending its known range by 100km (Image 3D).

Moreover, four species reported from only one or

two locations in Arunachal Pradesh from India and few other neighbouring countries were also recorded in the present study, extending their present known range to some extent. These are:

(1) *Lysionotus gamosepalus* W.T.Wang: Earlier reported only from Lohit and Upper Siang districts of Arunachal Pradesh from India (Akhil et al. 2019; Taram et al. 2020a), is also reported here from Pasighat of East Siang District (Image 3A). It is also distributed in China (POWO 2019).

(2) *Rhynchochotum parviflorum* Blume: So far known from Upper Siang District of Arunachal Pradesh (Taram et al. 2020d) and Andaman & Nicobar Islands in India (POWO 2019) is reported here from Sirki of East Siang District. Its distribution extends to New Guinea (POWO 2019).

(3) *Wallichia triandra* (J. Joseph) S.K.Basu: Earlier

Table 3. Rare and threatened plants recorded from the study area.

	Species name	Family	IUCN status
1	<i>Saurauia punduana</i> Wall.	Actinidiaceae	CR
2	<i>Larsenianthus arunachalensis</i> M. Sabu, Sanoj & Rajesh Kumar	Zingiberaceae	CR
3	<i>Piper pedicellatum</i> C.DC	Piperaceae	VU
4	<i>Phoenix rupicola</i> T.Anderson	Arecaceae	NT
5	<i>Zingiber zerumbet</i> (L.) Roscoe ex Sm.	Zingiberaceae	DD
6	<i>Amomum subulatum</i> Roxb.	Zingiberaceae	DD

reported only from Anjaw and Lohit District of Arunachal Pradesh (Henderson 2007) is reported here from Sirki of East Siang District. It is also distributed in China (POWO 2019).

(4) *Saurauia sinohirsuta* J.Q. Li & Soejarto: So far known from Upper Siang District of Arunachal Pradesh in India (Taram & Borah 2020) is reported here from Sirki of East Siang District. It is also distributed in China (POWO 2019).

Conservation

Out of the 508 taxa reported during this study, 108 species are designated under 'Least Concern', whereas most of them belong to 'Not Evaluated' category following IUCN 2020. Some rare and threatened plants recorded from the study area are listed in Table 3 along with their IUCN status.

Plants of medicinal and economic importance

During the present study it was found that 20 species (3.9%) are used to formulate different forms of traditional medicine, 126 species (24.8%) are edible, and 362 (71.3%) plants have no known uses (Appendix 1). Most of recorded usage show similarity to the earlier reports (Tag et al. 2008; Singh et al. 2011; Yumnam et al. 2011; Baruah et al. 2013; Yumnam & Tripathi 2013; Boko & Narsimhan 2014), except *Hornstedtia arunachalensis* S. Tripathi & V.Prakash, *Helixanthera parasitica* Lour., *Curculigo prainiana* (Deb) Bennet & Raizada, *Rhynchotechum obovatum* (Griff.) B.L.Burt, *Rhynchotechum parviflorum* Blume, *Brachystemma calycinum* D.Don and *Balanophora dioica* R. Br. ex Royle, which are new ethnobotanical records for the region.

Excluded taxa

Several studies conducted by Jeyaprakash K. (<https://orcid.org/0000-0001-8780-3487>) during 2015-2017 on the floristic diversity of East Siang District are not cited here complying with the journal's policy of curbing

publications in the predatory journals. These studies also included exotic, introduced and cultivated taxa to their list. We, however, have excluded them from the present checklist as our main aim was to document the native flora of the district. Moreover, *Champereia manillana* (Blume) Merr. reported by the same author as a new distributional record to Arunachal Pradesh was found to be an incorrect identification of *Lepionurus sylvestris* Blume, and hence, excluded from the present checklist.

CONCLUSION

There is still a gap in the research conducted so far to determine the approximate floral wealth of East Siang District. So far only one protected area, i.e., Daying Ering Memorial Wildlife Sanctuary has been designated in the district and much of the biodiversity finds its place in the private lands. We believe that our checklist will help in conservation planning of this sanctuary. The outburst of the human population, however, has created a need for new settlements as well as agricultural lands which pose serious threats to the present biodiversity in the area. Geographically, East Siang District is about 4% of Arunachal Pradesh, yet it represents around one-fourth of the state's flora. The tropical climate, along with its location in the eastern Himalaya biodiversity hotspot is the probable explanation for the high diversity. The present study should be considered a preliminary account of this floristically rich region, and more survey and research should be conducted to document its accurate floral wealth. Besides, the rich medicinal wealth of the district should also be conserved following several exercises such as Conservation Assessment and Management Prioritization (Ved et al. 2005), in association with the leading Indian institutions such as BSI, CIMAP, FRI, FRLHT, etc.

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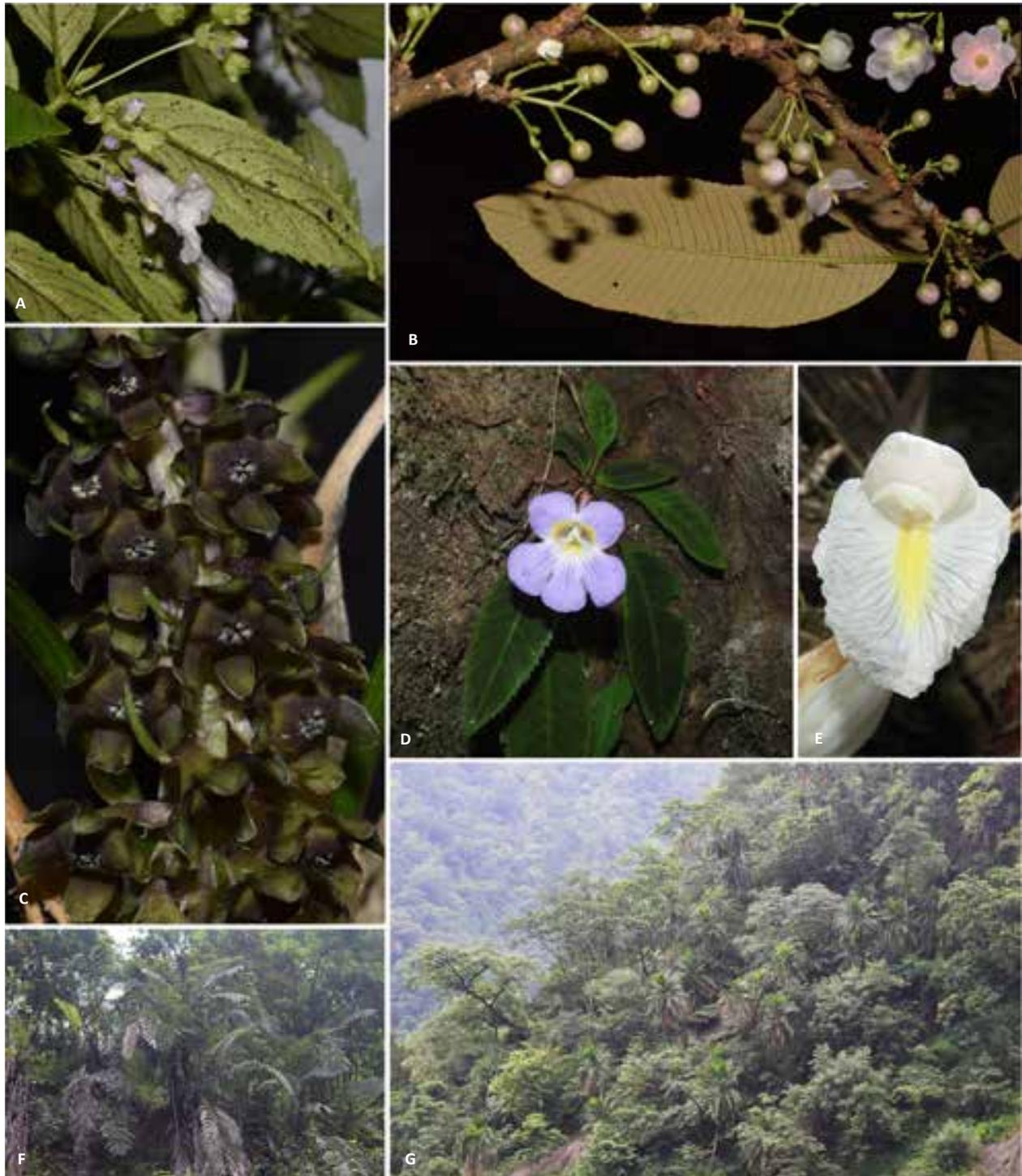


Image 3. Some endemic and endangered species found in East Siang: A—*Lysionotus gamosepalus* W.T. Wang | B—*Saurauia punduana* Wall. | C—*Peliosanthes ligniradicis* N. Tanaka, Taram & D. Borah | D—*Henckelia mishmiensis* (Debb. ex Biswas) D.J. Middleton & Mich.Möller | E—*Hornstedtia arunachalensis* S. Tripathi & V. Prakash | F—*Wallichia oblongifolia* Griff. | G—*Phoenix rupicola* T. Anderson. © A-C, F-G: Dipankar Borah; D-E: Momang Taram

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Appendix 1. Plant diversity of East Siang District.

Family	Taxon	Field no.	IUCN Status
Acanthaceae			
	<i>Codonacanthus pauciflorus</i> (Nees) Nees	MT2001	NE
	<i>Dicliptera babui</i> Karthik. & Moorthy	MT2500	NE
	<i>Justicia adhatoda</i> L.	MT2002	NE
	<i>Mackaya neesiana</i> (Wall.) Das	MT1742	NE
	<i>Phlogacanthus curviflorus</i> (Nees) Nees	MT2502	NE
	<i>Phlogacanthus gracilis</i> P.Anderson ex Burkill	NEIFM-306	E
	<i>Phlogacanthus thyrsoformis</i> (Roxb. ex Hardw.) Mabb.	MT2127	NE
	<i>Phlogacanthus vitellinus</i> (Roxb.) T.Anderson	MT2003	NE
	<i>Pseuderanthemum leptanthus</i> (C.B.Clarke) Lindau	MT2004	NE
	<i>Rhinacanthus calcaratus</i> (Wall.) Nees	MT2005	NE
	<i>Rungia pectinata</i> (L.) Nees	MT2501	NE
	<i>Strobilanthes hamiltoniana</i> (Steud.) Bosser & Heine	MT2128	NE
	<i>Strobilanthes mastersii</i> T.Anderson	MT2129	E
	<i>Strobilanthes oxycalycina</i> J.R.I. Wood	MT2007	E
	<i>Strobilanthes pauciflora</i> (Merr.) Y.F. Deng	MT2006, MT2101	NE
	<i>Strobilanthes secunda</i> T.Anderson	MT2008	NE
	<i>Strobilanthes tubiflos</i> (C.B.Clarke) J.R.I.Wood	MT2009	E
	<i>Thunbergia coccinea</i> Wall. ex D.Don	MT2010	NE
	<i>Thunbergia grandiflora</i> (Roxb. ex Rottler) Roxb.	MT2011	NE
Achariaceae			
	<i>Gynocardia odorata</i> R.Br.	MT1731	NE
Acoraceae			
	<i>Acorus calamus</i> L.	MT2533	LC
Actinidaceae			
	<i>Saurauia armata</i> Kurz #	MT1619	NE
	<i>Saurauia sinohirsuta</i> J.Q.Li & Soejarto	MT1829	NE
	<i>Saurauia napaulensis</i> DC. #	MT1590	LC
	<i>Saurauia punduana</i> Wall. #	MT1589	CR
Altingiaceae			
	<i>Liquidambar excelsa</i> (Noronha) Oken	MT1692	LC
Amaranthaceae			
	<i>Achyranthes aspera</i> L.	MT2134	NE
	<i>Achyranthes bidentata</i> Blume	MT2503	NE
	<i>Alternanthera sessilis</i> (L.) R.Br. ex DC.	MT2135	LC
	<i>Amaranthus viridis</i> L.	MT1564	NE
	<i>Chenopodium album</i> L. #	MT1700	NE
	<i>Chenopodium giganteum</i> D.Don. #	MT1617	NE
	<i>Cyathula prostrata</i> (L.) Blume	MT1235	NE
	<i>Deeringia amaranthoides</i> (Lam.) Merr. #	MT1747	NE

Family	Taxon	Field no.	IUCN Status
Amaryllidaceae			
	<i>Allium hookeri</i> Thwaites	MT1634	NE
Anacardiaceae			
	<i>Mangifera sylvatica</i> Roxb. #	MT1686	LC
	<i>Rhus chinensis</i> Mill. *	MT1580	LC
	<i>Spondias pinnata</i> (L.f) Kurz.	MT1530	NE
	<i>Choerospondias axillaris</i> (Roxb.) B.L.Burt & A.W.Hill #	MT2534	NE
	<i>Toxicodendron hookeri</i> (Sahni & Bahadur) C.Y.Wu & T.L.Ming	MT1626	NE
Annonaceae			
	<i>Fissistigma bicolor</i> (Roxb.) Merr.	MT1816	NE
	<i>Fissistigma polyanthum</i> (Hook.f and Thomson) Merr. #	MT1772	NE
	<i>Milium dioeca</i> (Roxb.) Chaowasku & Kessler	MT200	NE
	<i>Polyalthia suberosa</i> (Roxb.) Thwaites	NEIFM-394	NE
	<i>Trivalvaria costata</i> (Hook.f. & Thomson) I.M.Turner	NEIFM-513	NE
Apiaceae			
	<i>Centella asiatica</i> (L.) Urb *	MT1711	LC
	<i>Oenanthe javanica</i> (Blume) DC.	MT1821	LC
Apocynaceae			
	<i>Aganosma cymosa</i> (Roxb.) G.Don	NEIFM-483	NE
	<i>Alstonia scholaris</i> (L.) R.Br.	MT2133	LC
	<i>Beaumontia grandiflora</i> Wall.	MT1636	NE
	<i>Dischidia bengalensis</i> Colebr.	MT2134	NE
	<i>Hemidesmus indicus</i> (L.) R.Br.	MT2135	NE
	<i>Hoya arnottiana</i> Wight	NEIFM-215	NE
	<i>Hoya verticillata</i> (Vahl) G.Don	MT2012	NE
	<i>Rauvolfia verticillata</i> (Lour.) Baill.	NEIFM-353, NEIFM-486	NE
	<i>Tabernaemontana divaricata</i> (L.) R.Br. ex Roem. & Schult.	MT2136	NE
	<i>Wrightia coccinea</i> (Roxb. ex Hornem.) Sims	NEIFM-247	NE
Araceae			
	<i>Alocasia fornicata</i> (Kunth) Schott	MT2014	LC
	<i>Amorphophallus bulbifer</i> (Roxb.) Blume	NEIFM-560	NE
	<i>Amorphophallus napalensis</i> (Wall.) Bogner & Mayo	NEIFM-551	NE
	<i>Arisaema arunachalensis</i> A.Nangkar, A.P. Das & H.Tag	cf. Nangkar et al. 2017	E
	<i>Colocasia fallax</i> Schott	MT2013	LC
	<i>Homalomena aromatica</i> (Spreng.) Schott	MT2535	NE
	<i>Pothos scandens</i> L. *	MT1722	NE
	<i>Pothos chinensis</i> (Raf.) Merr.	NEIFM-213	NE
	<i>Rhaphidophora decursiva</i> (Roxb.) Schott	MT1571	NE
	<i>Rhaphidophora glauca</i> (Wall.) Schott	MT2130	NE
	<i>Rhaphidophora hookeri</i> Schott	MT1572	NE
	<i>Stuednera assamica</i> Hook.f.	MT2131	E

Family	Taxon	Field no.	IUCN Status
Araliaceae			
	<i>Aralia armata</i> (Wall ex. Don) Seem. #	MT1552	LC
	<i>Brassaiopsis glomerulata</i> (Blume) Regel	MT1579	LC
	<i>Eleutherococcus trifoliatus</i> (L.) S.Y.Hu	NEIFM-307	NE
	<i>Heteropanax fragrans</i> (Roxb.) Seem.	MT1656	NE
	<i>Hydrocotyle himalaica</i> P.K.Mukh.	MT1830	NE
	<i>Hydrocotyle sibthorpioides</i> Lam.	MT2504	LC
	<i>Hydrocotyle javanica</i> Thunb.	MT1712	NE
	<i>Schefflera bengalensis</i> Gamble	NEIFM-207	NE
	<i>Trevesia palmata</i> (Roxb. ex Lindl.) Vis. #	MT1679	LC
Arecaceae			
	<i>Calamus erectus</i> Roxb. #	MT1562	NE
	<i>Calamus flagellum</i> Griff. Ex Walp #	MT1541	NE
	<i>Caryota urens</i> L.	MT1570	LC
	<i>Phoenix rupicola</i> T.Anderson	DB2015	E, NT
	<i>Pinanga gracilis</i> Blume	MT2016	NE
	<i>Wallichia oblongifolia</i> Griff.	MT1538	NE
	<i>Wallichia triandra</i> (J.Joseph) S.K.Basu	MT1537	LC
Aristolochiaceae			
	<i>Aristolochia platanifolia</i> (Klotzsch) Duch	DB2152	NE
Asparagaceae			
	<i>Dracaena angustifolia</i> (Medik.) Roxb.	MT2407	NE
	<i>Dracaena petiolata</i> Hook.f.	MT2409	E
	<i>Peliosanthes ligniradicis</i> N.Tanaka, Taram & D. Borah	MT&DB 651	E
	<i>Peliosanthes macrophylla</i> Wall. ex Baker	MT2411	NE
	<i>Tupistra stoliczana</i> Kurz	MT2412	NE
Asteraceae			
	<i>Artemisia indica</i> Willd *	MT1646	NE
	<i>Blumea balsamifera</i> (L.) DC. *	MT1655	NE
	<i>Gnaphalium polycaulon</i> Pers. #	MT1758	NE
	<i>Grangea maderaspatana</i> (L.) Poir.	NEIFM-91	NE
	<i>Gynura cusimbua</i> (D.Don) S.Moore #	MT1743	LC
	<i>Laggera crispata</i> (Vahl) Hepper & J.R.I.Wood	MT2407	NE
	<i>Pseudognaphalium affine</i> (D.Don) Anderb. #	MT1757	NE
	<i>Youngia japonica</i> (L.) DC. #	MT1536	NE
Balanophoraceae			
	<i>Balanophora dioica</i> R. Br. ex Royle #	MT1558	NE
Balsaminaceae			
	<i>Impatiens arguta</i> Hook f. & Thomson	MT2019	NE
	<i>Impatiens latiflora</i> Hook.f. & Thomson	MT2018	NE
	<i>Impatiens porrecta</i> Wall. ex Hook.f. & Thomson	MT2017	NE
	<i>Impatiens siangensis</i> Gogoi	MT2417	E

Family	Taxon	Field no.	IUCN Status
Begoniaceae			
	<i>Begonia aborensis</i> Dunn #	MT1595	NE
	<i>Begonia acetosella</i> Craib. #	MT1638	NE
	<i>Begonia annulata</i> K.Koch	MT2019	NE
	<i>Begonia biserrata</i> Lindl. #	MT1639	NE
	<i>Begonia burkillii</i> Dunn	MT2020	NE
	<i>Begonia josephi</i> A.DC	MT2021	NE
	<i>Begonia roxburghii</i> (Miq.) A.DC #	MT1594	NE
	<i>Begonia silletensis</i> (A.DC.) C.B. Clarke #	MT2022	NE
	<i>Begonia xanthina</i> Hook.	MT2024	NE
Bignoniaceae			
	<i>Oroxylum indicum</i> (L.) Kurz	MT2406	NE
	<i>Radermachera gigantea</i> (Blume) Miq.	NEIFM-535	LC
	<i>Stereospermum chelonoides</i> (L.f.) DC	MT2025	NE
Boraginaceae			
	<i>Bothriospermum zeylanicum</i> (J.Jacq.) Druce	NEIFM-254	NE
	<i>Cordia dichotoma</i> G.Forst	MT1514	LC
	<i>Cynoglossum wallichii</i> G.Don	MT2405	NE
	<i>Ehretia acuminata</i> R.Br.	NEIFM-109, NEIFM-185	LC
	<i>Ehretia wallichiana</i> Hook.f. & Thomson ex C.B.Clarke	DB2023	NE
	<i>Rotula aquatica</i> Lour.	NEIFM-413	NE
Brassicaceae			
	<i>Cardamine hirsuta</i> L. #	MT1751	NE
	<i>Rorippa dubia</i> (Pers.) H.Hara #	MT1753	NE
Burseraceae			
	<i>Canarium strictum</i> Roxb. #	MT1687	NE
Campanulaceae			
	<i>Lobelia nummularia</i> Lam.	MT2403	NE
	<i>Lobelia zeylanica</i> L.	MT2404	LC
Capparaceae			
	<i>Capparis acutifolia</i> Sweet subsp. <i>sabiifolia</i> (J. D. Hooker & Thomson) Jacobs	MT2026	NE
	<i>Capparis assamica</i> Hook.f. & Thomson	MT2027	E
	<i>Capparis multiflora</i> Hook.f. & Thomson	MT1776	NE
	<i>Crateva magna</i> (Lour.) DC.	NEIFM-543	NE
	<i>Stixis suaveolens</i> (Roxb.) Baill. #	MT1613	NE
Caryophyllaceae			
	<i>Brachystemma calycinum</i> D.Don *	MT1746	NE
	<i>Stellaria media</i> (L.) Vill #	MT1694	LC
Celastraceae			
	<i>Microtropis discolor</i> (Wall.) Wall. ex Meisn.	MT2412	NE
	<i>Loeseneriella pauciflora</i> (DC.) A.C.Sm.	NEIFM-357	NE

Family	Taxon	Field no.	IUCN Status
Chloranthaceae			
	<i>Chloranthus elatior</i> Link	NEIFM-322	NE
Clusiaceae			
	<i>Garcinia anomala</i> Planch. & Triana #	MT2411	NE
	<i>Garcinia lanceifolia</i> Roxb. #	MT1560	NE
	<i>Garcinia pedunculata</i> Roxb. ex Buch.- Ham *	MT1586	NE
	<i>Garcinia xanthochymus</i> Hook.f. ex T.Anderson	NEIFM-487	NE
Colchicaceae			
	<i>Disporum longistylum</i> (H.Lév. & Vaniot) H.Hara	NEIFM-484	NE
Combretaceae			
	<i>Terminalia chebula</i> Retz. #	MT3003	NE
	<i>Terminalia myriocarpa</i> Van Heurck & Mull.Arg	MT3004	NE
Commelinaceae			
	<i>Amischotolype hookeri</i> (Hassk.) H.Hara	MT2505	NE
	<i>Commelina benghalensis</i> L	MT2401	LC
	<i>Floscopa scandens</i> Lour.	MT2402	LC
	<i>Murdannia nudiflora</i> (L.) Brenan	MT1693	NE
	<i>Rhopalephora scaberrima</i> (Blume) Faden	MT2400	NE
Convolvulaceae			
	<i>Argyreia argentea</i> (Roxb.) Sweet	NEIFM-407	NE
	<i>Argyreia nervosa</i> (Burm.f.) Bojer	MT2029	NE
Cornaceae			
	<i>Alangium chinense</i> (Lour.) Harms	MT2507	NE
Costaceae			
	<i>Hellenia speciosa</i> (J.Koenig) S.R.Dutta	MT2028	NE
Cucurbitaceae			
	<i>Cucumis maderaspatanus</i> L.	NEIFM-365	NE
	<i>Hodgsonia macrocarpa</i> (Blume) Cong. #	MT1552	NE
	<i>Solena heterophylla</i> Lour. *	MT2538	NE
	<i>Thladiantha cordifolia</i> (Blume) Cong. #	MT1738	NE
	<i>Trichosanthes nervifolia</i> L.	MT2299	NE
Cyperaceae			
	<i>Carex baccans</i> Nees	MT1677	LC
	<i>Cyperus digitatus</i> Roxb.	MT2536	LC
	<i>Cyperus rotundus</i> L.	MT2537	LC
	<i>Cyperus distans</i> L.f.	NEIFM-552	LC
	<i>Cyperus mindorensis</i> (Steud.) Huygh	NEIFM-129	NE
	<i>Cyperus pilosus</i> Vahl	NEIFM-555	NE
	<i>Eleocharis geniculata</i> (L.) Roem. & Schult.	MT2298	LC
	<i>Erioscirpus comosus</i> (Wall.) Palla	MT2291	NE
	<i>Fimbristylis bisumbellata</i> (Forssk.) Bubani	MT2288	LC
	<i>Fimbristylis dichotoma</i> (L.) Vahl	NEIFM-073, NEIFM-297	

Family	Taxon	Field no.	IUCN Status
Dilleniaceae			
	<i>Dillenia indica</i> L. #	MT1593	NE
Dioscoreaceae			
	<i>Dioscorea alata</i> L. #	MT1778	NE
	<i>Dioscorea bulbifera</i> L. #	MT1652	NE
	<i>Dioscorea esculenta</i> (Lour.) Burkill #	MT1832	NE
	<i>Dioscorea pentaphylla</i> L. #	MT1544	NE
	<i>Tacca integrifolia</i> Ker Gawl.	MT2030	NE
Ericaceae			
	<i>Agapetes bhutanica</i> N.P.Balacr. & Sud.Chowdhury	NEIFM-528	NE
	<i>Agapetes macrantha</i> var. <i>grandiflora</i> (Hook.f.) D.Banik and Sanjappa *	MT1502	NE
	<i>Agapetes serpens</i> (Wight) Sleumer	MT2031	NE
Euphorbiaceae			
	<i>Balakata baccata</i> (Roxb.) Esser	MT22188	LC
	<i>Bridelia montana</i> (Roxb.) Willd.	NEIFM-232	NE
	<i>Croton caudatus</i> Geiseler	MT2508	NE
	<i>Homonioia riparia</i> Lour.	NEIFM-100	LC
	<i>Mallotus paniculatus</i> (Lam.) Müll.Arg.	MT22187	LC
	<i>Mallotus tetracoccus</i> (Roxb.) Kurz	MT2509	NE
	<i>Ostodes paniculata</i> Blume	MT1556	LC
Fabaceae			
	<i>Albizia odoratissima</i> (L.f.) Benth	MT1550	LC
	<i>Archidendron chevalieri</i> (Kosterm.) I.C. Nielsen	NEIFM-478	NE
	<i>Bauhinia purpurea</i> L.#	MT2178	LC
	<i>Bauhinia variegata</i> L.#	MT1745	LC
	<i>Crotalaria spectabilis</i> Roth	MT2177	NE
	<i>Dalbergia rimosa</i> Roxb.	NEIFM-355	LC
	<i>Dalbergia sissoo</i> Roxb. ex DC.	MT2176	NE
	<i>Dalhousiea bracteata</i> (Roxb.) Graham ex Benth	MT2175	NE
	<i>Entada phaseoloides</i> (L.) Merr.	MT1773	NE
	<i>Erythrina variegata</i> L.	MT2173	LC
	<i>Gymnocladus burmanicus</i> C.E.Parkinson	cf. Singh et al. 2009	NE
	<i>Leptodesmia microphylla</i> (Thunb.) H.Ohashi & K.Ohashi	NEIFM-265, NEIFM-177	NE
	<i>Mastersia assamica</i> Benth.	MT2177	NE
	<i>Mucuna macrocarpa</i> Wall.	MT2510	NE
	<i>Ohwia caudata</i> (Thunb.) H.Ohashi	NEIFM-411	NE
	<i>Ototropis multiflora</i> (DC.) H.Ohashi & K.Ohashi	MT2178	NE
	<i>Pueraria montana</i> (Lour.) Merr. *	MT1775	NE
	<i>Senegalia catechu</i> (L.f.) P.J.H.Hurter & Mabb.	NEIFM-080, NEIFM-252	NE
	<i>Senegalia rugata</i> (Lam.) Britton & Rose	MT1501	NE
	<i>Tephrosia candida</i> DC	MT2171	NE

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Fagaceae			
	<i>Castanopsis indica</i> (Roxb. Ex Lindl.) A.DC.#	MT1602	LC
	<i>Castanopsis purpurella</i> (Miq.) N.P.Balakr. #	MT1618	NE
Gesneriaceae			
	<i>Aeschynanthus acuminatus</i> Wall. ex A.DC	MT2042	NE
	<i>Aeschynanthus gracilis</i> C.S.P.Paris ex C.B Clarke	MT1505	NE
	<i>Aeschynanthus micranthus</i> C.B. Clarke	MT1504	NE
	<i>Aeschynanthus monetarius</i> Dunn	MT1811	NE
	<i>Aeschynanthus parasiticus</i> C.B. Clarke	MT1503	NE
	<i>Aeschynanthus superbus</i> C.B.Clarke	MT2041	NE
	<i>Boeica filiformis</i> C.B.Clarke	MT2040	NE
	<i>Boeica fulva</i> C.B. Clarke #	MT1704	NE
	<i>Epithema carnosum</i> Benth	MT2039	NE
	<i>Henckelia siangensis</i> Taram, D.Borah & Tag	MT2300	E
	<i>Henckelia grandifolia</i> A.Dietr.	MT2038	NE
	<i>Henckelia mishmiensis</i> (Debb. ex Biswas) D.J.Middleton & Mich.Möller	MT2037, MT2172	NE
	<i>Henckelia oblongifolia</i> (Roxb.) D.J.Middleton & Mich. Möller	MT2170	NE
	<i>Henckelia pumila</i> (D.Don) A.Dietr	MT2036	NE
	<i>Lysionotus gamosepalus</i> W.T.Wang var. <i>gamosepalous</i>	MT2035	NE
	<i>Lysionotus serratus</i> D.Don	MT2034	NE
	<i>Rhynchotechum ellipticum</i> (Wall. ex D.Dietr.) A.DC. #	MT1705	NE
	<i>Rhynchotechum obovatum</i> (Griff.) B.L.Burt #	MT22167	NE
	<i>Rhynchotechum parviflorum</i> Blume #	MT1814, MT22168	NE
	<i>Rhynchotechum vestitum</i> Wall. ex C. B. Clarke #	MT1706, MT22169	NE
	<i>Stauranthera grandifolia</i> Benth.	MT2033	NE
	<i>Tetraphylloides bengalensis</i> (C.B.Clarke) Doweld	MT2032	NE
Gentianaceae			
	<i>Exacum teres</i> Wall.	MT2166	NE
Hydrangeaceae			
	<i>Hydrangea febrifuga</i> (Lour.) Y. De Smet & Granados	MT2043	NE
Hydroleaceae			
	<i>Hydrolea zeylanica</i> (L.) Vahl	MT2165	LC
Hypoxidaceae			
	<i>Curculigo capitulata</i> (Lour.) Kuntze *	MT1815	NE
	<i>Curculigo prainiana</i> (Deb) Bennet & Raizada (SE, Ethnomedicine)	MT2133	NE
Lamiaceae			
	<i>Achyropermum densiflorum</i> Blume	NEIFM-401	NE
	<i>Anisomeles indica</i> (L.) Kuntze	MT1894	NE
	<i>Callicarpa arborea</i> Roxb. #	MT2164	LC
	<i>Callicarpa macrophylla</i> Vahl	MT1517	LC
	<i>Clerodendrum chinense</i> (Osbeck) Mabb.	MT2512	LC
	<i>Clerodendrum colebrookeanum</i> Walp. #	MT2163	NE

Family	Taxon	Field no.	IUCN Status
	<i>Clerodendrum indicum</i> (L.) Kuntze	NEIFM-471	NE
	<i>Clerodendrum japonicum</i> (Thunb.) Sweet	MT2162	LC
	<i>Clerodendrum laevifolium</i> Blume	MT2044	NE
	<i>Elsholtzia ciliata</i> (Thunb.) Hyl.	NEIFM-359	NE
	<i>Gmelina arborea</i> Roxb. ex Sm	MT1545	LC
	<i>Isodon coetsa</i> (Buch.-Ham. ex D.Don) Kudô	NEIFM-240	NE
	<i>Leonurus japonicus</i> Houtt.	NEIFM-452	NE
	<i>Leucas chinensis</i> (Retz.) Sm.	NEIFM-065	NE
	<i>Leucas zeylanica</i> (L.) W.T.Aiton	NEIFM-296	NE
	<i>Leucosceptrum canum</i> Sm.	MT1733	NE
	<i>Perilla frutescens</i> (L.) Britt. #	MT2161	NE
	<i>Pogostemon brachystachyus</i> Benth.	NEIFM-312	NE
	<i>Pogostemon elsholtzioides</i> Benth.	NEIFM-148	NE
	<i>Pogostemon plectranthoides</i> Desf.	NEIFM-382	NE
	<i>Rothea serrata</i> (L.) Steane & Mabb.	MT 2511	NE
	<i>Tectona grandis</i> L.f	MT2160	NE
	<i>Teucrium viscidum</i> Blume	NEIFM-319, NEIFM-490	NE
Lauraceae			
	<i>Actinodaphne obovata</i> (Nees) Blume	MT1690	NE
	<i>Beilschmiedia assamica</i> Meisn.	NEIFM-541	NE
	<i>Cinnamomum bejolghota</i> (Bucc-Ham) Sweet #	MT1777	LC
	<i>Lindera communis</i> Hemsl.	NEIFM-443	NE
	<i>Litsea cubeba</i> (Lours.) Pers. #	MT2159	NE
	<i>Litsea glutinosa</i> (Lour.) C.B.Rob.	MT2540	LC
	<i>Litsea monopetala</i> (Roxb.) Pers.	MT1563	LC
	<i>Phoebe cooperiana</i> P.C.Kanjilal and Das #	MT2158	E
Linderniaceae			
	<i>Bonnaya antipoda</i> (L.) Druce	MT2157	NE
	<i>Bonnaya ciliata</i> (Colsm.) Spreng.	MT2156	NE
	<i>Torenia bicolor</i> Dalzell	MT2154	LC
	<i>Torenia crustacea</i> (L.) Cham. & Schuldt	MT2153	LC
	<i>Torenia fournieri</i> Linden ex E.Fourn.	MT2541	NE
Loranthaceae			
	<i>Helixanthera parasitica</i> Lour. #	MT2046	NE
Lythraceae			
	<i>Duabanga grandiflora</i> (Roxb. Ex DC) Walp.	MT2045	LC
	<i>Rotala rotundifolia</i> (Buch.-Ham. ex Roxb.) Koehne	MT20542	NE
Magnoliaceae			
	<i>Magnolia hodgsonii</i> (Hook.f. & Thomson) H.Keng	MT2047	LC
Malvaceae			
	<i>Abroma augustum</i> (L.) L.f.	MT1699	NE
	<i>Ayenia grandifolia</i> (DC.) Christenh. & Byng	NEIFM-216, NEIFM-463	NE
	<i>Bombax ceiba</i> L.	MT2050	NE

Family	Taxon	Field no.	IUCN Status
	<i>Grewia asiatica</i> L.	NEIFM-351	LC
	<i>Kydia calycina</i> Roxb.	MT1691	LC
	<i>Melochia corchorifolia</i> L.	NEIFM-270, NEIFM-325	NE
	<i>Pterospermum acerifolium</i> (L.) Willd	MT2049	LC
	<i>Pterospermum lanceifolium</i> Roxb. ex DC.	MT2048	NE
	<i>Sida acuta</i> Burm.f.	MT2054	NE
	<i>Sterculia lanceolata</i> var. <i>coccinea</i> (Jack) Phengklai #	MT1786	LC
	<i>Sterculia striatiflora</i> Mast. #	MT&DB 0206	NE
	<i>Sterculia villosa</i> Roxb. ex Sm	MT2052	NE
	<i>Urena lobata</i> L.	MT2051	NE
Marantaceae			
	<i>Phrynium pubinerve</i> Blume #	MT2050	NE
Mazaceae			
	<i>Mazus pumilus</i> (Burm.f.) Steenis	MT2150	NE
	<i>Mazus surculosus</i> D.Don	MT2513	NE
Melanthiaceae			
	<i>Paris polyphylla</i> Sm. *	MT2051	NE
Melastomataceae			
	<i>Melastoma malabathricum</i> L #	MT2149	NE
	<i>Osbeckia nepalensis</i> Hook.	MT2148	NE
	<i>Osbeckia nutans</i> Wall.	MT2147	NE
	<i>Oxyspora paniculata</i> DC.	MT2054	NE
	<i>Pseudodissochaeta assamica</i> (C.B.Clarke) Nayar	MT2052	NE
	<i>Sarcopyramis napalensis</i> Wall.	DB2053	NE
Meliaceae			
	<i>Chisocheton cumingianus</i> (C.DC.) Harms	DB2055	LC
	<i>Dysoxylum alliaceum</i> (Blume) Blume	NEIFM-488	LC
	<i>Melia azedarach</i> L.	MT2126	LC
	<i>Toona hexandra</i> (Wall.) M.Roem	MT2056	NE
Menispermaceae			
	<i>Stephania japonica</i> (Thunb.) Miers	MT2057	NE
	<i>Stephania rotunda</i> Lour.	MT2058	NE
	<i>Tinospora cordifolia</i> (Willd.) Hook.f. & Thomson	MT2514	NE
Molluginaceae			
	<i>Trigastrotheca pentaphylla</i> (L.) Thulin	NEIFM-345	NE
Moraceae			
	<i>Artocarpus heterophyllus</i> Lam. #	MT1779	NE
	<i>Artocarpus lacucha</i> Buch.Ham. #	MT1756	NE
	<i>Broussonetia papyrifera</i> (L.) L'Hér. ex Vent.	NEIFM-228	LC
	<i>Ficus auriculata</i> Lour. #	MT1601	LC
	<i>Ficus crassiramea</i> (Miq.) Miq.	MT1641	NE
	<i>Ficus drupacea</i> Thunb.	NEIFM-545	LC

Family	Taxon	Field no.	IUCN Status
	<i>Ficus hederacea</i> Roxb.	MT1790	NE
	<i>Ficus heteropleura</i> Blume	MT1764	NE
	<i>Ficus hispida</i> L.f.#	MT2124	LC
	<i>Ficus oligodon</i> Miq. #	MT1600	LC
	<i>Ficus religiosa</i> L.	MT1574	NE
	<i>Ficus semicordata</i> Buch-Ham ex Sm.#	MT1575	LC
	<i>Ficus simplicissima</i> Lour.	MT1599	NE
	<i>Ficus tinctoria</i> G.Forst.	MT1588	LC
	<i>Ficus variegata</i> Blume	MT1833	LC
	<i>Ficus virens</i> Aiton #	MT1808	LC
	<i>Maclura cochinchinensis</i> (Lour.) Corner #	MT1647	NE
	<i>Morus macrourea</i> Miq.	MT2123	NE
Musaceae			
	<i>Musa aurantiaca</i> G.Mann ex Baker #	MT1726	LC
	<i>Musa balbisiana</i> Colla #	MT1760	LC
	<i>Musa sanguinea</i> Hook.f. *	MT2069	LC
Myricaceae			
	<i>Myrica esculenta</i> Buch.Ham Ex D.Don #	MT2121	NE
Myrtaceae			
	<i>Syzygium formosum</i> (Wall.) Mason #	MT1826	NE
	<i>Syzygium fruticosum</i> DC. #	MT1828	NE
	<i>Syzygium aqueum</i> (Burm.f.) Alston #	MT2068	NE
	<i>Syzygium cumini</i> (L.) Skeels #	MT2120	LC
Nyctaginaceae			
	<i>Boerhavia diffusa</i> L.	NEIFM-537	NE
Olacaceae			
	<i>Erythralium scandens</i> Blume	MT2517	LC
Oleaceae			
	<i>Jasminum pentaneurum</i> Hand.-Mazz.	NEIFM-445	NE
	<i>Jasminum laurifolium</i> Roxb. ex Hornem. var. <i>laurifolium</i>	MT2067	NE
Orchidaceae			
	<i>Arundina graminifolia</i> (D.Don) Hochr	DB2137	NE
	<i>Bulbophyllum odoratissimum</i> (Sm.) Lindl. ex Wall.	MT2066	NE
	<i>Cheirostylis parvifolia</i> Lindl.	DB2138	NE
	<i>Coelogyne fimbriata</i> Lindl. var. <i>fimbriata</i>	MT2065	NE
	<i>Corymborkis veratrifolia</i> (Reinw.) Blum	DB2139	NE
	<i>Cymbidium aloifolium</i> (L.) Sw.	MT2064	NE
	<i>Dendrobium aphyllum</i> (Roxb.) C.E.C.Fisch.	MT2140	LC
	<i>Dendrobium chrysanthum</i> Wall. ex Lindl.	MT2141	NE
	<i>Dendrobium nobile</i> Lindl.	MT2142	NE
	<i>Dendrolirium lasiopetalum</i> (Willd.) S.C.Chen & J.J.Wood	MT2144	NE
	<i>Eulophia dabia</i> (D.Don) Hochr.	NEIFM-102, NEIFM-234	NE

Family	Taxon	Field no.	IUCN Status
	<i>Geodorum densiflorum</i> (Lam.) Schltr.	NEIFM-191	NE
	<i>Goodyera procera</i> (Ker Gawl.) Hook.	MT2063	NE
	<i>Nervilia macroglossa</i> (Hook.f.) Schltr.	NEIFM-214	NE
	<i>Pachystoma pubescens</i> Blume	DB2143	NE
	<i>Pholidota pallida</i> Lindl	MT2062	NE
	<i>Rhynchostylis retusa</i> (L.) Blume	MT2061	NE
	<i>Vanda bicolor</i> Griff.	MT2060	NE
	<i>Zeuxine flava</i> (Wall. ex Lindl.) Trimen	DB2059	NE
Orobanchaceae			
	<i>Lindenbergia hookeri</i> C.B.Clarke ex Hook.f. #	MT2119	NE
	<i>Aeginetia indica</i> L.	NEIFM-493	NE
Pandanaceae			
	<i>Pandanus furcatus</i> Roxb. #	MT1788	NE
Phyllanthaceae			
	<i>Actephila excelsa</i> (Dalzell) Mull.Arg	MT1508	NE
	<i>Baccaurea ramiflora</i> Lour. #	MT1629	NE
	<i>Bischofia javanica</i> Blume	NEIFM-495	LC
	<i>Breynia androgyna</i> (L.) Chakrab. & N.P.Balakr.#	MT2118	NE
	<i>Leptopus clarkei</i> (Hook.f.) Pojark.	NEIFM-387	NE
	<i>Phyllanthus assamicus</i> Müll.Arg.	NEIFM-121	LC
	<i>Phyllanthus fraternus</i> G.L.Webster	NEIFM-389, NEIFM-431	NE
	<i>Phyllanthus lanceolarius</i> (Roxb.) Müll.Arg.	NEIFM-298	NE
	<i>Phyllanthus reticulatus</i> Poir.	MT2070	NE
Piperaceae			
	<i>Piper attenuatum</i> Buch.-Ham. ex Miq.	NEIFM-174, NEIFM-362	NE
	<i>Piper griffithii</i> C.DC.	NEIFM-334	NE
	<i>Piper mullesua</i> Buch.-Ham. ex D.Don	MT1609	NE
	<i>Piper pedicellatum</i> C.DC #	MT2071	E,VU
	<i>Piper sylvaticum</i> Roxb.	NEIFM-435, NEIFM-386	NE
Plantaginaceae			
	<i>Plantago asiatica</i> L. #	MT1631	NE
Poaceae			
	<i>Arundinella nepalensis</i> Trin.	NEIFM-361	NE
	<i>Bambusa tulda</i> Roxb. *	MT2117	NE
	<i>Chrysopogon aciculatus</i> (Retz.) Trin	MT2555	NE
	<i>Cynodon dactylon</i> (L.) Pers.	MT1653	NE
	<i>Dendrocalamus giganteus</i> Munro *	MT1643	LC
	<i>Dendrocalamus hamiltonii</i> Nees and Arn.ex Munro *	MT1581	NE
	<i>Digitaria abludens</i> (Roem. & Schult.) Veldkamp	NEIFM-130, NEIFM-131	NE
	<i>Digitaria ciliaris</i> (Retz.) Koeler	NEIFM-133	NE
	<i>Dinochloa maccllellandii</i> (Munro) Kurz	MT2072	NE
	<i>Echinochloa colonum</i> (L.) Link	MT2553	LC

Family	Taxon	Field no.	IUCN Status
	<i>Echinochloa crus-galli</i> (L.) P. Beauv.	NEIFM-098	LC
	<i>Eragrostis uniolooides</i> (Retz.) Nees ex Steud.	NEIFM-344	LC
	<i>Eulalia hirtifolia</i> (Hack.) Kuntze	NEIFM-410	NE
	<i>Hemarthria compressa</i> (L.f.) R.Br.	NEIFM-291	LC
	<i>Oplismenus burmanni</i> (Retz.) P. Beauv.	MT2550	NE
	<i>Oplismenus compositus</i> (L.) P. Beauv.	MT 2551	NE
	<i>Phragmites karka</i> (Retz.) Trin. ex Steud.	MT1585	LC
	<i>Phyllostachys manii</i> Gamble	MT1566	NE
	<i>Pseudoraphis minuta</i> (Mez) Pilg.	NEIFM-290	NE
	<i>Saccharum arundinaceum</i> Retz. #	MT1782	NE
	<i>Saccharum spontaneum</i> L. #	MT1584	LC
	<i>Sacciolepis indica</i> (L.) Chase	NEIFM-161	NE
	<i>Setaria palmifolia</i> (J.Koenig) Stapf	MT2554	NE
	<i>Setaria pumila</i> (Poir.) Roem. & Schult.	MT2552	NE
	<i>Stapletonia seshagiriana</i> (R.B. Majumdar) H.B. Naithani	MT1820	NE
	<i>Themeda villosa</i> (Lam.) A. Camas	MT1708	NE
	<i>Thysanolaena latifolia</i> (Roxb. ex Hornem.) Honda	MT2116	NE
Polygonaceae			
	<i>Koenigia mollis</i> (D.Don) T.M. Schust. & Reveal #	MT1710	NE
	<i>Persicaria barbata</i> (L.) H. Hara	MT1632	LC
	<i>Persicaria capitata</i> (Buch - Ham Ex D.Don) H. Gross	MT1624	NE
	<i>Persicaria chinensis</i> (L.)H.Gross #	MT1787	NE
	<i>Persicaria hydropiper</i> (L.) Delarbre	MT1633	LC
	<i>Persicaria nepalensis</i> (Meisn.) H. Gross	MT1623	NE
	<i>Persicaria orientalis</i> (L.) Spach	NEIFM-89	NE
	<i>Persicaria strigosa</i> (R.Br.) H. Gross	NEIFM-540	NE
Primulaceae			
	<i>Ardisia solanacea</i> Roxb. #	MT1680	NE
	<i>Lysimachia debilis</i> Wall.	NEIFM-531	NE
	<i>Maesa indica</i> (Roxb.) Sweet #	MT1654	LC
Ranunculaceae			
	<i>Ranunculus cantoniensis</i> DC.	NEIFM-105	NE
Rafflesiaceae			
	<i>Sapria himalayana</i> Griff.	MT3010	NE
Rosaceae			
	<i>Potentilla indica</i> (Andrews) Th. Wolf #	MT1648	NE
	<i>Rubus ellipticus</i> Sm. #	MT1759	NE
	<i>Rubus moluccanus</i> L. #	MT1567	NE
	<i>Rubus niveus</i> Thumb #	MT1542	NE
	<i>Rubus paniculatus</i> Sm. #	MT1569	NE
	<i>Rubus rosifolius</i> Sm. #	MT1831	NE
	<i>Rubus sumatranus</i> Miq. #	MT1713	NE

Family	Taxon	Field no.	IUCN Status
Rubiaceae			
	<i>Argostemma sarmentosum</i> Wall.	MT2081	NE
	<i>Argostemma verticillatum</i> Wall.	MT2080	NE
	<i>Catunaregam spinosa</i> (Thunb.) Tirveng.	NEIFM-466	NE
	<i>Chassalia curviflora</i> var. <i>ophioxylodes</i> (Wall) Deb & B.Krishna	DB2079	NE
	<i>Coffea benghalensis</i> B.Heyne ex Schult.	MT1697	LC
	<i>Dentella repens</i> (L.) J.R.Forst. & G.Forst.	NEIFM-95	LC
	<i>Dimetia scandens</i> (Roxb.) R.J.Wang	DB1111	NE
	<i>Exallage auricularia</i> (L.) Bremek.	NEIFM-125	NE
	<i>Gomphostemma lucidum</i> Wallich ex Bentham	MT3006	NE
	<i>Ixora polyantha</i> Wight	MT2078	NE
	<i>Mussaenda glabra</i> Vahl #	MT2077	NE
	<i>Mussaenda roxburghii</i> Hook.f. #	MT2076	NE
	<i>Mycetia mukerjiana</i> Deb & Ratna Dutta	MT2075	E
	<i>Oldenlandia umbellata</i> L.	MT2519	NE
	<i>Paederia foetida</i> L. *	MT2115	NE
	<i>Psychotria monticola</i> Kurz	NEIFM-425	NE
	<i>Psychotria silhetensis</i> Hook.f.	NEIFM-426	NE
	<i>Scleromitron diffusum</i> (Willd.) R.J.Wang	NEIFM-158	NE
	<i>Uncaria macrophylla</i> Wall.	MT3005	NE
	<i>Wendlandia budleioides</i> Wall. ex Wight & Arn.	MT2114	NE
Rutaceae			
	<i>Citrus indica</i> Yu. Tanaka #	MT2074	E
	<i>Citrus latipes</i> (Swingle) Yu.Tanaka #	MT1688	NE
	<i>Citrus medica</i> L. #	MT2073	NE
	<i>Glycosmis pentaphylla</i> (Retz.) DC.	NEIFM-424	LC
	<i>Murraya koenigii</i> (L.) Spreng. #	MT2113	NE
	<i>Murraya paniculata</i> (L.) Jack	MT1675	NE
	<i>Murraya tetramera</i> C.C.Huang	NEIFM-429	NE
	<i>Toddalia asiatica</i> (L.) Lam.	MT1535	NE
	<i>Zanthoxylum armatum</i> D.C #	MT1534	LC
	<i>Zanthoxylum oxyphyllum</i> Edgew. #	MT1533	NE
	<i>Zanthoxylum rhetsa</i> (Roxb.) DC. #	MT2112	LC
Sabiaceae			
	<i>Sabia lanceolata</i> Colebr.	MT2111	NE
Salicaceae			
	<i>Casearia vareca</i> Roxb.	MT1597, MT2110	NE
Sapindaceae			
	<i>Aesculus assamica</i> Griff.	MT2109	NE
	<i>Nephelium lappaceum</i> L. #	MT2100	LC
Saurauraceae			
	<i>Houttuynia cordata</i> Thunb. *	MT2108	NE

Family	Taxon	Field no.	IUCN Status
Schrophulariaceae			
	<i>Buddleja asiatica</i> Lour	DB2107	LC
Simaroubaceae			
	<i>Brucea mollis</i> Wall. ex Kurz	NEIFM-243	NE
	<i>Ailanthus integrifolia</i> Lam.	MT2106	LC
Smilacaceae			
	<i>Smilax ovalifolia</i> Roxb. ex D.Don	MT2105	NE
	<i>Smilax zeylanica</i> L	MT2104	NE
Solanaceae			
	<i>Solanum spirale</i> Roxb. #	MT1524	NE
	<i>Solanum villosum</i> Mill. #	MT2103	NE
	<i>Solanum violaceum</i> Ortega #	MT1583	NE
	<i>Lycianthes biflora</i> (Lour.) Bitter	NEIFM-402	NE
Stemonaceae			
	<i>Stemona tuberosa</i> Lour	MT2099	NE
Styracaceae			
	<i>Styrax serrulatus</i> Roxb.	MT2098	NE
Tamaricaceae			
	<i>Tamarix dioica</i> Roxb. ex Roth	MT2520	NE
Theaceae			
	<i>Pyrenaria barringtoniifolia</i> (Griff.) Seem.	NEIFM-180	NE
	<i>Schima wallichii</i> (DC.) Korth.	MT2089	LC
Typhaceae			
	<i>Typha angustifolia</i> L.	MT1740	LC
Urticaceae			
	<i>Boehmeria penduliflora</i> Wedd. ex D.G.Long	MT1739	NE
	<i>Boehmeria pilosiuscula</i> (Blume) Hassk.	MT1741	LC
	<i>Debregezia longifolia</i> (Burm.f.) Wedd.	MT1762	NE
	<i>Dendrocnide sinuata</i> (Blume) Chew	MT1825	NE
	<i>Elatostema dissectum</i> Wedd. #	MT2088	NE
	<i>Elatostema sessile</i> J.R.Forst. & G.Forst.	MT1755	NE
	<i>Gonostegia hirta</i> (Hassk.) Miq. #	MT1568	NE
	<i>Gonostegia pentandra</i> (Roxb.) Miq.	NEIFM-173, NEIFM-264	NE
	<i>Pilea insolens</i> Wedd. #	MT1749	NE
	<i>Pilea umbrasa</i> Wedd. ex Blume #	MT1744	NE
	<i>Poikilospermum suaveolens</i> (Blume) Merr. #	MT1701	NE
	<i>Pouzolzia calophylla</i> W.T.Wang & C.J.Chen	NEIFM-412	NE
	<i>Pouzolzia zeylanica</i> (L.) Benn.	MT1750	NE
	<i>Urtica ardens</i> Link. *	MT1729	NE
	<i>Urtica dioica</i> L. #	MT2097	LC
Viburnaceae			
	<i>Sambucus adnata</i> Wall. ex DC.	MT2096	NE

Family	Taxon	Field no.	IUCN Status
Violaceae			
	<i>Viola betonicifolia</i> Sm.#	MT1540	NE
	<i>Viola pilosa</i> Blume #	MT1539	NE
	<i>Viola thomsonii</i> Oudem.	NEIFM-249	NE
Vitaceae			
	<i>Ampelocissus hoabinhensis</i> C.L.Li	NEIFM-514	NE
	<i>Causonis trifolia</i> (L.) Mabb. & J.Wen	NEIFM-455	NE
	<i>Cissus assamica</i> (M.A.Lawson) Craib	NEIFM-256	NE
	<i>Leea indica</i> (Burm.f.) Merr.	MT2093	LC
	<i>Parthenocissus semicordata</i> (Wall.) Planch.	NEIFM-198, NEIFM-454	NE
	<i>Tetrastigma leucostaphylum</i> (Dennst.) Alston	MT2087 ; MT2095	NE
Zingiberaceae			
	<i>Alpinia nigra</i> (Gaertn.) Burt #	MT1683	LC
	<i>Alpinia roxburghii</i> Sweet #	MT1591	NE
	<i>Amomum maximum</i> Roxb. #	MT1578	LC
	<i>Amomum pterocarpum</i> Thwaites #	MT1592	LC
	<i>Amomum subulatum</i> Roxb. #	MT2086	DD
	<i>Globba multiflora</i> Wall. ex Baker	MT2085	NE
	<i>Hedychium coccineum</i> Buch.-Ham. ex Sm	MT2084	NE
	<i>Hedychium stenopetalum</i> G.Lodd.	MT1627	NE
	<i>Hornstedtia arunachalensis</i> S. Tripathi & V. Prakash #	MT2083	E
	<i>Larsenianthus arunachalensis</i> M. Sabu, Sanoj & Rajesh Kumar	cf. Mibang & Das 2017	E, CR
	<i>Larsenianthus assamensis</i> S. Dey, Mood & S. Choudhury	cf. Mibang & Das 2017	E
	<i>Larsenianthus careyanus</i> (Benth. & Hook.f.) W.J. Kress & Mood	DB2082	NE
	<i>Zingiber sianginensis</i> Tatum & A.K. Das *	MT2083	E
	<i>Zingiber zerumbet</i> (L.) Roscoe ex Sm.	MT2084	DD

*—Ethnomedicine | #—Edible

CR—Critically Endangered | LC—Least Concern | NT—Near Threatened | E—Endemic | VU—Vulnerable | DD—Data Deficient.

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Crepuscular hunting of swiftlets (Family: Apodidae) by Besra (Family: Accipitridae) in the urban areas of the Andaman Islands, India

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Abstract: We report the crepuscular hunting behavior by the Besra *Accipiter virgatus*, on the Glossy Swiftlets *Collocalia esculenta affinis* and the Edible-nest Swiftlets *Aerodramus fuciphagus inexpectatus* in urban areas the Andaman & Nicobar Islands. Unlike other raptors in the islands, the Besra hunts at twilight often in the absence of moonlight or/and artificial light. Glossy and Edible-nest Swiftlets have been ranched in human habitations and their nests harvested for livelihood support of local communities under an ex situ conservation program. Using the focal animal sampling method, we recorded the hunting behavior of the Besra (the predator) on the swiftlets (the prey) for 40h (120 min/day for 20 days) at the ex situ swiftlet colony established in a house in the Middle Andamans. The Besra made 84 hunting attempts, with the highest success rate (15.4%) between 17.00–18.00 h. The catch rate was a mean of 4±11 (SD) per day. The maximum time that was used for attempt to kill the prey was two hours. Depredation of the Edible-nest Swiftlet by the Besra could affect ex situ conservation efforts, which can also lead to economic losses and retaliation against the raptor. Restricting perch sites for the raptor around ranching houses might reduce predation risks for the swiftlets.

Keywords: Andaman & Nicobar Islands, Besra, crepuscular hunting, Edible-nest Swiftlet, ex situ conservation, predatory behavior.

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Author contribution: AD—conceptualization, data curation, methodology, writing – original draft; DK—conceptualization, data curation, formal analysis, methodology, writing – original draft; PG—conceptualization, methodology, data curation, writing – original draft; SM—funding acquisition, investigation, project administration, resources, conceptualization, supervision, validation, writing – review & editing.

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INTRODUCTION

Availability of food and potential nesting sites play an essential role in adapting birds to survive urban environments (Marzluff 2016). Glossy Swiftlet *Collocalia esculenta affinis* and the Edible-nest Swiftlet *Aerodramus fuciphagus inexpectatus* (Cranbrook et al. 2013) are examples of birds that are adapted to urban habitats in the Andaman & Nicobar Islands (Manchi & Mane 2012). This successful adaptation is due to the availability of nesting locations and food. Some cave-dwelling swiftlets such as the Edible-nest Swiftlet are of human interest and famous worldwide for 'bird's nest soup,' a delicacy in Chinese cuisine, and traditional Chinese medicine (Koon & Cranbrook 2002; Chantler & Boesman 2019). The nests exclusively built using bird's saliva have a high market value due to their medicinal properties. These birds are farmed for mass production of the edible nests to sustain international demand (Hobbs 2004; Thorburn 2015).

In the past several decades, the nests have been illegally harvested from caves in the Andaman Islands, impacting the population. For sustainable use of the birds' nest and to reduce the pressure on wild populations, an ex situ conservation program was started in urban houses, which habituated the Edible-nest Swiftlet to nest in human-made structures. The Glossy Swiftlet also nests in abandoned houses, bridges, and jetties. As part of the conservation program, the swiftlets are attracted to human habitation and reared in an artificial structure, known as ex situ swiftlet house (Manchi & Mane 2012). Scientists and managers with years of efforts, have successfully attracted the Edible-nest Swiftlet population to breed in one such ex situ structure (henceforth 'swiftlet house') in the Middle Andamans by using the sympatric Glossy Swiftlet as foster parents to hatch the eggs and rear the chicks (Sankaran & Manchi 2008). This swiftlet house supports a colony of Glossy Swiftlet (~1,000 individuals) and Edible-nest Swiftlet (8–10 individuals). The Edible-nest Swiftlet (ENS) shares the nesting site with the Glossy Swiftlet. Some young ENS birds are also known to build nests on existing Glossy Swiftlet nests.

During our study, the swiftlet population in the house was hunted by the raptor, Besra *Accipiter virgatus abdulalii*, at dawn and dusk. Besra is a small, diurnal bird of prey of the Order Falconiformes. Out of 11 subspecies of *A. virgatus*, *A.v. abdulalii* is an endemic subspecies restricted to the Andaman Islands (Naoroji 2006; Clark & Marks 2020). Due to the easy availability of the prey, and its nesting locations, Besra has adapted to the urban

habitats in the Andaman Islands. As the other members of the genus *Accipiter*, the Besra is a swift and silent hunter. The species is quick on the wing in quest of prey, often twisting and turning to chase the bird and escape hindrances. Besra breeds from March to May and nests on roadside trees, near human habitation, mangrove forests, and Padauk plantations (Ali & Ripley 1978). Being an opportunistic hunter, the diet of Besra includes small birds, insects, and some mammals and reptiles, depending on the season and availability of prey. It primarily preys on birds during the breeding season (Huang et al. 2004). In the Andaman Islands, the Besra predated on poultry near human habitation. It is also a potential predator of swiftlets breeding in the limestone caves of the Andaman Islands, according to Manchi & Sankaran (2009).

Foraging in twilight is generally regarded as an end or beginning of diurnal or nocturnal activity. While crepuscular foraging is uncommon, many nocturnal predators such as owls, nightjars, and waders begin hunting at dusk. The circumstances under which crepuscular hunting occurs are not well understood or documented (Martin 1990). Among diurnal raptors other than owls, the smaller species of the genus *Falco* (Lesser Kestrel *F. naumanni*, Kestrel *F. tinnunculus*, Hobby *F. Subbuteo* and Sooty Falcon *F. concolor*) and large-bodied falcons such as the peregrine falcon hunt at night under artificial lights such as street lamps, and moonlight (Ratcliffe 1980; Pierson & Donahue 1983).

As the swiftlet ranching in the ex situ houses aims to provide livelihood support for the economic development of the local populations in the Andaman Islands, the threat from Besra can have significant implications. Therefore, we studied the crepuscular hunting behavior of Besra. Further, we hypothesized that the successful hunting of swiftlets by Besra is associated with i) time of day (diurnal vs. crepuscular), ii) availability of perch sites, and iii) flock size of swiftlets.

MATERIALS AND METHODS

Study Area

The Andaman group of islands is in the northeastern Indian Ocean, along the southern extension of the Arakan Yoma mountain range, are peaks of a submerged continuous mountain ridge extending up to Sumatra in the south, between latitude 6.75–13.68 °N and 92.20–93.95 °E. The Andaman group of islands are divided into (a) South Andaman, (b) Middle Andaman, (c) North Andaman, (d) Baratang, and (e) Rutland. The forest

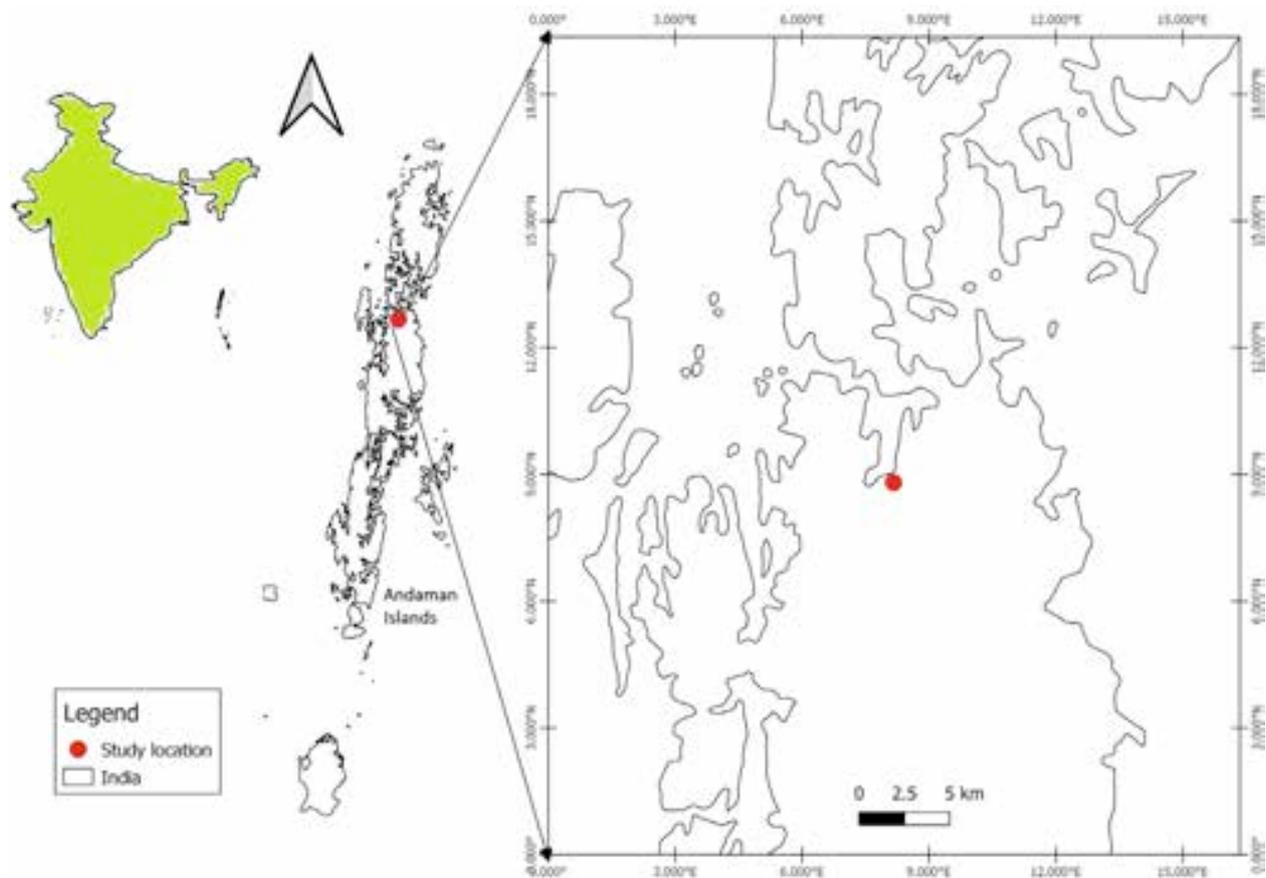


Figure 1. Geographical location of the study site.

types range from the tropical wet evergreen forest towards the south to tropical moist deciduous forest in the North Andaman group of islands (Davidar et al. 2001; Champion & Seth 2005). Parts of the islands also have human settlements and agricultural fields surrounded by deciduous forests. This island group, with a high proportion of endemic flora and fauna, is one of the global biodiversity hotspots in the world (Conservation International 2005). The archipelago has 19 identified Important Bird and Biodiversity Areas (Rahmani et al. 2016) and is also recognized as an endemic bird area (Birdlife International 2019). Including endemic species such as Andaman Serpent Eagle *Spilornis elgini*, Great Nicobar Serpent Eagle *Spilornis klossi*, Central Nicobar Serpent Eagle *Spilornis minimus*, and Nicobar Sparrowhawk *Accipiter butleri*; the Andaman & Nicobar Islands have 22 raptor species.

We conducted the present study in the northernmost part of Middle Andaman in Tugapur near Mayabunder town (Figure 1). The temperature in Tugapur during April was between 27°C and 35°C (Accuweather 2018). Sunrise and sunset were between 05.01h & 05.16h and

17.31h & 17.33h, respectively (Time and Date 2018). Human habitation and small patches of deciduous forest surround the swiftlet house. Plant species such as *Azadirachta indica* A. Juss., *Gliricidia sepium* (Jacq.) Walp., *Calamus* sp. L., *Tectona grandis* L.f., and *Ficus religiosa* L. are seen around the swiftlet house. We made daily visits to the swiftlet house during the late summer season from 02 April to 28 April 2018 between 16.00h and 18.00h. The study period coincided with the breeding season of the Besra (Ali & Ripley 1978) and swiftlets (Manchi & Sankaran 2014). The activities of the Besra near the swiftlet house were studied using focal animal sampling (Altmann 1974). We made observations from the moment the individual arrives around the swiftlet house until it leaves the site. Simultaneously, swiftlets around the house were observed to understand the interaction between the prey and predator. Three observers were stationed at different observation stations around the house to observe and note the behavior of the Besra along with the time (Figure 2). With 120 min of observation every day, we collected data for 2,400 min in 20 days. Occasionally, we prolonged the observations

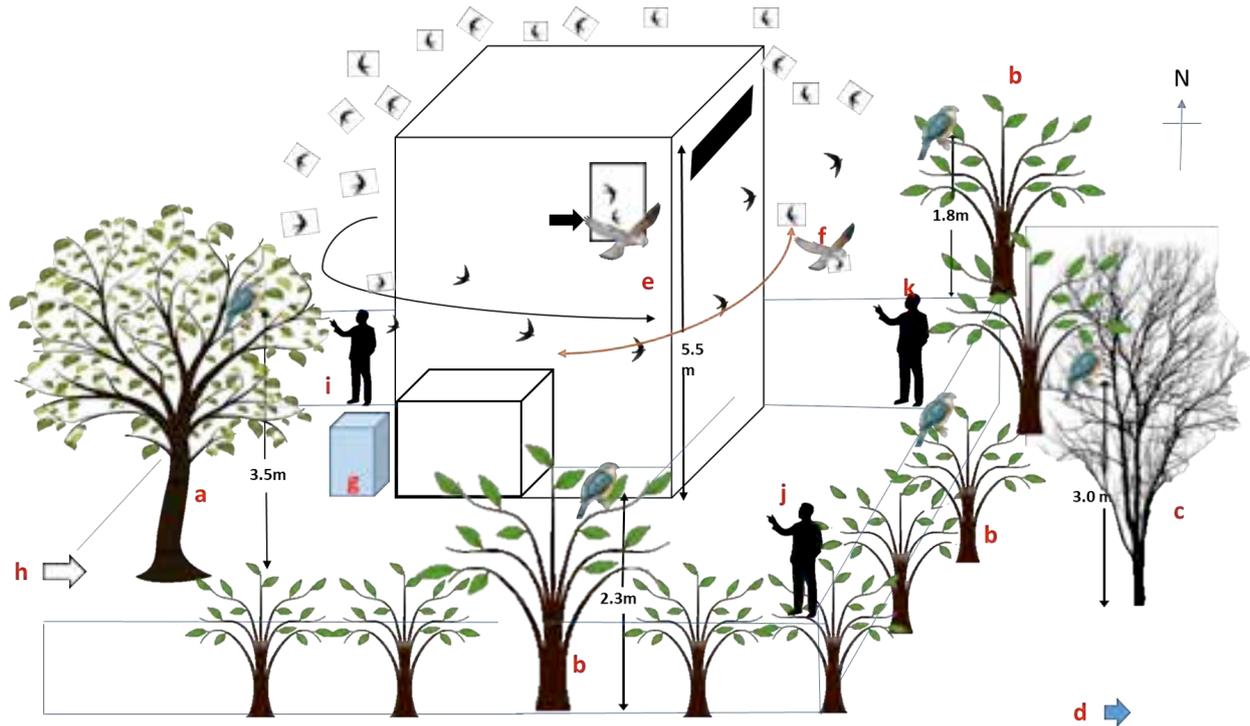


Figure 2. Study site: a—*Azadirachta indica* | b—*Gliricidia sepium* | c—*Tectona grandis* | d—towards *Ficus religiosa* | e—entrance of the swiftlet house for swiftlets | f—frequently used path by Besra | g—water tank | h—entry of the study site | i–j–k—observation stations.

until 18.30h to check the presence and activity of Besra. Because of unfavorable circumstances, we could make only one observation during the morning hours (04.45–07.00 h). To process the collected data, we used XL-STAT software (Ver. 2020, Addinsoft 2020). We excluded the set of data from the single morning visit during analysis. Ethograms of Besra was made based on behavioral observations. No ethical approval was obligatory for this study.

RESULTS

The present study confirms the Besra as a predator of the swiftlets. Swiftlets being diurnal foragers, the breeding individuals keep returning to the breeding location to feed their nestlings, and the arrival of the Besra was at 16.00h. During the 20 days of observations, the Besra made 84 hunting attempts per day to catch swiftlets with 4 ± 11 (Mean \pm SD). It preyed upon 1 ± 0.5 swiftlets per day, with 14.92% (n=84) successful attempts. The most hunting attempts were between 17.31–17.50 h (67.80 %), and most successful (11.48%) and unsuccessful (56.32%) attempts were also made during the same period (Figure 3). Comparing the sunset timings (17.31h to 17.33h on observation days)

and the period of most hunting attempts were during the crepuscular (twilight) hours (Time and Date 2018). It confirmed the crepuscular hunting behavior of Besra, the predator. Out of the total time the besra spent near the ex situ house, 48% was spent attempting a kill while 35% time was spent on the perch (Figure 4).

All the hunting attempts, successful or unsuccessful, were on the flock of swiftlets swarming in the lower canopy (2–5 m above ground) and never above the canopy. The number of attempts increased as the swiftlet flock size increased (Table 1). The Besra captured and carried its kill with its claws. Immediately after capture, the predator brought each kill to the same perch of attack (mostly *Azadirachta indica*) or the nearest landing site (mostly *Tectona grandis* within 10m). As the swiftlets gathered around the house in small groups (15–20 birds), the Besra would chase the selected prey with athletic ease, twisting, turning, and swooping close to the ground (6 in Figure 2) as the aerodynamic swiftlet tried to break away from its grasp. The predator and the prey are swift flyers, and thus sometimes swiftlets managed to evade the predator successfully. The height at which the chase was made, particular areas around the swiftlet house, and the time of the hunt influenced a successful kill.

Table 1. Table 1. The hunting attempts by the Besra (the predator) at a given time in relation to the flock size of swiftlets (the prey) (Note: flock size: <50—small | 50–200—medium | 200–500—large)

Time (h)	Flock size of swiftlets	Total hunting attempts (% success)
16.20–16.30	small	1 (0)
16.30–16.40	small	1 (0)
16.40–16.50	small	6 (50)
16.50–17.00	large	3 (0)
17.00–17.10	large	0
17.10–17.20	large	6 (0)
17.20–17.30	large	11 (0)
17.30–17.40	large	35 (14.28)
17.40–17.50	small	21 (23.80)
17.50–18.00	small	0
	Total	84

The Besra usually ripped off the feathers of its prey before consuming the kill. Occasionally, it would swallow the feathers too. The Besra spent 16% of its time in consuming the kill (Figure 4). By examining the kill remnants, we confirmed that the prey species was Glossy Swiftlet, however, once the population of the Edible-nest Swiftlet grows, we cannot deny the possibility of it being hunted by the Besra.

The most hunting attempts by Besra were from a particular perch in the canopy of a large *Azadirachta indica* (46.17%), beside the swiftlet house, however, the Besra was observed using other perching sites around the swiftlet house as well (Table 2). After each attempt, it often perched (30.76%) on *Gliricidia sepium* (2 in Figure 2). The predator returned to the same perch on *Azadirachta indica* for the next attempt to catch a swiftlet.

DISCUSSION

Swiftlets are known to flock near their breeding and roosting sites during dawn and dusk while leaving or returning to their roosts (Tarburton 2009; Mane & Manchi 2017). The cave-dwelling swiftlets are usually known to be vulnerable while entering and exiting the caves (Mane & Manchi 2017). The present study, however, is the first to document aerial predation of swiftlets by the Besra around human-habitation. Flying in flocks equips the swiftlets to successfully evade predators. The time of arrival at the roost (Mane & Manchi 2017, 2019) and the speed used for entering the

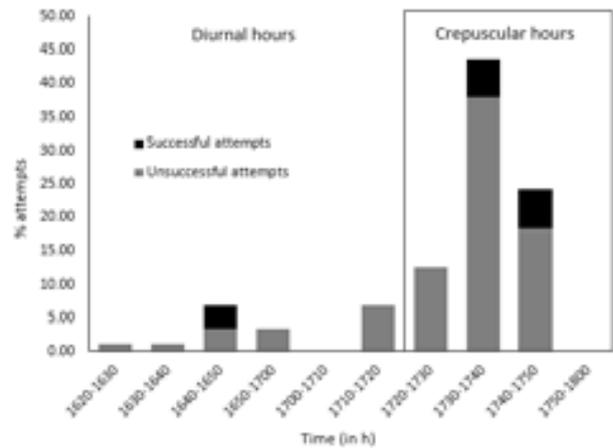


Figure 3. Hunting attempts (in %) by the Besra in relation to the time of the day.

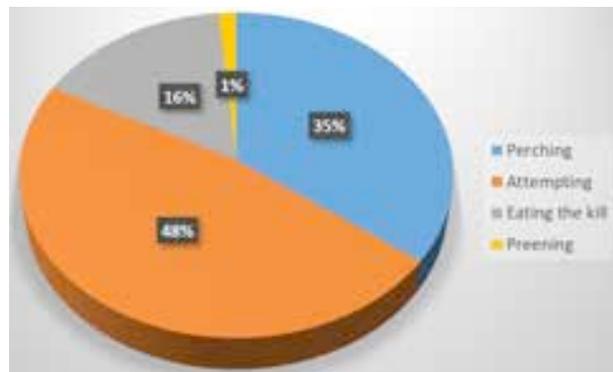


Figure 4. Activity time budget of Besra near the ex situ swiftlet house, Middle Andaman Island.

breeding/roosting sites (Tarburton 2009) are counted as the anti-predatory behaviors of the swiftlets.

Approximately 20km from the swiftlet house, we located a Besra chasing the swiftlets during morning and evening twilight hours. This location hosts thousands of Glossy Swiftlets. Here, we observed the swiftlets forming to chase away the predator. Therefore, we can see two anti-predatory mechanisms in the swiftlets, one to escape the predator by forming groups, and secondly by mobbing the predator by forming larger groups. The anti-predatory behavior may limit the number of prey obtained by the Besra. Apart from the Besra, we noticed that the Shikra *Accipiter badius* attempting to hunt swiftlets near the swiftlet house between 16.00 and 16.30 h from the *Ficus benghalensis* L. trees which was the tallest in the study site. The Shikra chased the swiftlets swarming in the open air but was unsuccessful in catching a bird.

As the number of Glossy Swiftlets in the swiftlet

Table 2. Perch sites used by Besra while hunting the swiftlets near the ex situ swiftlet house, Middle Andaman.

	Scientific name of the perch tree	Common name of the perch tree	Height of perch (meters)	Total hunting attempts (% success)
1	<i>Azadirachta indica</i> A. Juss	Neem	3.5	13 (46)
2	<i>Gliricidia sepium</i> (Jacq.) Walp	Gliricidia	2.3	13 (15)
3	<i>Gliricidia sepium</i> (Jacq.) Walp	Gliricidia	1.8	13 (8)
4	<i>Gliricidia sepium</i> (Jacq.) Walp	Gliricidia	2.3	13 (8)
5	<i>Tectona grandis</i> L.f.	Teak	3	7 (14)
6	<i>Ficus religiosa</i> L.	Peepal	~7	26 ((8)

house is greater than the Edible-nest Swiftlet, we can say that the Besra may be preying on the more easily available target. It makes the Edible-nest Swiftlet as vulnerable to predation as the Glossy Swiftlets, putting the already low population at risk.

Sparrow-hawks use at least a 15cm circumference branch of a tree or a moderately flat surface for perching (Owen 1932). Although we could not confirm the same for the Besra, the repetitive use of the same perch shows some preference for perches. To avoid predation of swiftlets, it is essential to avoid such ideal ‘perches’ or trees near the swiftlet house. Removal of these ideal available perches might drastically reduce the rate of predation of the swiftlets by the Besra, at least till the individual finds and learn to use any other perching site and hunting strategy. The predation behavior depends mostly on the cognitive ability of an individual; it is challenging at present to ascertain the change in the predation rates by the Besra after removal of the perch trees around the house.

Many raptors have shown adaptations in hunting and foraging behavior according to the prey resources. The Bat Hawk *Macheiramphus alcinus* in sub-Saharan Africa, South Asia, and New Guinea, is known to hunt bats during the twilight hours. The bat hawk’s adaptation in morphology and behavior results from the prey’s biology (Black et al. 1979). The Bat Falcon *Falco ruficularis*, found in Mexico, and Central and South America, is another bird that hunts bats at dusk. The Bat Falcon has also adapted to hunting in urban environments (Seijas 1996). The Peregrine Falcon, a diurnal bird, is also reported to prey on bats (Pierson & Donahue 1983), and some even during dusk or at night (Rejt 2001). Mester & Oliver (2018) reported an unusual fishing behavior of the Eurasian Sparrowhawk during winter when the prey are scarce. Their report suggested, ‘Raptors may be able to adopt unusual hunting or foraging behavior and prey selection in response to changing environmental conditions and changing availability of potential prey

resources.’ Therefore, with a more focused study and observations throughout the islands, it can be concluded whether *A. v. abdulalii* of Andaman Islands has adapted to crepuscular hunting owing to the swiftlet’s ecology, and whether this behavior is limited to urban areas (in ex situ structures) or exists everywhere.

CONCLUSION

Since the swiftlet-farming efforts are in the direction of livelihood development and economic growth in the islands, any hindrance to such a program will ultimately affect the local human populations. The raptor-human conflict that may arise from this is a threat to the survival of both (prey and predator) species. Based on the observations discussed, we now recommend maintaining the premises of the swiftlet houses by restricting the vegetation growth for better management of the swiftlet populations in urban areas and also to avoid future raptor-human interaction.

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A study on diversity of mammalian species using camera traps and associated vegetation in Mizoram University Campus, Aizawl, Mizoram

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Abstract: Fragmented forests often have conservation value, serving as a refuge or corridors for small mammalian species. In the study, the diversity of mammals was studied within Mizoram University (MZU) campus. Forty-eight plant species from 25 families were recorded on different sites. Thirty quadrates were nested at the locations of occurrence of mammalian species for vegetation type analysis. *Schima wallichii* was the most dominant plant species with the highest IVI values of (31.7%), followed by *Aporosa octandra* (22.93%) and *Castanopsis tribuloide* (21.17%). Camera trap method was used to collect information about the mammalian diversity in the campus. The mammalian species recorded in this study makes 15% of mammalian fauna of the state of Mizoram. Twelve mammal species and six bird species were recorded by the camera traps. With proper awareness among residents of the campus and a planned approach for developmental activities, the findings of our study can make an important extension for the coexistence of mammalian species and long term survivability within MZU campus.

Keywords: Camera traps, coexistence, mammalian diversity, Mizoram University, Urban biodiversity.

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Author details: J.H. ZOTHANPUII is a MSc student of Department of Zoology, Mizoram University and the work is a part of her dissertation work. She is currently preparing for future possible research opportunities. SUSHANT GOUDA is a PhD scholar and is currently engaged in a NMHS based project under the Department of Zoology, Mizoram University. He is active in the field of wildlife biology and conservation education. His major research area is ecology and distribution of Himalayan bear. ABINASH PARIDA is a senior research fellow in the NMHS project in department of zoology. He is a vigorous field biologist and working on faunal diversity and conservation. His major research area is primate diversity, ecology, and conservation. G.S. SOLANKI is Professor in Zoology and Principal Investigator of Nation Mission on Himalayan Studies (NMHS) project in the Department of Zoology. This project is funded by G.B. Pant National Institute Himalayan Environment and Sustainable Development, Almora, Uttarakhand. Major research area is ecology and conservation of wilderness species with main emphasis on primates.

Author contribution: JHZ and AP have carried out the research work and collected the data. SG helped in preparation of this the manuscript. GSS was Principal Investigator and guiding teacher, and provided work plan and all logistic support required.

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INTRODUCTION

The northeastern region of India shares two biodiversity hotspots and has rich mammalian and avian diversity (Choudhury 2006). The state of Mizoram is part of the Indo-Myanmar Biodiversity Hotspot region and harbours 126 species of mammals including 37 threatened species (Lalthanzara 2017). The rapid expansion of shifting cultivation and urbanization in the region has, however, led to a contemporary changes in landscape, forest fragmentation, and ecosystem modification (Teegalapalli et al. 2009; Yadav 2013). Such anthropogenic activities have created a mosaic of remnant forest patches of varying size, demarcated by a network of roads, concerts, and settlements areas (Mazumdar et al. 2011). Recent researches on biodiversity conservation had put urban areas having rich biodiversity into the limelight for the long-term persistence of native species (Ordenana et al. 2010; Lopucki & Kitowski 2017; Hill et al. 2018). Urbanization is often considered to have several negative impacts on the native flora and fauna; nevertheless, it also serves as valuable habitat and corridor for dispersal of certain animal species (Opdam et al. 2003; Fernandez & Simonetti 2013). Urban and semi-urban green forest patches serves as surrogate and refuge habitat to be utilized for dispersal and migration for birds and many small to medium size mammals (Gallo et al. 2017). The role of degraded forest landscapes and patches within the campus of academic institutions as a potential habitat for small mammals and birds has also been acknowledged in some studies (Vallejo et al. 2008; Mazumdar et al. 2011; Voon et al. 2014; Nerlekar et al. 2016).

The camera trap is a useful technique widely used for collecting information on elusive species (Kelly et al. 2008; Linkie & Ridout 2011). Information on species diversity and distribution is considered to be the primary need for conservation activities (Geldmann et al. 2013; Brncic et al. 2015). While information on biodiversity in megacities and large urban areas are easily available, the same is scarce in small cities and educational campuses (Lopucki et al. 2013; Lopucki & Kitowski 2017). Mizoram University (MZU) campus at Aizawl, Mizoram encompasses a large area with lush green landscape having several patches of natural forest vegetation. Although research on herpetofauna (Vanlalhlimpua 2012), butterfly (Baruah 2017), spiders (Lalthafamkima 2017), and birds (Sailo et al. 2019) are available from MZU campus, no study has been undertaken on the distribution of mammalian species. Therefore, a study

was planned to determine the mammalian diversity using mainly camera traps and direct observations within the campus.

MATERIALS AND METHODS

Study area

Mizoram University (MZU) is located at the western end of the Aizawl town, the state capital of Mizoram at a distance of about 15km. The campus of MZU is spread in an area of 978.1988 acres and lies between 23.7394°N & 92.6651°E. The elevation ranges between 300–880 m. University campus encompasses the administrative block, academic blocks, hostels, and other facilities which are sparsely embedded within regenerating tropical wet evergreen and semi-evergreen forests (Image 1). The campus also harbors a protected forested water catchment reserve in the north and a small biodiversity park. In the vicinity of the campus, there are settlement areas, where residents undertake agriculture practice (jhum cultivation), which is the primary source of income and livelihood. Several small natural and seasonal streams also flow through the campus. Vegetation profile is dominated by trees comprised of 384 species of vascular plants from 290 genera and 107 families (Lalchhuanawma 2008; Rai 2016; Sailo et al. 2019). A map of the university is presented in Figure 1.

Vegetation analysis

A preliminary survey of campus area was carried out to select the sampling site for vegetation study. Transects were set up in areas where signs and evidence of animals' presence were observed. For vegetation sampling 10m x 10m quadrates were plotted at every 200m interval along transects for tree diversity analysis. Thirty quadrates were plotted in different sites and only trees having DBH more than 10cm were considered for the analysis (Daniels et al. 1996). Vegetation analysis was done for the relative frequency of occurrence of the species; relative density, relative abundance and important value index (IVI) were determined following methods of Lalchhuanawma (2008) and Ahmed (2012).

Camera trapping

The camera traps are often used for understanding the secretive behavior or nocturnal activity, and estimating animal population comprehensively even at relatively low densities of animals (Ordenana et al. 2010; Gouda et al. 2020). Camera deployment points were chosen based on the presence of visible animal trails, footprints,

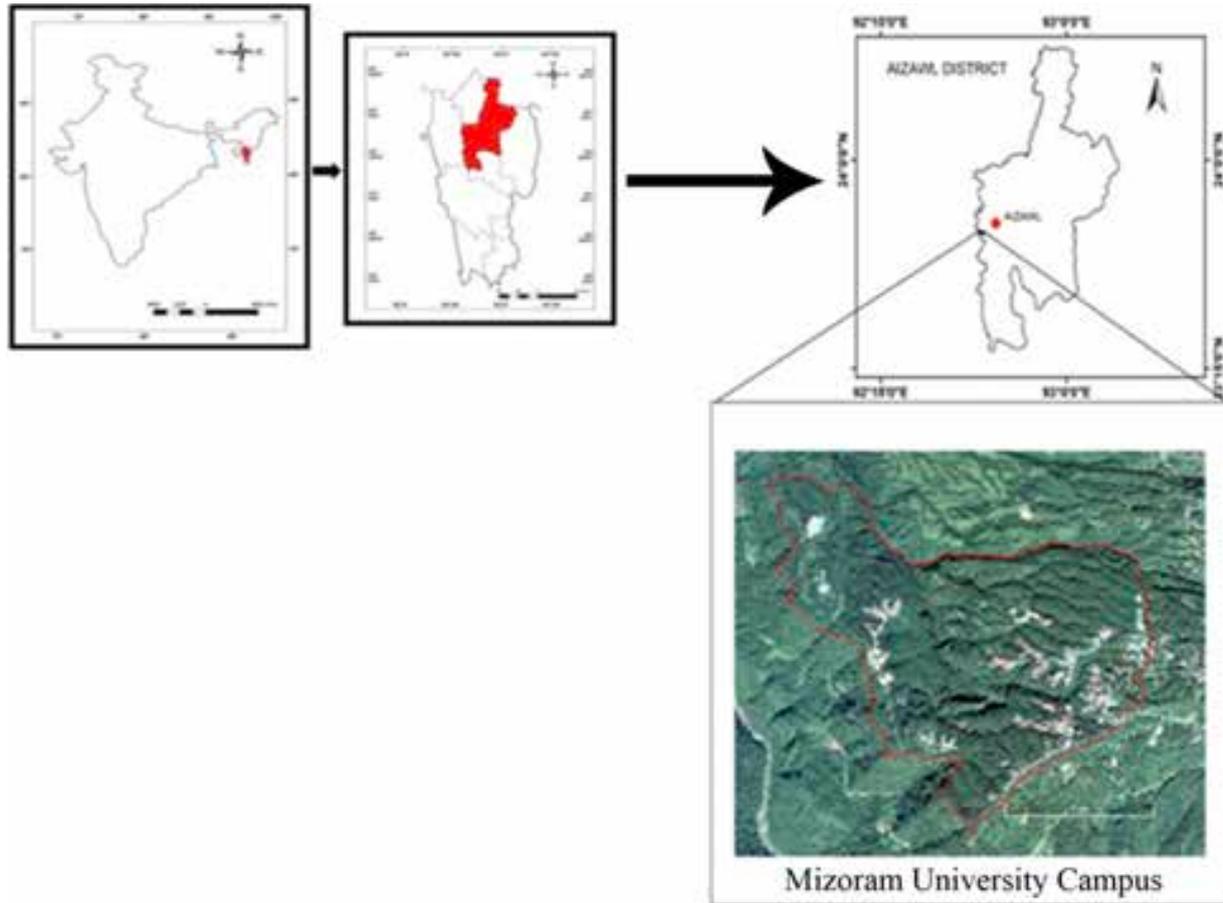


Figure 1. Location of Aizawl City and Mizoram University campus.

scats, activity areas such as dens or close to streams (Sasidhran et al. 2016). Nine camera traps (Cuddeback) were randomly deployed at 15 sampling points within the study sites for 45 days as presented in Figure 2. The cameras were installed on a tree trunk at a suitable height of 30–50 cm above the ground at optimum angles based on slope conditions for viewing the animal trails without the camera view being blocked by any objects (Image 1). Camera traps were programmed to take sequential photographs with five seconds delay registering date and time for each exposure. On average the camera traps were operational for five days and were checked every two days for photos and battery replacement. GPS (Garmin map 78S) was also used to record coordinates, elevation, and slope of camera trap locations. The relative abundance index (RAI) value was calculated using the method suggested by Jenks et al. (2011). Whenever possible we also recorded directly observed species especially arboreal ones around the camera trap locations.

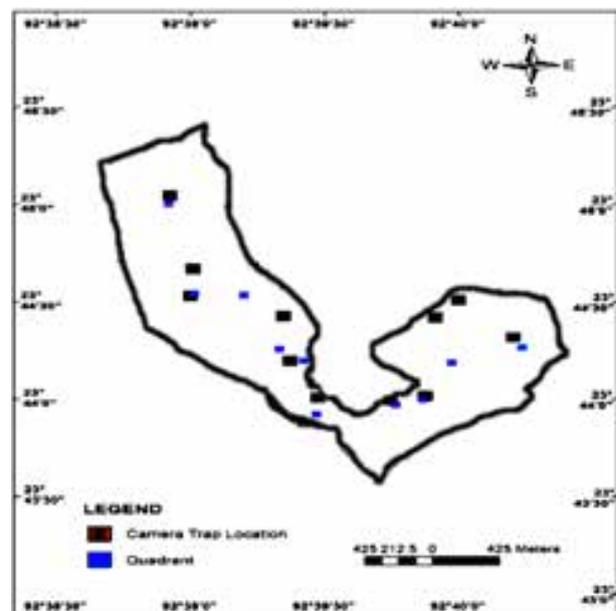


Figure 2. Location of camera traps and quadrates for vegetation sampling.



Image 1. a & b—view of MZU campus | c—installation of camera traps | d—habitat with water source, footprints, and tracks of different animal species. © Abinash Parida.

RESULTS

Forty-eight plant species which belongs to 25 families were recorded along the transect lines during the vegetation study. Vegetation samples were analyzed for the parameters mentioned in Table 1. The analysis showed that *Schima wallichii* is the most dominant plant species having the highest IVI values (31.7%), followed by *Aporosa octandra* (22.93%), *Castanopsis tribuloide* (21.17%), and *Syzygium praecox* (17.86%). Plant species such as *Hibiscus macrophyllus*, *Bischofia javanica*, *Cinnamomum cassia*, and *Acer laevigatum* were some of the least available species (Table 1). The family Euphorbiaceae contributed for the highest IVI, followed by Theaceae and Fagaceae while the family Pandanaceae had the lowest IVI values among the recorded plant families (Figure 3).

The mammalian diversity on MZU campus is 19 mammalian species; 12 were captured through camera traps (Table 2) (Image 2) and seven were recorded by direct observations (Table 3). Six bird species were also

photo-captured during this study (Table 4) (Image 3). Among the mammalian species, civets are predominant with 30.17% followed by felids (25%). The presence of Marbled Cat *Pardofelis marmorata* (Near Threatened), Clouded Leopard *Neofelios nebulosa* (Vulnerable), and Golden Jackal *Canis aureus* (Least Concern) were important and majestic mammalian species. Such valuable observations from the fragmented habitat of an urban landscape are significant for species conservation. Other recorded species are categorised as Least Concern. The relative abundance index (RAI) value was highest (16.67) for Crab-eating Mongoose *Herpestes urva* followed by Common Palm Civet *Paradoxurus hermaphroditus* (13.3), and Leopard Cat *Prionailurus bengalensis* (10) (Table 2). Sites with minimal human disturbance and water bodies within the campus had higher RAI values. Among directly observed mammalian species squirrels and shrews were the major ones (Table 3). Seven avian species namely Khalij Pheasant (male and female), Blue Whistling Thrush, Forktail, Asian Barred Owlet, Long-tailed Nightjar, and Large Cuckoo

Table 1. Vegetation composition and analysis.

	Name of species	Mizo name	Family	Relative density (%)	Relative Frequency (%)	Relative Dominance (%)	Important Value Index(IVI)
1	<i>Acer laevigatum</i>	Thingkhim	Aceraceae	0.54	0.36	0.38	1.28
2	<i>Albizia chinensis</i>	Vang	Mimosaceae	2.20	1.84	1.93	5.97
3	<i>Albizia richardiana</i>	Theichhawl	Mimosaceae	2.20	1.65	3.23	7.08
4	<i>Albizia procera</i>	Kangtek	Mimosaceae	1.37	1.10	2.48	4.95
5	<i>Anogeissus acuminata</i>	Zairum	Combretaceae	0.29	0.73	0.69	1.71
6	<i>Aporosa octandra</i>	Chhawntual	Euphorbiaceae	10.09	4.61	8.23	22.93
7	<i>Artocarpus lakoocha</i>	Theitat	Moraceae	0.83	1.10	1.01	2.94
8	<i>Balakata baccata</i>	Thingvawkpui	Euphorbiaceae	1.37	3.68	1.84	6.89
9	<i>Bischofia javanica</i>	Khuangthli	Euphorbiaceae	0.12	0.01	0.15	0.28
10	<i>Bombax insigne</i>	Pang	Bimbaceae	1.78	1.47	3.29	6.54
11	<i>Callicarpa arborea</i>	Hnahkiah	Verbenaceae	4.4	3.13	1.95	9.48
12	<i>Castanopsis indica</i>	Sehawr	Fagaceae	3.74	1.84	1.40	6.98
13	<i>Castanopsis lanceifolia</i>	Vawmbuh	Fagaceae	0.41	1.29	0.87	2.57
14	<i>Castanopsis tribuloides</i>	Thingsia	Fagaceae	8.43	3.13	9.61	21.1
15	<i>Cedrelatoona</i>	Tei	Meliaceae	0.29	0.36	0.30	0.95
16	<i>Cinnamomum cassia</i>	Thakthing	Lauraceae	0.12	0.01	0.13	0.26
17	<i>Cinnamomum tamala</i>	Tespata	Lauraceae	2.36	1.47	1.81	5.64
18	<i>Colona floribunda</i>	Hnahthap	Tiliaceae	0.12	0.36	0.17	0.65
19	<i>Cordia wallichi</i>	Muk	Boraginaceae	0.83	2.02	2.05	4.9
20	<i>Derris robusta</i>	Thingkha	Fanaceae	1.12	1.29	0.92	3.33
21	<i>Derris thrysiflora</i>	Hulhu	Papilionaceae	0.12	0.01	0.19	0.32
22	<i>Drymicarpus racemosus</i>	Vawmbal	Anacardiaceae	3.44	1.65	5.24	10.3
23	<i>Duabanga grandiflora</i>	Zuang	Sonneratiaceae	0.12	0.01	0.28	0.41
24	<i>Embilica officinales</i>	Sunhlu	Euphorbiaceae	0.54	0.55	0.51	1.6
25	<i>Erythrina variegata</i>	Fartuah	Papilionaceae	0.41	0.73	1.99	3.13
26	<i>Ficus hirta</i>	Sazutheipui	Moraceae	0.54	0.62	0.67	2.13
27	<i>Glochidion heyneanum</i>	Thingpawchhia	Euphorbiaceae	2.78	2.02	2.15	6.95
28	<i>Gmlina arborea</i>	Thlanvawng	Verbenaceae	0.71	0.55	1.22	2.48
29	<i>Haldina cordifolia</i>	Lungkhup	Rubiaceae	2.20	1.10	1.62	4.92
30	<i>Hibiscus macrophyllus</i>	Vaiza	Malvaceae	0.12	0.01	0.07	0.20
31	<i>Ilex godajam</i>	Thinguihahni	Aquifoliaceae	0.41	0.73	0.82	1.96
32	<i>Lithocarpus elegans</i>	Thingpuithing	Fagaceae	0.70	0.36	0.72	1.78
33	<i>Litsea mono petala</i>	Nauthak	Lauraceae	2.07	1.84	1.60	5.51
34	<i>Macaranga indica</i>	Hnahkhar	Euphorbiaceae	4.86	2.94	3.74	11.54
35	<i>Macropanax undulatum</i>	Phuanberh	Araliaceae	0.41	0.92	0.85	2.18
36	<i>Messua ferrea</i>	Hershse	Guttiferae	0.70	0.36	0.31	1.37
37	<i>Pandanus fascicularis</i>	Ramlakhuih	Pandanaceae	0.41	1.47	0.33	2.21
38	<i>Protium serratum</i>	Bil	Burseraceae	1.37	0.92	1.45	3.74
39	<i>Saurauia punduana</i>	Tiar	Actinidiaceae	2.78	1.29	1.96	6.03
40	<i>Schima khasiana</i>	Khiangzo	Theaceae	0.64	0.92	0.67	2.23
41	<i>Schima wallichii</i>	Khiang	Theaceae	13.29	4.05	14.36	31.70
42	<i>Sterculia villosa</i>	Khaupui	Sterculaceae	1.12	0.92	1.33	3.37
43	<i>Stereospermum tetragonum</i>	Zinghal	Bignoniaceae	0.41	1.17	0.75	2.63

	Name of species	Mizo name	Family	Relative density (%)	Relative Frequency (%)	Relative Dominance (%)	Important Value Index(IVI)
44	<i>Syzigiumcumini</i>	Lenhmui	Myrtaceae	0.71	1.47	1.43	3.61
45	<i>Syzigium praecox</i>	Hmuifang	Myrtaceae	6.35	2.76	8.75	17.86
46	<i>Toona ciliate</i>	Teipui	Meliaceae	0.41	0.55	0.94	1.90
47	<i>Trema orientalis</i>	Belphuar	Cannabaceae	0.41	0.92	9.03	10.36
48	<i>Wendlandia budieioides</i>	Batling	Rubiaceae	3.19	2.21	1.73	7.13

Table 2. Mammalian diversity based on photo captured and their status

	Family	Common name	Scientific name	IUCN status	RAI
1	Felidae	Clouded Leopard	<i>Neofelis nebulosa</i>	Vulnerable	3.33
2	Felidae	Marbled Cat	<i>Pardofelis marmorata</i>	Near threatened	3.33
3	Felidae	Leopard Cat	<i>Prionailurus bengalensis</i>	Least concern	10.00
4	Suidae	Wild Pig	<i>Sus scrofa</i>	Least concern	3.33
5	Viverridae	Large Indian Civet	<i>Viverra zibetha</i>	Least Concern	13.33
6	Viverridae	Small Indian Civet	<i>Viverricula indica</i>	Least concern	10.00
7	Viverridae	Common Palm Civet	<i>Paradoxurus hermaphroditus</i>	Least concern	13.33
8	Herpestidae	Crab-eating Mongoose	<i>Herpestes urva</i>	Least concern	16.67
9	Canidae	Golden Jackal	<i>Canis aureus</i>	Least concern	3.33
10	Viverridae	Himalayan Palm Civet	<i>Paguma larvata</i>	Least concern	6.67
11	Mustelidae	Ferret Badger	<i>Melogale sp.</i>	Least concern	3.33
12	Tupaiidae	Tree Shrew	<i>Cladobats belangari</i>	Least concern	3.33

Shrike were also captured by camera traps in different areas of the campus (Table 4, Image 3).

DISCUSSION

This study can make an important extension in documentation and range of faunal species available within a university campus. The IVI and other quantitative values obtained for different plant species in the study coincide with the findings of Lalramenga (2006) and Lalchuanawma (2008) that were carried out within the campus. The rich floral diversity plays a key role in supporting both the avifauna and mammalian fauna in the campus. The presence of roadside plants like *Ficus benghalensis*, *F. religiosa*, *Trema orientalis*, *Lantana camara*, *Musa paradise*, and *Casia auriculata* in MZU campus can serve as good habitat and feeding grounds for birds and small mammals as reported by Lalchuanawma (2008) and Rai (2016).

The undisturbed forest patches within MZU campus appears to support a diverse group of mammalian

species. The record of 19 mammalian species in the campus is an evidence of its rich mammalian diversity, which contributes for 15% of the mammalian fauna of the state of Mizoram. Family viverridae is highly diverse in MZU campus; five species of family viverridae were also reported earlier in Mizoram (Lalthanzara 2017). Presence of felids (three species) in the campus also forms 37.5%, as eight fields were previously reported by Lalthanzara (2017) in Mizoram. Species such as Clouded Leopard, Marbled Cat, Golden Jackal, and Ferret Badger not only indicates the potential of the university campus to support the cohabitation and co-existence with mammals but also highlights the values of urban diversity.

The relative abundance index (RAI) indicates that ground dwelling birds such as Khalij Pheasant (male and females) were dominant (RAI= 13.33) and also are the prey species for carnivorous mammals. The mammalian species evident in MZU campus are very rare and may be difficult to record even in a protected forest. Dense shrub forests, tall fruiting trees, and low lying natural streams along the edges of the campus that provides ideal forest

Table 3. Mammalian species observed directly during study and their status.

	Family	Common name	Scientific name	IUCN status
1	Mustelidae	Yellow-throated Martin	<i>Martes flavigula</i>	Least Concern
2	Cervidae	Barking Deer	<i>Muntiacus muntjak</i>	Least Concern
3	Sciuridae	Pallas's Squirrel	<i>Callosciurus erythraeus</i>	Least Concern
4	Sciuridae	Himalayan Hoary-bellied Squirrel	<i>Callosciurus pygerythrus</i>	Least Concern
5	Tupaiaidae	Northern Tree Shrew	<i>Tupaia belangeri</i>	Least Concern
6	Muridae	Black Rat	<i>Rattus rattus</i>	Least Concern
7	Soricidae	House Shrew	<i>Suncus murinus</i>	Least Concern

Table 4. Avian species photo-captured during the study and their status.

	Family	Common name	Scientific name	IUCN	RAI
1	Phasianidae	Khalij Pheasant	<i>Lophura leucomelanos</i>	Least Concern	13.33
2	Muscicapidae	Blue Whistling Thrush	<i>Myophonus caeruleus</i>	Least Concern	3.33
3	Muscicapidae	Forktail	<i>Enicurus leschenaulti</i>	Least Concern	3.33
4	Strigidae	Asian-barred Owlet	<i>Glaucidium cuculoides</i>	Least Concern	6.67
5	Campephagidae	Large Cuckoo Shrike	<i>Coracina macei</i>	Least Concern	3.33
6	Caprimulgidae	Long-tailed Nightjar	<i>Caprimulgus climacurus</i>	Least Concern	3.33

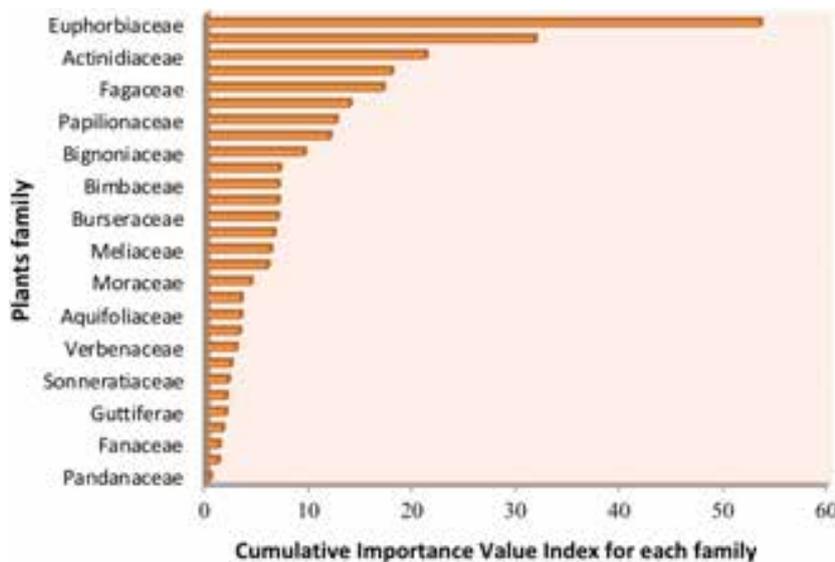


Figure 3. Plant families and their IVI values.

cover and feeding opportunities are contributing to the abundance of these species within the campus. The low resident human population (little over 2000 individuals) with a good sense of conservation value and long forest corridors are also key factors for the distribution of such majestic species within the campus. Anthropogenic activities like hunting, jhumming or shifting cultivation in the vicinity of the campus, collection of NTFP's and

the ever-increasing chain of construction works can have negative impacts on biodiversity in general and particularly on mammals of MZU campus. Information on mammalian diversity recommends that a long-term and systematic study on biodiversity profile of the MZU campus is needed.



Image 2. a—Clouded Leopard | b—Golden Jackal | c—Himalayan Palm Civet | d—Crab-eating Mongoose | e—Leopard Cat | f—Large Indian Civet | g—Small Indian Civet | h—Himalayan Hoary-bellied Squirrel | i—Common Palm Civet | j—Ferret Badger | k—Wild Boar | l—Marbled Cat.

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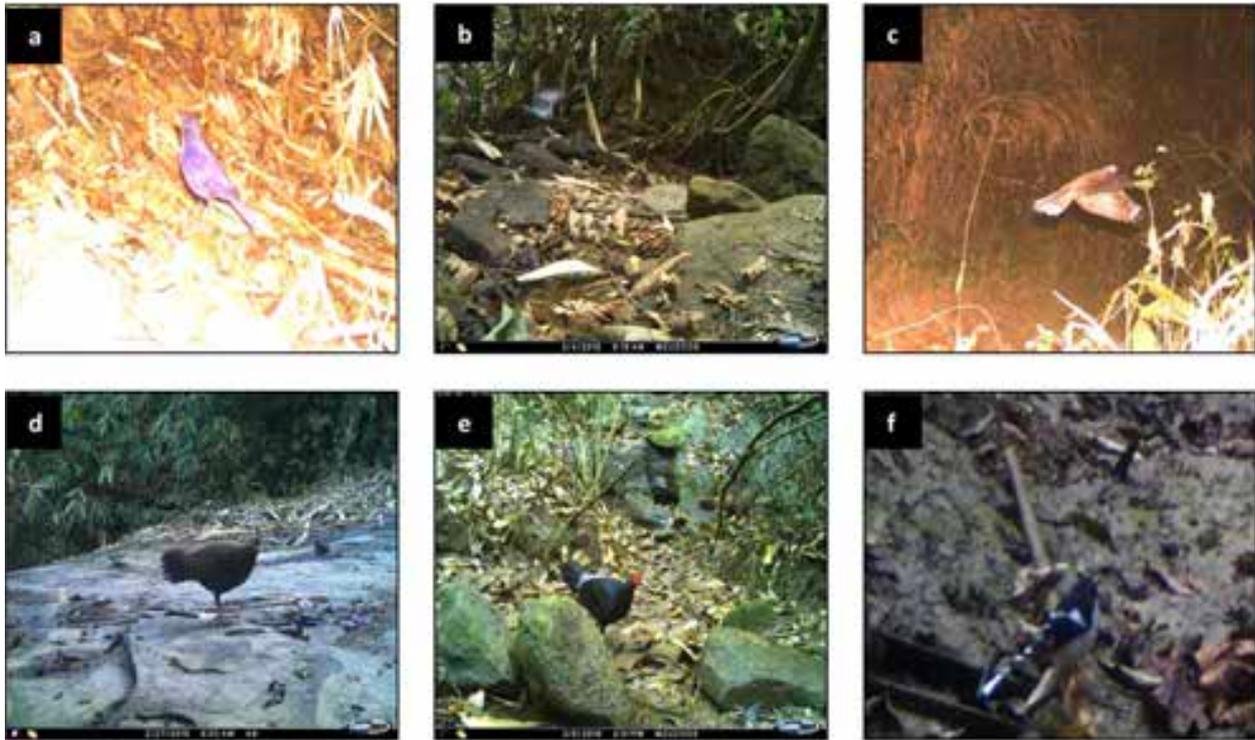


Image 3. Birds captured by camera trap: a—Blue Whistling Thrush | b—Large Cuckoo Shrike | c—Long-tailed Nightjar | d—Khalij Pheasant female | e—Khalij Pheasant male | f—Forktail.

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Distribution of *Syzygium travancoricum* Gamble (Myrtaceae), a Critically Endangered tree species from Kerala part of Western Ghats, India

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Abstract: *Syzygium travancoricum* Gamble is an endemic, Critically Endangered tree species found in marshy swamps of the Western Ghats. An assessment was conducted on existing distribution pattern and population structure of this species in Kerala. Most of the existing populations are mainly located in sacred groves and this tree is an ideal example on role of sacred groves in conserving biodiversity especially IUCN red listed species. A total of 465 individuals were recorded in which maximum number of sapling density was recorded in Kalasamala Sacred grove followed by patches of *Myristica* swamps in Kulathupuzha. Fragmentation of populations, habitat modification, over exploitation, pollution and invasion of exotics are major threats to *S. travancoricum*. Hence urgent intervention is required for restoration programmes, management of habitat and subsequently delisting this species from IUCN red list of threatened species.

Keywords: Endemic, Kalasamala, Myrtaceae, population structure, southern Western Ghats.

The genus *Syzygium* Gaertner (Myrtaceae) comprises more than 1,200 species all over the world, distributed in tropical regions of Asia, Africa, Australia, and in southwestern Pacific regions (Parnell et al. 2007; Govaerts et al. 2008). As an economically and ecologically important genus, *Syzygium* species are commercially cultivated for their fruits (*S. cumini* (L.) Skeels, *S. jambos*

(L.) Alston), timber (*S. aqueum* (Burm. f.) Alston, *S. malaccense* (L.) Merr. & L. M. Perry), or for medicinal properties (*S. aromaticum* (L.) Merr. & L. M. Perry) or as spices. In India, 54 species have been so far reported (Govaerts et al. 2008) with the highest concentration in the Western Ghats with 48 species, among which 27 species are exclusively endemic to this region (Sujanapal & Kunhikannan 2017). The forests of the Western Ghats, especially tropical wet evergreen and high-altitude Shola peaks are ideal habitats for the *Syzygium*, however, most of these species are threatened and as per the IUCN (v1.18-2019) Red List, three species are Critically Endangered, eight as Endangered, five Vulnerable and four as Least Concern. The species like *S. palghatense* and *S. courtallense* are known from single locality and *S. beddomei* and *S. rubicundum* are known only by their type collections. *Syzygium travancoricum* Gamble is an evergreen tree species endemic to the Western Ghats and it was first discovered in the swampy lowlands (altitude <65m) of Travancore by Bourdillon (1908). Later Gamble described it in 1918 in Kew Bulletin and Flora of the Presidency of Madras in 1919. According to Byng et

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al. (2015), *S. travancoricum* is conspecific with *S. stocksii* (Duthie) Gamble, as both share long petioles, elliptic leaves and highly branched inflorescences, however, Sujanapal & Kunhikannan (2017) treated this as different species and commented that detailed studies based on molecular and population data are essential for fixing the status of *S. travancoricum*. This is a medium sized tree growing up to 25m in height. The bark surface is longitudinally fissured, grayish-brown in colour and inner bark is grey. Leaves are simple, opposite, 8–16.5 × 5–8.5 cm, exstipulate and large in size; lateral nerves are 10–15 pairs, parallel, distant, with intra-marginal nerve; petiole 10–20 mm long, grooved above. Flowers are bisexual, white, mostly in axillary lax cymose corymbose with peduncle 4.5–5 (8) cm long, their branches are also long and ascending. Stamens are numerous, free, bent inwards at middle when in bud. Ovary is inferior, two-celled with many ovules. Fruit is a berry, oblong-obtuse on sides, 1 × 0.5 cm, deep violet, and pericarp juicy containing single seed (Image 1). As per IUCN Red list (2019), only 200 trees are found in the Western Ghats and treated as Critically Endangered C2a ver 2.3. It is also reported that the sacred grove of Aikad reported to harbour four individuals and another subpopulation of 15 to 20 trees has been seen at Guddrikal. Nair & Mohanan (1981) quoted that “Apparently no tree is surviving in the type locality. Recently only four trees have been spotted in a sacred grove of Aikad in Quilon District”. In Kerala, *S. travancoricum* have been reported from evergreen and semi-evergreen forests and few sacred groves in Thiruvananthapuram, Kollam, Pathanamthitta, Alapuzha, Thrissur, and Kasaragod districts (Sasidharan 2004) and locally known as Poriyal, Vathamkollimaram or Kulavetty. Ray (2011) reported this species from three groves namely, Devaravattibana (Mattigar), Kadkod Choudammabana (Aralihonda) and Choudammabana (Dugdmane) with high seedlings and sapling density in Karnataka region. It has also been reported from the freshwater Myristica swamps of Kaan forests of Uttara Kannada (Chandran et al. 2008 & 2010), Kulathupuzha (Robi 2009) and from a relic Myristica swamp at Brahma Karmali of Valpoi Taluka in Goa (Prabhugaonkar et al. 2014). From Tamil Nadu, this species was reported in five sites, viz., Pandiar, Nadugani, Tropical Gene Pool Garden, Kilnadugani and Poonoor forest areas in Nadugani Village of Gudalur Taluk in Nilgiri District (Udhayavani et al. 2013) and Megamalai Wildlife Sanctuary (Karuppusamy & Ravichandran 2016). Udhayavani et al. (2013) also recorded *S. travancoricum* from the Nilgiri District growing in association with species like *Aporosa lindleyana*, *Sterculia guttata*,

Syzygium cumini and *Litsea coriacea*. The population of *S. travancoricum* has declined considerably mainly because of over exploitation and habitat destruction. A vast stretch of virgin forest is being converted for agriculture purpose and pressure from exotics plants, grazing and forest fire also caused decline in regeneration (Udhayavani et al. 2013). The swampy wetland habitat has been widely drained and converted into paddy fields have also caused the reduction of its population, leading to a critically endangered state (IUCN 2019). The plants are under great threat due to fragmentation of populations, high rate of inbreeding and habitat specificity which accounts for poor regeneration and seed viability (Sujanapal & Kunhikannan 2017; present observation). In this context, a thorough understanding of existing population structure and composition of *S. travancoricum* will be helpful in developing a proper conservation plan.

MATERIAL AND METHODS

The forest areas and sacred groves in Kerala part of the Western Ghats were thoroughly surveyed during 2016, samples were collected and identified using relevant literature (Gamble 1919; Sasidharan 2004). All specimens were deposited at Kerala Forest Research Institute herbarium (KFRI). Several new locations were identified and total number of individuals in a population, site characteristics, important threats, and pattern of regeneration in each site were noted. Moreover, all previous literature available in Kerala part of Western Ghats were also compiled and revisited.

RESULTS

It was found that among newly identified populations in Kerala, except for the population at Kalasamala Sacred grove, others are represented with few isolated, scattered individuals. In this context, we conducted a detailed investigation on Kalasamala population including total enumeration of individuals with detailed vegetation analysis, especially on population structure, species composition and contribution of individuals. This population might be the second largest population represented with mature reproductively viable individuals distributed outside in the protected areas in the state.

Kalasamala is situated at Thrissur District (Figure 1) and the area lies between 11.128° N & 76.150° E at 30 m above the sea level with an annual average rainfall of over 2,600 mm. The major proportion (77%) of annual rainfall occurs from the south-west monsoon which usually starts in early June and continues until

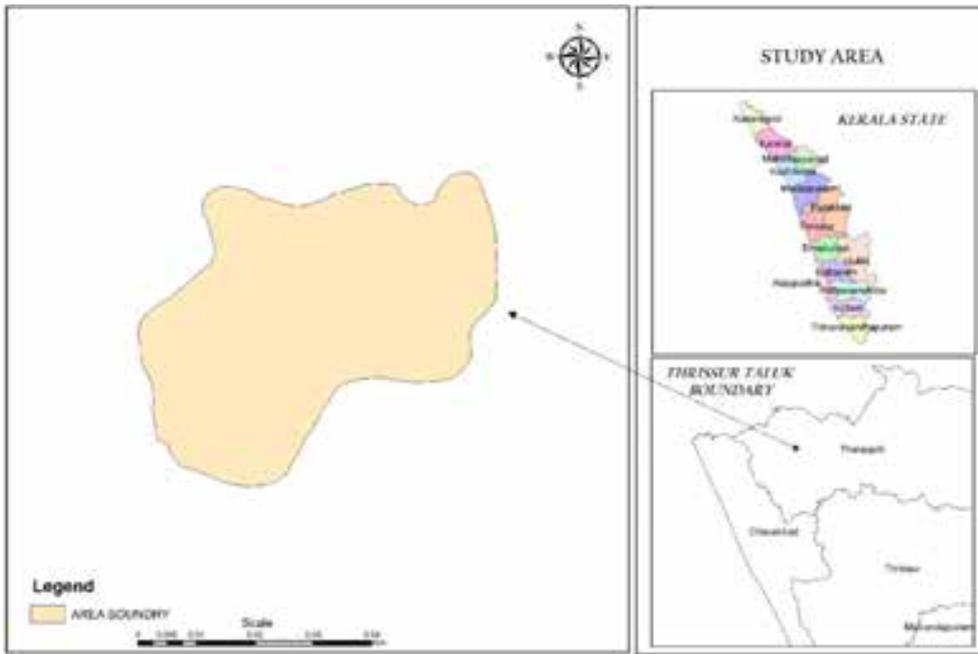


Figure 1. Study area of *Syzygium travancoricum* in Kalasamala Sacred grove.

September. The north-east monsoon starts in October and lasts until the end of November. The mean annual maximum and minimum temperature is 37°C and 19°C, respectively. There is a small sacred grove under private ownership in this area with an extent of 3.5 acres of land crammed with an isolated patch of *S. travancoricum*. For getting effective protection measures, the last author initiated legal procedures to declare this as a 'biodiversity heritage site' and the process has almost in completion. Total stem count was carried out for the population with an area of 0.21ha which in turn divided into 21 blocks of 10m × 10m size for enumeration of tree species. All individuals above 10cm GBH (Girth at Breast Height) girth at 1.37m above ground were measured and identified. Population density, frequency, basal area, IVI (Important value index) and girth class distribution were calculated.

Since *S. travancoricum* is a critically endangered species, our survey throughout the Kerala State located six new localities in which five were represented with very small population size (<10 individuals) except for Kalasamala. The results of total enumeration of individuals in Kalasamala population shows presence of 246 stems (>10.1cm) representing in 236 individuals highlights conservation significance of this area (Table 1). This population covers a total area of 0.21ha and results of vegetation structure shows that the vegetation is over dominated by *S. travancoricum* which represent 86.74% of total Important Value Index (Table 1). The basal area

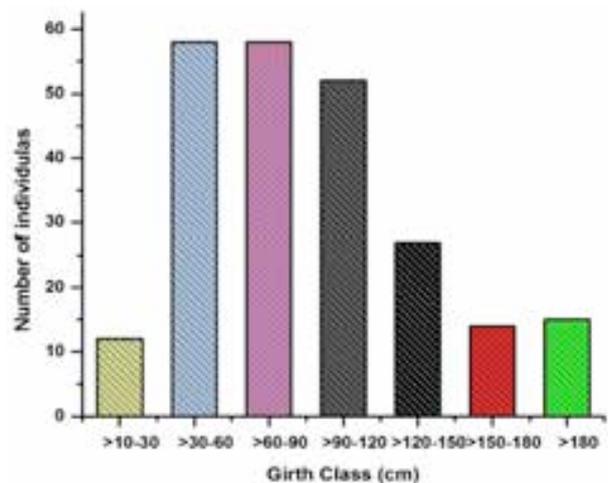


Figure 2. Girth class distribution of *Syzygium travancoricum* population in Kalasamala, Kerala, India.

of *S. travancoricum* was accounting for 109.37m²/ha. The basal area value is higher than the normal value may be due to the presence of more number of higher girth class trees and higher density value. *S. travancoricum* also represents the 100% frequency of distribution (Table 1) and associated species are *Alstonia scholaris*, *Caryota urens*, *Holigarna arnottiana*. In Uttara Kannada (Chandran et al. 2008, 2010), reported the position of *S. travancoricum* in terms of IVI as 17th and the values range from 0.1065 to 0.5483. The mean IVI of 0.3044 is slightly higher than 0.2988, the mean of the IVI recorded

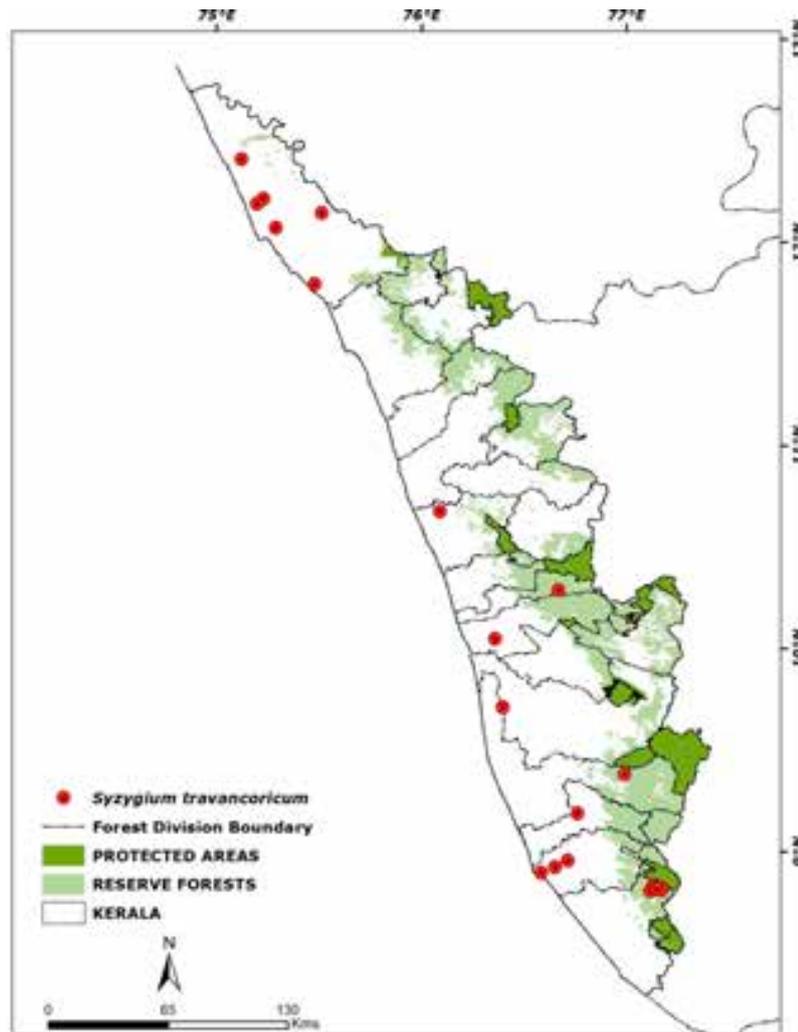


Figure 3. Distribution of *Syzygium travancoricum* in Kerala.

Table 1. Plant species and their contribution to the population dominated by *Syzygium travancoricum* in Kalasamala, Kerala, India.

Species	Basal area m ² /ha	Density (Individuals/ha)	Percentage of Frequency	IVI	Relative IVI
<i>Alstonia scholaris</i>	3.46	14	10	11.20	3.73
<i>Caryota urens</i>	6.48	14	10	13.65	4.55
<i>Holigarna arnottiana</i>	3.17	19	14	14.92	4.98
<i>Syzygium travancoricum</i>	109.37	1171	100	260.23	86.74

from transects in six swamps of southern Kerala. In Uttara Kannada (Chandran et al. 2008), reported that *Aglaia anamallayana*, *Calophyllum apetalum*, *Diospyros paniculata*, *D. pruriens*, *Dipterocarpus indicus*, *Gymnocranthera canarica*, *Holigarna grahamii*, *Hydnocarpus pentandra*, *Hopea ponga*, *Mastixia arborea*, *Myristica fatua* var. *magnifica*, and *Pinanga dicksonii* as the main associates of *S. travancoricum*. Robi (2009) surveyed different swamp patches in Kulathupuzha and

counted only 20 trees from 17 transects in six specific swamps like Emponge, Karinkurinji, Marappalam Major, PerumPadappy, PlavuChal, and Pullu Mala. Density was 11 trees/ha and sixth most important position with an IVI of 0.1198 girth class distribution shows that there was less number of individuals in the smaller girth size which means the lowest size class of 10–30cm gbh was less abundant and formed 5.08% of the total density. The highest number of individuals were record in girth



Image 1. A—habit | B—flowering branch; C—inflorescence | D & E—fruiting twigs. (© A-C - M.S. Sanil | D-E - V.B. Sreekumar)

class 30–60 and 60–90 cm gbh which is accounting for 49.14% of the total density and the density of large trees (girth >180cm) was 6.36%. The study on size class

distribution in five sites of Nadugani Village (Udhayavani et al. 2013) shows that 37% of mature stems, 50% sub-adults, 13% saplings. Here in the present study, the size

Table 2. Details of existing status *Syzygium travancoricum* reported from forest areas and sacred groves in Kerala.

	Number of individuals	Place	Current status of the population	References
1	236	Kalasamala, Kunnankulam, Thrissur	Alteration of swampy habitat, moderate regeneration, very low seedling establishment.	Present study
2	2	Konginichal Kavu, Alakkadu, Kannur	No regeneration, poor fruit set	Present study
3	2	Mazhuvancheril Sree Khandakarna Kavu, Vaikom, Kottayam	No regeneration, poor fruit set	Present study
4	2	Paliyerikavu, Kannur	No regeneration, poor fruit set	Present study
5	2	Ponnakkudam Sacred grove, Ernakulum; Kerala	No regeneration, poor fruit set	Present study
6	16	Vazhachal forests, Thrissur; Kerala	No regeneration.	Present study
7	2	Andallur Kavu, Dharmadam Kannur	No regeneration, poor fruit set	Present study
8	5	Pariyaram, Kannur	Habitat modification,	Present study
9	4	Aickad Sacred grove Kodumon, Pathanamthitta	No regeneration and unhealthy trees	IUCN, 2017
10	20	Goodrical, Kerala	Population was not traced	IUCN, 2017
11	153	Emponge, Karinkurinj, Marappalam Major, PerumPadappy, PlavuChal, Pullu Mala and other swampy patches, Kulathupuzha (Kerala)	Habitat modification, moderate regeneration and fruit set.	Robi 2009
12	8	Kuzhumathikkad Chirakkadavu Bhagavathi Temple Sacred grove, Kollam	No regeneration and unhealthy trees	Present study
13	5	Mampuzha Shri Indaliyappan Temple, Kottamkara, Kollam	Invasion of exotics, very low fruit set and regeneration	Present study
14	3	Moonamkadavu Sacred grove, Kasaragod district	Habitat modification	Present study
15	5	Asramam, Kollam	Habitat modification	Present study

class distribution of tree species shows that inadequate or poor regeneration status with most of the trees in medium to large size class and few are young trees (Figure 2) which need to be investigated thoroughly. In Kulathupuzha, the girth class distribution for *S. travancoricum* (Robi 2009) shows reduced number of individuals in the lower girth classes and the graph plotted is an almost perfect 'J' instead of the expected inverted 'J'. Whereas, in the Uttara Kannada sub-population of *S. travancoricum*, graphical representations of girth class distribution return an almost perfect inverted 'J' indicating a healthy regeneration pattern (Chandran et al. 2010). Similarly, the seedlings density reported from the sacred groves of Karnataka (Ray 2011) is 8235.44/ha (Mattigar, Devarabatti Bana), 4549.52/ha (Aralihonda, Kadkod Choudamma Bana) and 52.736/ha (Dugdmane, Choudi Bana), however, in the present study, observation of phenological pattern indicates normal fruiting and seed germination patterns for *S. travancoricum*, whereas germinated seedlings do not survive after the two-leaf seedling stage as other studies (Chandran et al. 2008; Robi 2009) and consider that swamps might be inundated at the time of fruit fall and seed germination. It was also presumed that fungal attack in the intense humid nature of swamp conditions is the major problem of the survival of *S. travancoricum*. In Kalasamala, seedlings are totally absent and the main threat to

the population is degradation of swampy habitat due to intense soil erosion from the top hill. Table 2 shows the compilation of number of individuals recorded so far from different localities of Kerala and a total of 465 individuals were counted in which the maximum number of sapling density was found in Kalasamala followed by *Myristica* patches of Kulathupuzha. Figure 3 shows that the current distribution locations of *S. travancoricum* in Kerala region. As a part of this study, we could revisit all populations reported by the previous authors (Table 2), except the population in Goodrical which is not traceable. In most cases, alteration of swampy habitat is the major factor for poor seedling establishment which gradually result in declining total number of the individuals.

In conclusion, disjunct distribution pattern, alteration of habitats, changes in landuse pattern and poor fruit set affected the distribution pattern of *S. travancoricum*. In Kalasamala, the swampy habitat of this species has altered due to deposition of silt and waste as a result of construction activities and other human interventions. The State Biodiversity Board has already taken urgent steps to declare this site as 'heritage site', and to support the temple management for protection of this highly threatened species. As a first step towards eco-restoration programmes, Kerala Forest Research Institute (KFRI) has raised sufficient seedlings, around

100 seedlings were restored at Kalasamala initially and planning to restore 2,000 seedlings in different sites of sacred groves involving officials of concerned panchayat, temple management, students as well as through public support.

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Butterflies (Lepidoptera: Rhopalocera) of the undivided Midnapore District, West Bengal, India: a preliminary report

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Abstract: A butterfly study in the biodiversity rich biogeography transition zone of the undivided Midnapore District of West Bengal was carried out from March 2014 to March 2017. A total of 98 species of butterflies under five families, 19 subfamilies, and 70 genera were documented by the study. Maximum species diversity was found in Nymphalidae family with 31 species (31.63%), followed by Lycaenidae 24 species (24.48%), Hesperidae-17 (17.34%), Pieridae 16 (16.32%), and the least by Papilionidae 10 (10.28%). Among the families, the highest common species (n=10) was recorded under the family Nymphalidae and maximum rare species (n=4) under the family Lycaenidae. Highest species richness has been recorded in post-monsoon season (97 species) followed by monsoon (78 species), winter (66 species), and pre-monsoon (47 species). During the study, five species of butterflies (*Castalius rosimon*, *Neptis jumbah*, *Discophora sondaica*, *Lethe europa*, and *Papilio clytia*) under Schedule-I, three species (*Euchrysops cnejus*, *Mahathala ameria*, and *Cepora nerissa*) under Schedule-II and three species (*Baoris farri*, *Hyarotis adrastus*, and *Euploea core*) under Schedule-IV were legally protected under the Indian Wildlife (Protection) Act, 1972. A good species to genera ratio (1.4: 1) along with 10 newly recorded species and their range extension provide information for better understanding of the ecology and distribution pattern of the butterfly fauna. The information of the study will thus help to develop conservation strategies for management of the unique bio-geographical transitional zone.

Keywords: Checklist, new records, species diversity, seasonality, biogeography transition zone.

Butterflies are highly sensitive to changes in the environment (Landres et al. 1988; Simberloff 1988), act as ecological indicators (New 1991; Pollard & Yates 1993;

Thomas 2005; Bonebrake et al. 2010), and serve as surrogate species for floral and faunal diversity (Ehrlich & Raven 1964) by improving community structure. Further, they are primary consumers in forest ecosystems (Rosenberg et al. 1986) and their diversity increases with an increase in habitat scale and vegetation structure complex (Price 1975). Therefore, the conservation of butterflies is necessary to understand their natural history and ecology for the maintenance of ecosystem health.

Transition zones are defined as the boundaries between biogeographical regions, which represent areas of biotic overlap, and being favoured by ecological and historical changes allow a combination of taxa belonging to various biotic components (Morrone 2004). As the boundaries between these regions are areas of high biotic interaction (Ruggiero & Ezcurra 2003), with increased richness and abundance as well as unique ecotonal species (Odum 1953), they deserve special attention. The undivided Midnapore District of West Bengal which lies in the transition zone of three distinct biogeographic regions of India (Deccan Peninsula: Chotta-Nagpur biotic province, Gangetic Plains: Lower Gangetic Plain biotic province, and Coast: Eastern Coastal biotic province) makes it a perfect habitat for species diversity (Payra et

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al. 2017; Samanta et al. 2017; Paria et al. 2018; Biswas et al. 2019) and acts as a connecting path of species distribution and migration from Western Ghats zone to northeastern zone (Abdulali 1949; Ali 1949; Hora 1949). Dry peninsular Sal forest and dry deciduous scrub vegetation along with different orchard plantation and agriculture practices in this region act as refuges for butterflies (Mahata et al. 2019).

Limited studies have been carried out on butterfly fauna in the undivided Midnapore District, and previous records from the region include the report of 112 species from the coastal area of Purba Medinipur (Payra et al. 2017) and 82 species from Midnapore Municipality area of West Midnapore District (Biswas et al. 2019) and remaining part was untouched. The present study aims to document butterflies in the undivided Midnapore district and prepare a checklist along with habitat ecology and management status for undertaking proper conservation action in near future.

MATERIALS AND METHODS

Study sites

The study was conducted at undivided Midnapore district (22.953°–21.610° N & 88.211°–86.564° E) of West Bengal, India. The undivided Midnapore district is now divided into three districts (Purba Medinipur, Paschim Medinipur, and Jhargram) and covers an area of 14,081km². Of which 2,971km² is under forest cover contributing 3.35% of the total geographical area of the state (FSI 2017). This region covers northern dry mixed deciduous forest of Chotta-Nagpur Plateau ecozone, Eastern Highlands moist deciduous forests of the eastern part of Eastern Ghats along with dry peninsular Sal Forest, and deciduous scrub forest of Lower Gangetic Plain and saltwater mixed forest (Champion & Seth 1968; Figure 1). The topography of this region is highly variable from undulating southwestern region to Gangetic Plains and wetlands towards the eastern side. This undivided Midnapore District is characterized by

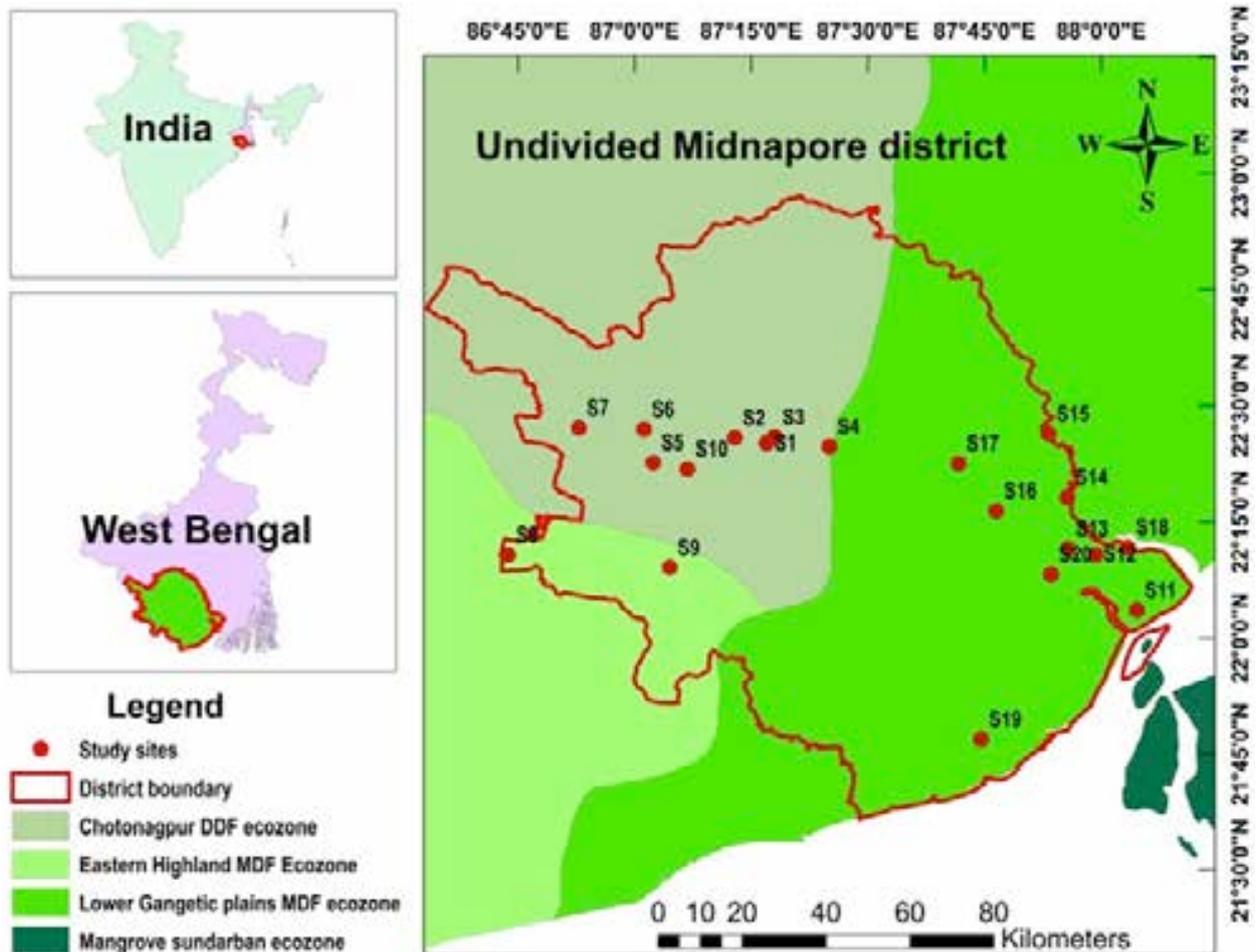


Figure 1. The map of study sites in undivided Midnapore District, West Bengal, India (DDF- Dry Deciduous Forest, MDF- Moist Deciduous Forest)

a variety of soil ranging from lateritic red soil towards the western region, alluvial along the rivers and sandy saline soil in the coastal belt. The average temperature in this region lies in between 10°C and 39°C with average annual precipitation of 1,752mm (Anonymous 2011a, b). The climate is characterized by tropical hot summer, cold winter, abundant rainfall and humidity in monsoon. Floods are quite regular in monsoon (Anonymous 2011a,b). A total of 20 different sites have been surveyed on the basis of different vegetation patterns to document the butterfly fauna. Elevations of these sites vary between 2–90 m (Table 1).

Data collection

The study was carried out over a period of three years during March 2014 to March 2017 in four distinct seasons: pre-monsoon (March to May), monsoon (June to August), post-monsoon (September to November), and winter (December to February). We adopted opportunistic sampling methods by walking through trails and recorded the species visually by using close focusing binocular (Olympus WP II, Olympus Corporation, Japan). Butterfly sampling was carried out only in good weather conditions (>10°C, <18 miles/hour wind speed and no rain) from 09.00 to 13.00 h. Butterflies were identified in the field with the help of field guides (Wynter-Blyth 1957; Kunte 2000; Kehimkar 2008). Butterflies which could not be identified in flight, were caught with butterfly-net, identified and then released safely. Photograph of the documented species was taken during the survey with the help of a DSLR camera (Nikon D5300 with Nikon 70–300 mm lens, Nikon Corporation, Japan) for further confirmations. Taxonomy and species names of butterflies were followed after Varshney & Smetacek (2015). A checklist of butterfly fauna was prepared along with their legal protection status. The status of a butterfly was decided based on the frequency of sighting of the butterfly during the total study period. Status of the butterfly was categorized under five categories: rare (R), not rare (NR), not common (NC), common (C), and very common (C). Butterflies having <20 % sightings were marked as rare butterflies, 20–40 % not rare, 40–60 % not common, 60–80 % common, and >80% very common (Payra et al. 2017; Samanta et al. 2017).

RESULTS AND DISCUSSION

A total of 98 species of butterflies under five families, 19 subfamilies and 70 genera were documented from the study area (Table 2, Image 1, 2, 3, 4 & 5). Maximum species richness was observed in the family Nymphalidae with 31 species (31.63%), followed by Lycaenidae 24

species (24.48%), Hesperidae 17 species (17.34%), Pieridae 16 species (16.32%), and least by Papilionidae 10 species (10.20%) (Table 2). The Nymphalidae family was very common in occurrence (10 species) whereas the maximum number of rare species were found in the Lycaenidae family (four species) (Table 2, Figure 3). Highest species richness was recorded in post-monsoon season (97 species), followed by monsoon (78 species), winter (66 species) and pre-monsoon (47 species) (Figure 4). Among the species recorded, 11 species are under schedule category of Indian Wildlife (Protection) Act 1972. Under Schedule-I, there are five species (*Castalius rosimon*, *Neptis jumbah*, *Discophora sondaica*, *Lethe europa*, and *Papilio clytia*), under Schedule-II, three species (*Euchrysops cnejus*, *Mahathala ameria*, and *Cepora nerissa*) and under Schedule-IV three species (*Baoris farri*, *Hyarotis adrastus*, and *Euploea core*) (Table 2). Five species are placed under Least Concern (LC) category and the rest 93 species are not assessed (IUCN 2020) (Table 2, Image 1, 2, 3, 4 & 5).

Among the recorded 17 species of Hesperidae (Image 1), *Ampittia dioscorides* was very common, whereas *Badamia exclamationis* and *Matapa aria* were rare. *Baoris farri* and *Pelopidas subochracea* were recorded for the first time from the region. Among 24 species of Lycaenidae family (Image 2), *Arhopala atrax*, *Iraota timoleon*, *Mahathala ameria*, *Rathinda amor*, and *Virachola isocrates* were found to be rare. *Zizina otis* and *Arhopala atrax* were recorded for the first time from this region. *Iraota timoleon* and *Rathinda amor* were recorded only during post monsoon period. Out of 31 species recorded under Nymphalidae (Image 3), *Ariadne ariadne*, *A. merione*, *Danaus chrysippus*, *Danaus genutia*, *Euploea core*, *Acraea violae*, *Phalanta phalantha*, *Junonia almana*, *J. iphita*, and *J. lemonias* were very common whereas *Tirumala septentrionis* and *Discophora sondaica* were rare and *T. septentrionis*, *Phaedyma columella*, and *Ypthima baldus* were recorded for the first time from this region. Among 10 species under Papilionidae family (Image 4), *Papilio demoleus* was very common, whereas *Papilio crino* was rare and *Graphium nomius* was newly recorded from this region. Out of 16 species under Pieridae family (Image 5), *Delias eucharis* was very common and two species (*Eurema laeta* and *E. sari*) were recorded for the first time from undivided Midnapore region (Table 2, 3).

In our study, the presence of *Symphaedra nais*, *Danaus melanippus indicus* and *Ypthima baldus* showed the species range extension from central India towards eastern India. Our study also recorded several Eastern Ghats species such as *Pseudozizeeria maha*,

Table 1. Location and Habitat conditions of the study sites.

Site No.	Site name	GPS Location	Elevation (m)	Habitat type	Vegetation type (dominant)	Management status
1	Gopegarh Eco-park (S1)	22.419°N 87.283°E	52.43	Managed park surrounded by natural vegetation	Non-native flowering plants and <i>Anacardium occidentale</i>	Managed by park authority, no grazing
2	Gurguripal Eco-park (S2)	22.430°N 87.215°E	48.77	Degraded park with natural vegetation	<i>Shorea robusta</i> and <i>Eucalyptus tereticornis</i>	Moderately anthropogenic activity, no grazing
3	VU Campus, Midnapore (S3)	22.432°N 87.299°E	58.22	Natural vegetation surrounded by Non-native flowering plants	<i>Anacardium occidentale</i> and <i>Eucalyptus tereticornis</i>	Least management, no grazing
4	Pathra (S4)	22.412°N 87.418°E	33.18	Natural rural vegetation with agriculture field	<i>Eucalyptus tereticornis</i> , <i>Bambusa</i> spp.	Grazing along with high human interference
5	Amlachati Vesaj Udyan (S5)	22.377°N 87.039°E	89	Managed herbal garden surrounded by sal forest.	Medicinal plants and <i>Shorea robusta</i>	Highly managed, no grazing
6	Jhargram Mini Zoo (S6)	22.449°N 87.020°E	81.38	Managed mini zoo in sal forest	<i>Shorea robusta</i>	Managed, no grazing
7	Chilkigarh (S7)	22.451°N 86.881°E	73.85	Rural area	<i>Shorea robusta</i>	Moderate grazing with anthropogenic activity
8	Jhilli Pakhiralaya (S8)	22.179°N 86.729°E	66.45	Forest area along marshy water body	<i>Shorea robusta</i>	No management, moderate grazing with anthropogenic activity
9	Kodopal fruit garden (S9)	22.151°N 87.075°E	35.66	Fruit orchards along with ornamental flowering plant	<i>Mangifera indica</i> , <i>Psidium guajava</i> , <i>Citrus</i> spp., <i>Zizyphus vulgaris</i>	Managed. No grazing and anthropogenic activity
10	Banantika (S10)	22.363°N 87.113°E	79.55	Fruit orchards	<i>Mangifera indica</i> , <i>Psidium guajava</i> , <i>Phyllanthus emblica</i>	Poorly managed. grazing and anthropogenic activities limited
11	Haldia (S11)	22.059°N 88.077°E	07.32	Park Garden	Ornamental flowering plants	Managed park
12	Mahishadal (S12)	22.178°N 87.988°E	06.10	Shrub land with natural fruit orchards	<i>Mangifera indica</i>	Rural area, No management
13	Nandakumar (S13)	22.192°N 87.928°E	04.87	Shrub land beside rail tracks	<i>Lantana camara</i>	No management
14	Tamluk (S14)	22.302°N 87.927°E	06.10	Natural Vegetation	Ornamental flowering plants	Urban area, no management
15	Kolaghat (S15)	22.439°N 87.887°E	09.45	Shrub land and agriculture land	<i>Hibiscus rosa-sinensis</i> , <i>Tabernaemontana divaricate</i>	Rural area, less grazing and anthropogenic activity
16	Moyna (S16)	22.273°N 87.775°E	04.57	Bogs and Marshy area	Grass and <i>Eucalyptus tereticornis</i>	Rural area, less grazing and anthropogenic activity
17	Khirai (S17)	22.375°N 87.694°E	06.10	Riverside agriculture land	<i>Oryza sativa</i> , <i>Tagetes</i> spp.	Rural area, less grazing and anthropogenic activity
18	Geonkhali (S18)	22.195°N 88.057°E	02.13	Riverside Park	Natural shrub and ornamental flower	Rural park, less management
19	Contai (S19)	21.782°N 87.742°E	07.32	Shrub land and agricultural fields	<i>Lantana Camara</i> and <i>Oryza sativa</i>	Rural area, less grazing and anthropogenic activity
20	Narghat (S20)	22.137°N 87.892°E	06.10	Riverside semi urban park	<i>Casuarina equisetifolia</i>	Semi urban area, less grazing and anthropogenic activity

Anthropogenic Activities (AA) are based upon number of human encounters and grazing seen during 30 minutes transect walk. Least: 0–1 time, moderately AA: 3–4 times, highly AA: 8–10 times encounter

Junonia orithya, *Catopsilia pyranthe*, *Cepora nerissa*, *Spindasis vulcanus*, *Papilio demoleus*, *Arhopala atrax*, *Tarucus nara*, *Papilio hector*, and *Graphium nomius*. The study further revealed the presence of *Papilio crino* known from northeastern India as well as species such as *Colotis amata*, *Catopsilia pomona*, *Catopsilia pyranthe*, *Cepora nerissa*, *Danaus chrysippus*, *Tirumala limniace*, *Zizeeria karsandra*, and *Catochrysops strabo* known from northwestern India (Kehimkar 2008). This species checklist represents the species distribution

information for better understanding the butterfly fauna and making conservation strategies for similar biogeographical ecozones.

CONCLUSION

The present study provides a checklist of butterfly fauna in a biogeographically transitional region of undivided Midnapore District; the first attempt at a relatively larger scale documentation of butterflies of this region. A record of 98 species of butterfly along

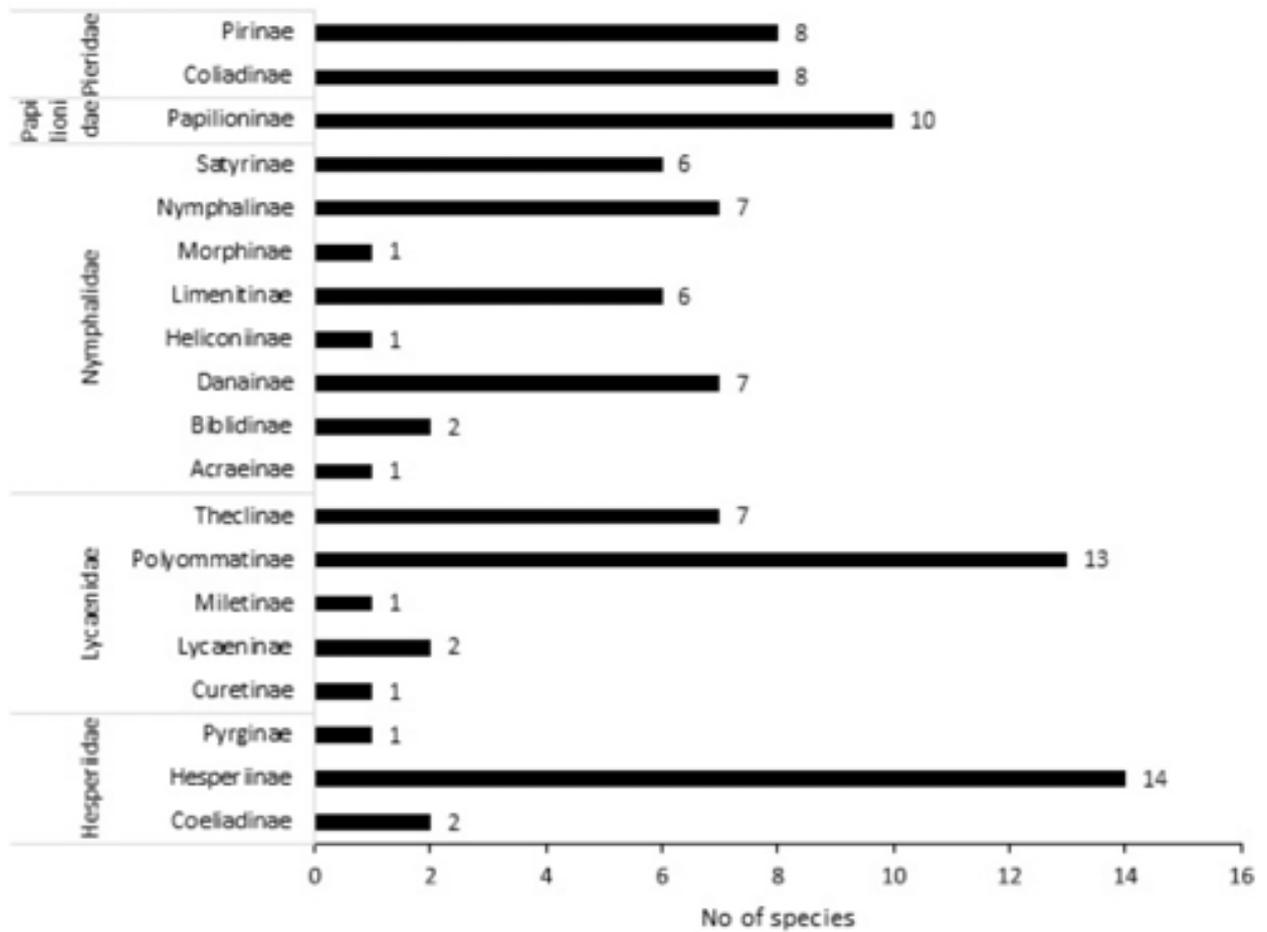


Figure 2. Subfamily wise recorded butterfly species in study sites.

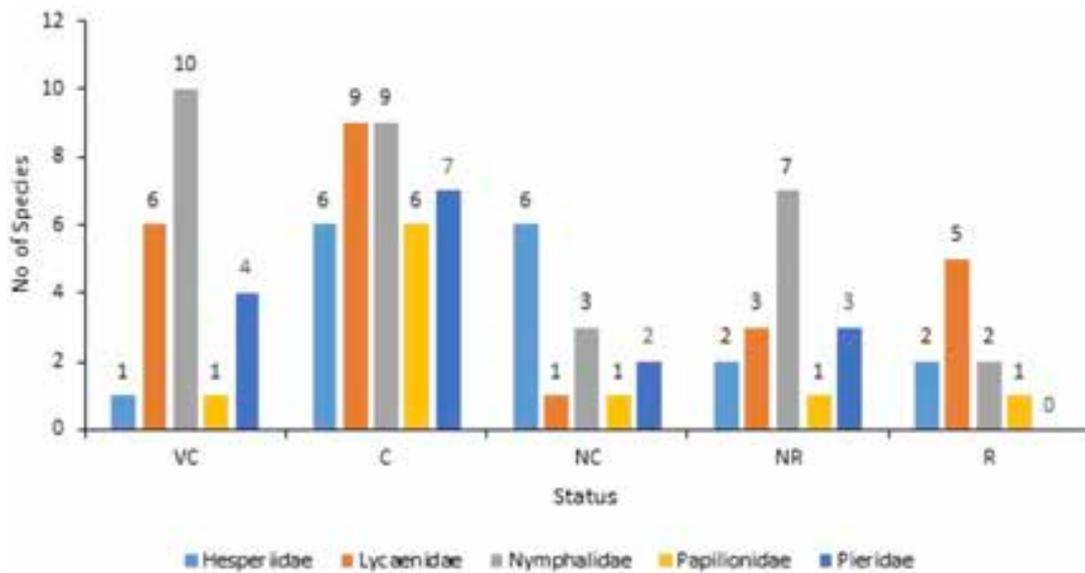


Figure 3. Status of butterfly in the study: VC—Very Common | C—Common | NC—Not common | NR—Not Rare | R—Rare.

Table 2. Checklist of butterfly fauna recorded in the study.

	Scientific name	Common name	Status	Flight period	IUCN/IW(P)A
Family HesperIIDae					
Subfamily Coeliadinae					
1	<i>Badamia exclamationis</i> (Fabricius, 1775)	Brown Awl	R	2, 3	NE
2	<i>Hasora chromus</i> (Cramer, [1780])	Common Banded Awl	NR	1, 2, 3, 4	NE
Subfamily HesperIinae					
3	<i>Suastus gremius</i> (Fabricius, 1798)	Indian Palm Bob	C	1, 3	NE
4	<i>Ampittia dioscorides</i> (Fabricius, 1793)	Bush Hopper	VC	3, 4	NE
5	<i>Baoris farri</i> (Moore, 1878)*	Paintbrush Swift	NC	2, 3, 4	NE/ Sch IV
6	<i>Borbo cinnara</i> (Wallace, 1866)	Rice Swift	NC	2, 3, 4	NE
7	<i>Erionota thrax</i> (Linnaeus, 1767)	Palm Redeye	NR	2, 3	NE
8	<i>Hyarotis adrastus</i> (Stoll, [1780])	Tree Flitter	NC	1, 3	NE/ Sch IV
9	<i>Iambrix salsala</i> (Moore, [1866])	Chestnut Bob	C	1, 2, 3, 4	NE
10	<i>Matapa aria</i> (Moore, [1866])	Common Redeye	R	1, 4	NE
11	<i>Parnara guttatus</i> (Bremer & Grey, [1852])	Straight Swift	NC	2, 3, 4	NE
12	<i>Pelopidas subochracea</i> (Moore, 1878)*	Moore's Swift	C	1, 2, 3, 4	NE
13	<i>Pelopidas thrax</i> (Huebner, [1821])	Small Branded Swift	C	2, 3, 4	NE
14	<i>Telicota colon</i> (Fabricius, 1775)	Common Palm Dart	NC	2, 3, 4	NE
15	<i>Telicota bambusae</i> (Moore, 1878)	Dark Palm Dart	C	3, 4	NE
16	<i>Udaspes folus</i> (Cramer, [1775])	Grass Demon	NC	2, 3, 4	NE
Subfamily Pyrginae					
17	<i>Spialia galba</i> (Fabricius, 1793)	Indian Skipper	C	1, 2, 3, 4	NE
Family Lycaenidae					
Subfamily Curetinae					
18	<i>Curetis thetis</i> (Drury, 1773)	Indian Sunbeam	C	3, 4	NE
Subfamily Lycaeninae					
19	<i>Pseudozizeeria maha</i> (Kollar, [1844])	Pale Grass Blue	VC	1, 2, 3, 4	NE
20	<i>Zizina otis</i> (Fabricius, 1787)*	Lesser Grass Blue	VC	1, 2, 3, 4	NE
Subfamily Miletinae					
21	<i>Spalgis epius</i> (Westwood, 1852)	Common Apefly	NR	2, 3	NE
	Subfamily Polyommatinae				NE
22	<i>Anthene emolus</i> (Godart, 1824)	Common Ciliate Blue	C	2, 3, 4	NE
23	<i>Castalius rosimon</i> (Fabricius, 1775)	Common Pierrot	VC	1, 2, 3	NE/ Sch I (Part IV)
24	<i>Catochrysops strabo</i> (Fabricius, 1793)	Forgetmenot	C	1, 2, 3, 4	NE
25	<i>Chilades lajus</i> (Stoll, [1780])	Lime Blue	C	2, 3, 4	NE
26	<i>Euchrysops cnejus</i> (Fabricius, 1798)	Gram Blue	NR	2, 3, 4	NE/ Sch II (Part II)
27	<i>Jamides bochus</i> (Stoll, [1782])	Dark Cerulean	C	1, 2, 3, 4	NE
28	<i>Jamides celeno</i> (Cramer, [1775])	Common Cerulean	C	3, 4	NE
29	<i>Luthrodes pandava</i> (Horsfield, [1829])	Plains Cupid	VC	1, 2, 3, 4	NE
30	<i>Neopithecops zalmora</i> (Butler, 1870)	Quaker	VC	2, 3, 4	NE
31	<i>Prosotas dubiosa</i> (Semper, [1879])	Tailless Lineblue	C	1, 2, 3	NE
32	<i>Tarucus nara</i> (Kollar, 1848)	Striped/ Rounded Pierrot	C	2, 3, 4	NE

	Scientific name	Common name	Status	Flight period	IUCN/IW(P)A
33	<i>Zizeeria karsandra</i> (Moore, 1865)	Dark Grass Blue	C	1,2,3,4	NE
34	<i>Zizula hylax</i> (Fabricius, 1775)	Tiny Grass Blue	VC	1,2,3,4	NE
Subfamily Theclinae					
35	<i>Arhopala atrax</i> (Hewitson, 1862)*	Indian Oakblue	R	2,3	NE
36	<i>Iraota timoleon</i> (Stoll, [1790])	Silverstreak Blue	R	3	NE
37	<i>Mahathala ameria</i> (Hewitson, 1862)	Falcate Oakblue	R	3	NE/ Sch II (Part II)
38	<i>Rapala manea</i> (Hewitson, 1863)	Slate Flash	NR	2,3	NE
39	<i>Rathinda amor</i> (Fabricius, 1775)	Monkey Puzzle	R	3	NE
40	<i>Spindasis vulcanus</i> (Fabricius, 1775)	Common Silverline	NC	2, 3	NE
41	<i>Virachola isocrates</i> (Fabricius, 1793)	Common Guava Blue	R	1, 3, 4	NE
Family Nymphalidae					
Subfamily Biblidinae					
42	<i>Ariadne</i> (Linnaeus, 1763)	Angled Castor	VC	1, 2,3,4	NE
43	<i>Ariadne merione</i> (Cramer, [1777])	Common Castor	VC	1, 2, 3, 4	NE
	Subfamily Danainae				NE
44	<i>Danaus chrysippus</i> (Linnaeus, 1758)	Plain Tiger	VC	1,2,3,4	NE
45	<i>Danaus genutia</i> (Cramer, [1779])	Common Tiger/ Striped Tiger	VC	1,2,3,4	NE
46	<i>Danaus melanippus indicus</i> (Fruhstorfer, 1899)	White Tiger	C	2,3,4	NE
47	<i>Euploea core</i> (Cramer, [1780])	Common Crow	VC	1, 2, 3, 4	LC/Sch IV
48	<i>Euploea klugii</i> kollari (C. & R. Felder, 1865)	King Crow	C	3,4	NE
49	<i>Tirumala limniace</i> (Cramer, [1775])	Blue Tiger	NC	1, 2, 3	NE
50	<i>Tirumala septentrionis</i> (Butler, 1874)*	Dark Blue Tiger	R	2,3	NE
Subfamily Acraeinae					
51	<i>Acraea violae</i> (Fabricius, 1793)	Tawny Coster	VC	1, 2, 3,4	NE
Subfamily Heliconiinae					
52	<i>Phalanta phalantha</i> (Drury, [1773])	Common Leopard	VC	1, 2,3, 4	NE
Subfamily Limenitinae					
53	<i>Euthalia aconthea</i> (Cramer, [1777])	Common Baron	NR	2, 3	NE
54	<i>Moduza procris</i> (Cramer, [1777])	Commander	C	2, 3	NE
55	<i>Neptis hylas</i> (Linnaeus, 1758)	Common Sailer	C	2, 3, 4	NE
56	<i>Neptis jumbah</i> Moore, [1858]	Chestnut-streaked Sailer	NR	2, 3	NE/ Sch I (Part IV)
57	<i>Phaedyma columella</i> (Cramer, [1780])*	Short-banded Sailer	NC	2,3,4	NE
58	<i>Symphhaedra nais</i> (Forster, 1771)	Baronet	C	1, 2, 3	NE
Subfamily Morphinae					
59	<i>Discophora sondaica</i> Boisduval, 1836	Common Duffer	R	3	NE/ Sch I (Part IV)
Subfamily Nymphalinae					
60	<i>Hypolimnas bolina</i> (Linnaeus, 1758)	Great Eggfly	c	1,2,3,4	NE
61	<i>Junonia almana</i> (Linnaeus, 1758)	Peacock Pansy	VC	2,3,4	LC
62	<i>Junonia atlites</i> (Linnaeus, 1763)	Grey Pansy	c	2,3,4	NE
63	<i>Junonia hierta</i> (Fabricius, 1798)	Yellow Pansy	NC	2,3	LC
64	<i>Junonia iphita</i> (Cramer, [1779])	Chocolate Pansy	VC	1, 2, 3, 4	NE
65	<i>Junonia lemanias</i> (Linnaeus, 1758)	Lemon Pansy	VC	1,2,3,4	NE
66	<i>Junonia orithya</i> (Linnaeus, 1758)	Blue Pansy	NR	2, 3	NE

	Scientific name	Common name	Status	Flight period	IUCN/IW(P)A
Subfamily Satyrinae					
67	<i>Elymnias hypermnestra</i> (Linnaeus, 1763)	Common Palmfly	NR	2, 3, 4	NE
68	<i>Lethe europa</i> (Fabricius, 1775)	Bamboo Treebrown	NR	3,4	NE/Sch I (Part IV)
69	<i>Melanitis leda</i> (Linnaeus, 1758)	Common Evening Brown	C	1, 2, 3, 4	NE
70	<i>Mycalesis perseus</i> (Fabricius, 1775)	Common Bush Brown	NR	2,3,4	NE
71	<i>Ypthima baldus</i> (Fabricius, 1775)*	Common Fivering	NR	2, 3, 4	NE
72	<i>Ypthima huebneri</i> Kirby, 1871	Common Fourring	C	1,2,3,4	NE
Family Papilionidae					
Subfamily Papilioninae					
73	<i>Graphium agamemnon</i> (Linnaeus, 1758)	Tailed Jay	C	1,2,3	NE
74	<i>Graphium doson</i> (C. & R. Felder, 1864)	Common Jay	C	1, 2, 3	NE
75	<i>Graphium nomius</i> (Esper, 1799)*	Spot Swordtail	C	1,2,3	NE
76	<i>Pachliopta aristolochiae</i> (Fabricius, 1775)	Common Rose	C	1, 2, 3, 4	NE
77	<i>Papilio clytia</i> Linnaeus, 1758	Common Mime	C	2, 3	NE/Sch I (Part IV)
78	<i>Papilio crino</i> Fabricius, 1793	Common Banded Peacock	R	3, 4	NE
79	<i>Papilio demoleus</i> Linnaeus, 1758	Lime Butterfly	VC	1,2,3,4	NE
80	<i>Papilio hector</i> (Linnaeus, 1758)	Crimson Rose	NR	1,2,3,4	NE
81	<i>Papilio polymnestor</i> Cramer, [1775]	Blue Mormon	NC	2,3,4	NE
82	<i>Papilio polytes</i> Linnaeus, 1758	Common Mormon	C	2, 3, 4	NE
Family Pieridae					
Subfamily Coliadinae					
83	<i>Catopsilia pomona</i> (Fabricius, 1775)	Common Emigrant	C	1, 2, 3, 4	NE
84	<i>Catopsilia pyranthe</i> (Linnaeus, 1758)	Mottled Emigrant	VC	1,2,3,4	NE
85	<i>Eurema andersoni</i> Moore, 1886	One-spot Grass Yellow	C	1,2,3,4	LC
86	<i>Eurema blanda</i> (Boisduval, 1836)	Three-spot Grass Yellow	VC	1,2,3,4	NE
87	<i>Eurema brigitta</i> (Stoll, [1780])	Small Grass Yellow	VC	1,2,3,4	LC
88	<i>Eurema hecabe</i> (Linnaeus, 1758)	Common Grass Yellow	C	2, 3, 4	NE
89	<i>Eurema laeta</i> (Boisduval, 1836)*	Spotless Grass Yellow	C	1,2,3,4	NE
90	<i>Eurema sari</i> (Horsfield)*	Chocolate Grass Yellow	C	2, 3, 4	NE
Subfamily Pirinae					
91	<i>Appias libythea olferna</i> (Swinhoe, 1890)	Striped Albatross	NR	2,3	NE
92	<i>Cepora nerissa</i> (Fabricius, 1775)	Common Gull	C	2, 3	NE/Sch II (Part II)
93	<i>Colotis amata</i> (Cramer, 1775)	Small Salmon Arab	NC	1,2,4	NE
94	<i>Delias eucharis</i> (Drury, 1773)	Common Jezabel	VC	1, 2, 3, 4	NE
95	<i>Ixias pyrene</i> (Linnaeus, 1764)	Yellow Orange Tip	NC	1,2,3	NE
96	<i>Leptosia nina</i> (Fabricius, 1793)	Psyche	C	1,2,4	NE
97	<i>Pareronia valeria</i> (Cramer, [1776])	Common Wanderer	NR	2,3	NE
98	<i>Pieris canidia</i> (Linnaeus, 1768)	Indian Cabbage White	NR	2,3	NE

VC—very common | C—common | NC—not common | NR—not rare | R—rare | 1—pre-monsoon | 2—monsoon | 3—post-monsoon | 4—winter | NE—not evaluated | LC—Least Concern | Sch—schedule.

* indicates newly recorded species from this region

Table 3. List of newly recoded species along with their distribution and legal protection status.

	New records	Distribution ^a	Legal protection ^b	Reference
Family Hesperidae				
1	<i>Baoris farri</i>	West Bengal (Alipurduar, South 24 Parganas, Bankura, Purulia district), NEI, A & N Is., SI, CI	NE/ Sch IV	Saji & Manoj (2020)
2	<i>Pelopidas subochracea</i>	West Bengal (northern Bengal, southwestern Bengal), NEI, SI, CI	NE	Saji (2020)
Family Lycaenidae				
3	<i>Zizina otis</i>	India	NE	Saji & Mahajan (2020)
4	<i>Arhopala atrax</i>	West Bengal (Alipurduar, Bankura, Purulia district) NEI, CI, NI	NE	Anonymous (2020a)
Family Nymphalidae				
5	<i>Tirumala septentrionis</i>	NEI, NI, SI, CI	NE	Saji et al. (2020)
6	<i>Phaedyma columella</i>	West Bengal (Alipurduar, Bardhaman district), NEI, NI, SI, CI, A&N Is	NE	Churi & Bagli (2020)
7	<i>Ypthima baldus</i>	West Bengal (Northern Bengal to southern Bengal), NEI, NI, CI, SI	NE	Ogale & Saji (2020)
Family Papilionidae				
8	<i>Graphium nomius</i>	West Bengal (Bardhaman, Bankura, Purulia district), SI, CI, NEI	NE	Churi (2020)
Family Pieridae				
9	<i>Eurema laeta</i>	West Bengal (Bankura, Purulia district), SI, CI, NEI, NI	NE	Bhakare & Bhagwat (2020)
10	<i>Eurema sari</i>	NEI, SI	NE	Anonymous (2020b)

^a Emphasis on regional distribution along with National level distribution (NEI: Northeastern India, NI: Northern India, SI: Southern India, CI: Central India, A&N Is: Andaman & Nicobar island of newly recoded species in this study.

^b Legal protection of newly recorded butterfly species based on IUCN Red list (NE: Not Evaluated)/IW(P)A (Sch IV: Schedule IV).

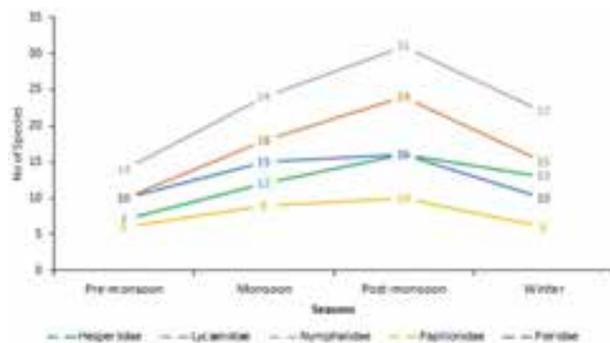


Figure 4. Flight period of butterfly fauna during entire study.

with good species to genera ratio (1.4: 1) indicates good butterfly diversity in the region. Presence of 11 legally protected species under the Indian Wildlife (Protection) Act, 1972 and 10 newly recorded species from this region draws attention towards the conservation priorities of this flagship species. The present comprehensive checklist of the butterfly fauna along with their habitat types and seasonality provide information for better understanding the ecology and distribution pattern of the butterfly fauna so that it can help to develop specific conservation measures for this unique biogeographical transitional zone.

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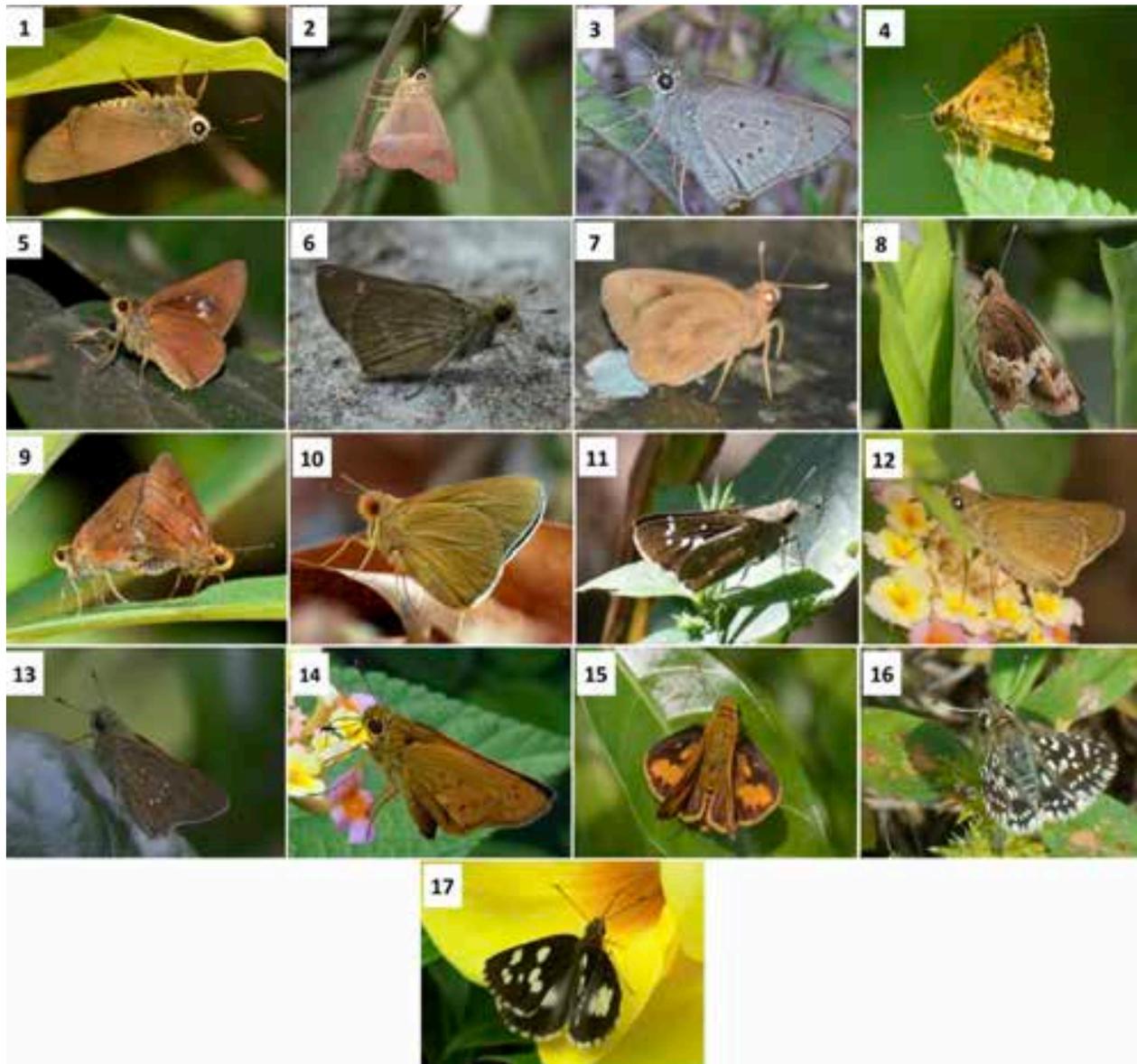


Image 1. Species of Hesperidae family recorded in the present study: 1—*Badamia exclamationis* | 2—*Hasora chromus* | 3—*Suastus gremius* | 4—*Suastus gremius* | 5—*Baoris farri* * | 6—*Borbo cinnara* | 7—*Erionota thrax* | 8—*Hyarotis adrastus* | 9—*Iambrix salsala* | 10—*Matapa aria* | 11—*Parnara guttatus* | 12—*Pelopidas subochracea* * | 13—*Pelopidas thrax* | 14—*Telicota colon* | 15—*Telicota bambusae* | 16—*Udaspes folus* | 17—*Spialia galba*.

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Image 2. Species of Lycaenidae family recorded in the present study: 18—*Curetis thetis* | 19—*Pseudozizeeria maha* | 20—*Zizina otis* * | 21—*Spalgis epius* | 22—*Anthene emolus* | 23—*Castalius rosimon* | 24—*Catochrysops strabo* | 25—*Chilades lajus* | 26—*Euchrysops cnejus* | 27—*Jamides bochus* | 28—*Jamides celeno* | 29—*Luthrodes pandava* | 30—*Neopithecops zalmora* | 31—*Prosotas dubiosa* | 32—*Tarucus nara* | 33—*Zizeeria karsandra* | 34—*Zizula hylax* | 35—*Arhopala atrax* * | 36—*Iraota timoleon* | 37—*Mahathala ameria* | 38—*Rapala manea* | 39—*Rathinda amor* | 40—*Spindasis vulcanus* | 41—*Virachola Isocrates*.



Image 3. Species of Nymphalidae family recorded in the present study: 42—*Ariadne ariadne* | 43—*Ariadne merione* | 44—*Danaus chrysippus* | 45—*Danaus genutia* | 46—*Danaus melanippus indicus* | 47—*Euploea core* | 48—*Euploea klugii kollari* | 49—*Tirumala limniace* | 50—*Tirumala septentrionis* * | 51—*Acraea violae* | 52—*Phalanta phalantha* | 53—*Euthalia aconthea* | 54—*Moduza procris* | 55—*Neptis hylas* | 56—*Neptis jumbah* | 57—*Phaedyma columella* * | 58—*Symphhaedra nais* | 59—*Discophora sondaica* | 60—*Hypolimnas bolina* | 61—*Junonia almanac* | 62—*Junonia atlites* | 63—*Junonia hierta* | 64—*Junonia iphita* | 65—*Junonia lemonias* | 66—*Junonia orithya* | 67—*Elymnias hypermnestra* | 68—*Lethe europa* | 69—*Melanitis leda* | 70—*Mycalesis perseus* | 71—*Ypthima baldus* * | 72—*Ypthima huebneri*.



Image 4. Species of Papilionidae family recorded in the present study: 73—*Graphium Agamemnon* | 74—*Graphium doson* | 75—*Graphium nomius** | 76—*Pachliopta aristolochiae* | 77—*Papilio clytia* | 78—*Papilio crino* | 79—*Papilio demoleus* | 80—*Papilio hector* | 81—*Papilio polymnestor* | 82—*Papilio polytes*.



Image 5. Species of Pieridae family recorded in the present study: 83—*Catopsilia Pomona* | 84—*Catopsilia pyranthe* | 85—*Eurema andersoni* | 86—*Eurema blanda* | 87—*Eurema brigitta* | 88—*Eurema hecabe* | 89—*Eurema laeta** | 90—*Eurema sari** | 91—*Appias libythea olferna* | 92—*Cepora nerissa* | 93—*Colotis amata* | 94—*Delias eucharis* | 95—*Ixias pyrene* | 96—*Leptosia nina* | 97—*Pareronia valeria* | 98—*Pieris canidia*.

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Occurrence of *Corica soborna* Hamilton, 1822 (Clupeiformes: Clupeidae) in the Godavari basin, India

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Abstract: We record for the first time, *Corica soborna* Hamilton, 1822 from Godavari River, based on a single specimen collected from stream near Talai Village, Kumaram Bheem Asifabad District of Telangana State. In addition to a detailed description of its morphological characters, we also provide details on distribution, habitat and threats to the species.

Keywords: Bejjur, Clupeoid, freshwater fish, Pranahita River sub-basin, Telangana State.

Approximately 420 species of clupeoids are known from around the world, distributed in marine, estuarine and freshwater habitats (Fricke et al. 2020). In India, more than 80 species of clupeoid fishes have been recorded (Froese & Pauly 2019), of which around 34 species are distributed in freshwaters and estuaries (Jayaram 2010; Froese & Pauly 2019). The Ganges River Sprat, *Corica soborna* Hamilton, 1822 is a small freshwater clupeid primarily known from the Ganges river basin, with a record from the Bharathapuzha River basin in Kerala (Bijukumar & Sushama 2000). Other records are from Indonesia, Malaysia, and Singapore in southeastern Asia (Hardenberg 1936; Herre &

Myers 1937; Whitehead 1967, 1972). This species was described from Aiyargunj in erstwhile Bengal Presidency (Hamilton 1822; also see Britz 2019), at the confluence of Punarbhaba River with Mahananda River, located in current day Bangladesh (Bhattacharya 1974). Previous records of the species are mostly from the Ganges and its tributaries – Dehri-on-Sone on river Sone in Bihar (Motwani & David 1957), Goribaba on river Ken in Uttar Pradesh (Srivastava et al. 1970), Dighwara on the Ganges in Bihar (Karamchandani 1962), from several other locations on the Ganges in Patna and Bhagalpur districts in Bihar, and Sahibganj district in Jharkhand (Kamal & Ahsan 1978). The only record of this species in India, outside the Ganges River system is from the Ponnani Estuary, Kerala (Bijukumar & Sushama 2000). Whitehead (1972) inadvertently mentioned ‘Mahanadi’ river as type locality instead of ‘Mahananda’ river from where the species was described by Hamilton (1822). It has so far not been recorded from the Godavari River basin (Talwar & Jhingran 1991; Krishnan & Mishra 2001; Jayaram 1981, 2010), or from anywhere in Telangana

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State (Barman 1993; Chaudhry 2010; Jayaram 2010). Through this communication, we report for the first time, the occurrence of *Corica soborna* in the Godavari River basin in Telangana State, India.

MATERIAL AND METHODS

While conducting ichthyological surveys in Telangana State, we collected a single specimen of a fish that resembled *C. soborna* with a dragnet operated by a local fisher at Talai Village, Kumaram Bheem Asifabad District. The voucher specimen was photographed, labeled, and fixed in 4% formalin (Jayaram 2010), and deposited in the Natural History Museum, Department of Zoology, University College of Science, Osmania University, Hyderabad, Telangana State, India (NHMOU). Morphometric measurements were taken using Mitutoyo digital calipers following Jayaram (2010) and Armbuster (2012), and the fish was subsequently identified as *C. soborna* following Whitehead (1972, 1985), Talwar & Jhingran (1991), and Jayaram (1981, 2010).

Corica soborna Hamilton, 1822 (Image 1)

Materials examined: NHM.OU.F-993, 05.i.2019, a stream near Talai Village, Pranahita sub-basin of Godavari, Kumarambheem Asifabad District, Telangana, India, 19.296°N & 79.952°E, 110m, coll. Kante Krishna Prasad & Md. Younus.

Distinguishing characters: *Corica soborna* is a small

fish, attaining about 50mm in standard length. The following characters distinguish the species from its congeners: body moderately elongate; abdomen keeled; 10 pre-pelvic and eight post-pelvic scutes; mouth terminal; second supra-maxilla equal to, or as long as maxilla blade; teeth absent or minute; gill rakers of first gill 19–21 (Whitehead 1972; Talwar & Jhingran 1991); dorsal fin inserted above pelvic origin with two simple (Talwar & Jhingran 1991) and 13–14 branched rays (Hamilton 1822; Talwar & Jhingran 1991); pectoral with 12–13 rays (Hamilton 1822; Talwar & Jhingran 1991); pelvic fin insertion in advance of dorsal fin origin with one simple and seven branched rays (Talwar & Jhingran 1991); anal fin with two simple and 12–13 branched rays and two rays in a distinct finlet (Talwar & Jhingran 1991); caudal fin deeply forked with 18 rays and faint dark edges (Hamilton 1822); scales small; lateral line absent; lateral series scales 40 to 42 (Jayaram 2010).

The present specimen matches the general description provided by Hamilton (1822), Talwar & Jhingran (1991), and Jayaram (2010). Morphometric measurements and meristic counts are presented in Table 1. Minor variations with respect to simple and branched rays could likely be due to the manner in which counts were taken by previous researchers with those made in the present study. In the present specimen, dorsal fin is inserted above pelvic origin, with three simple and 13 branched rays; pectoral fin with one simple and 13 branched rays; pelvic fin with one simple and seven branched rays; anal fin with three simple and 11 branched rays and two branched rays in the finlet;

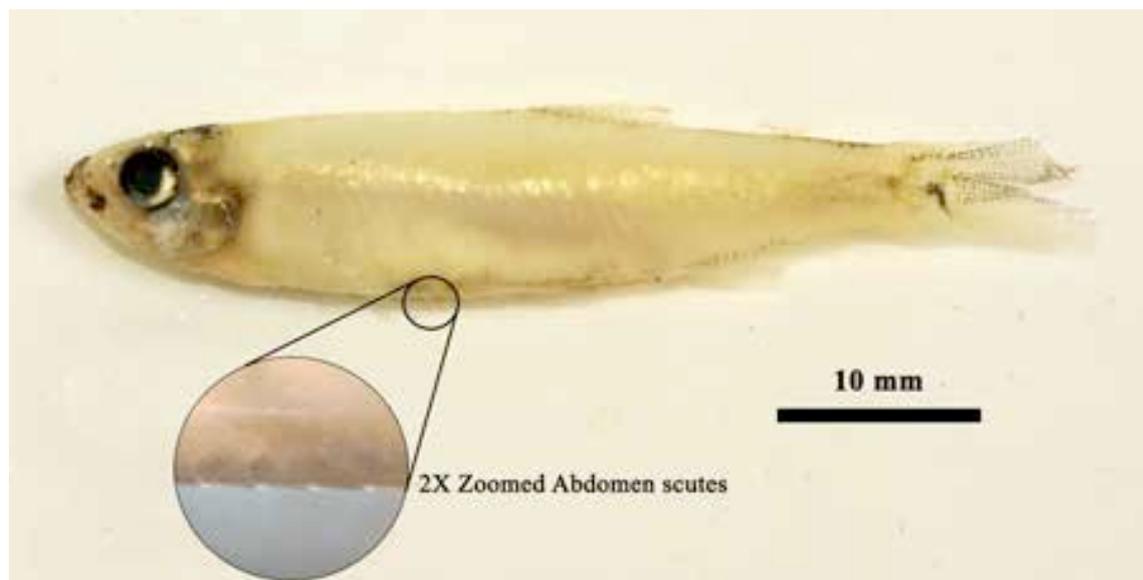


Image 1. Lateral view of the preserved specimen of *Corica soborna* (NHM.OU.F-993). © K. Krishna Prasad

Table 1. Morphometric characters and meristic counts of *Corica soborna* from Godavari River basin, Telangana State, India.

Morphometric Characters	Specimen voucher NHM.OU.F-993
Total length (mm)	50.5
Standard length (mm)	41.6
Head length (mm)	9.0
% of Standard length	
Body depth	20.4
Head length	21.6
Head depth	16.9
Head width	9.2
Eye diameter	7.9
Snout length	6.4
Inter orbital width	4.8
Dorsal fin base length or dorsal fin width	14.6
Pre-dorsal distance	50.8
Dorsal fin length	19.1
Dorsal fin origin to hypural distance	46.2
Pectoral fin length	17.3
Pelvic fin length	13.7
Caudal peduncle length	11.0
Caudal peduncle depth	9.2
Pre-pelvic distance	49.0
Pre-anal distance	70.7
Anal fin base length	17.7
Anal fin length	12.1
% of Head length	
Head depth	78.3
Head width	42.5
Eye diameter	36.6
Snout length	29.8
Inter orbital width	22.2
Meristic counts	
Scales in Lateral Series	41
Transverse scale rows	9
Pre pelvic scutes	10
Post pelvic scutes	8
Pre-dorsal scales	17
Pre-pelvic scales	14
Pre-anal scales	23
Dorsal fin rays	iii+13
Pectoral fin rays	i+13
Pelvic fin rays	i+7
Anal fin rays + (finlet)	iii+11+(2)
Principal caudal fin rays	18
Procurrent caudal fin rays	12

caudal fin deeply forked with 18 principal rays and 12 procurrent rays; lateral series scales 41; and 22 gill rakers on lower arch of first gill.

Habitat: One individual of *Corica soborna* was collected in an eighth-order stream (Strahler 1957) of six-meter depth, with rapid water flow (1.2 m/s) and bedrock, submerged boulders, sand, silt and detritus swamp as substrates. Riparian vegetation on the left bank of the stream was occupied with shrubs and scattered trees of *Tectona grandis*, *Cassia fistula*, *Albizia amara*, and *Acacia leucophloea*, and right bank with cultivated lands. Co-occurring fish fauna included *Amblypharyngodon mola*, *Barilius barila*, *Chanda nama*, *Channa marulius*, *C. striata*, *Glossogobius giuris*, *Labeo calbasu*, *Macrognathus pancalus*, *Mastacembelus armatus*, *Mystus bleekeri*, *M. vittatus*, *Puntius sophore*, *Rasbora daniconius*, *Sperata seenghala*, and *Systemus sarana*.

DISCUSSION

The present record of *Corica soborna* in the Godavari River basin extends its distribution to the middle of peninsular India. This species is currently known from few locations in the Ganga basin from Uttar Pradesh to West Bengal in India, and in Bangladesh (Mahananda, Ken, and Sone river drainages), as well as in the Bharathapuzha River (in Ponnani Estuary) in Kerala. The presence of *C. soborna* in the Godavari River basin suggest that this species may be more widespread in peninsular India than previously thought. In addition to the inland waters of India and Bangladesh (Rahman 1989, 2005; Payne et al. 2004; Wahab 2007; Alam et al. 2013; Arefin et al. 2018), *C. soborna* also occurs in the estuaries of Southeast Asian countries (Fig. 1) including Indonesia (Hardenberg 1936; Whitehead 1967, 1972), Malaysia (Whitehead 1967, 1972), and Singapore (Herre & Myers 1937; Whitehead 1967). This species has also specifically been recorded from Mekong and Bangpakong rivers in Thailand (Taki 1978; Suvatti 1981); however, Chaudhry (2010) doubted the validity of these records. The discontinuous distribution of this species may suggest data-deficiency in surveys and non-availability of verified checklists from other parts of southern and Southeast Asian countries, indicating 'Wallacean shortfall'.

Alteration, fragmentation, and destruction of critical habitats as a result of the construction of irrigation and hydropower projects in the middle and upper reaches of the Godavari, as well as indiscriminate fishing could be detrimental to the survival of the newly detected population of *C. soborna* in Telangana State. We



Figure 1. Global distribution of *Corica soborna*. Verified records in green circles, present record in red circle, and unverified records as red question mark. Numerical numbers in the circles represent ID numbers mentioned in Table 2.

Table 2. Global distribution of *Corica soborna*. Location names in brackets are historic names.

ID	Location	River Basin/ Estuary	Country	Source
1	Aiyargunj, Mahananda River, Bangladesh (Aiyargunj, Mahananda River, India)*	Ganga	Bangladesh	Hamilton 1822; Britz 2019
2	Dehri-on-Sone, Sone River, Bihar	Ganga	India	Motwani & David 1957
3	Dighwara, Ganga River, Bihar	Ganga	India	Karamchandani 1962
4	Goribaba, Ken River, Banda, Uttar Pradesh	Ganga	India	Srivastava et al. 1970
5	Rajmahal, Ganga River, Jharkhand	Ganga	India	Kamal & Ahsan 1978**
6	Ponnani Estuary, Kerala	Bharathapuzha	India	Bijukumar & Sushama 2000
7	Talai, Kumarambheem Asifabad District, Telangana State	Godavari	India	Present Study
8	Halda River, Chittagong, Bangladesh	Karnaphully	Bangladesh	Alam et al. 2013
9	Narsunda River, Kishoreganj District, Bangladesh	Surma-Meghna	Bangladesh	Arefin et al. 2018
10	Padma River, Bangladesh	Ganga	Bangladesh	Payne et al. 2004
11	Kapuas River, West Kalimantan (S.W. Borneo)	Kapuas	Indonesia	Whitehead 1967, 1972
12	Musi River, South Sumatra, Indonesia (S.E. Sumatra)	Musi	Indonesia	Whitehead 1967
13	Padang Tikar Bay, Indonesia	Estuary	Indonesia	Hardenberg 1936; Whitehead 1967
14	Pamangkat, Indonesia (Pemangkat, S.W. Borneo)	Estuary	Indonesia	Whitehead 1967, 1972
15	Perak River, Malaysia (Malaya)	Perak	Malaysia	Whitehead 1967, 1972
16	Singapore	Estuary	Singapore	Herre & Myers 1937; Whitehead 1967

*Type locality; **Kamal & Ahsan (1978) report *C. soborna* being collected from various locations of Ganga River in Patna and Bhagalpur districts in Bihar, and Sahibganj District in Jharkhand, India.

propose further research on understanding the micro-level distribution, population status and dynamics, and threats (both direct and indirect) of this species in Telangana State.

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Strobilanthes affinis (Acanthaceae): a new addition to the flora of Manipur, India

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The genus *Strobilanthes* Blume (1826: 781) from the family Acanthaceae (Wood & Scotland 2009) is a genus of perennial flowering herbs and shrubs with about 453 species worldwide (Biju et al. 2017; Mabberly 2017; Thomas et al. 2019, 2020; Wood et al. 2019). In India, the genus is represented by ca. 150 species (Scotland 1998; Gopalan & Chitra 2008; Karthikeyan et al. 2009; Wood 2014; Biju et al. 2017), and it is found mostly in evergreen forests. The name *Strobilanthes* is derived from the Latin word “*Strobilos*” meaning cone and “*anthos*” meaning flower or shoot. For the first time, the genus was scientifically described by Christian Gottfried Daniel Nees Von Esenbeck in the 19th century for the plants which bloom after long intervals, while some species of *Strobilanthes* bloom annually (Preethi & Suseem 2014; Josekutty et al. 2016, 2017, 2018; Augustine et al. 2017). The genus is distributed in the tropical regions of Asia starting from Afghanistan through India and Southeast Asia to Australia (Josekutty et al. 2018). During 2017–2018, while surveying floral diversity around Imphal, Manipur, when the weather was clear and the vegetation was on the verge of blooming, an exciting view of a blue patch of herbs was recorded. The site was an isolated place near an abandoned old temple-like structure. It

appeared like a small hill at the periphery of a human inhabitation which might have been used in the past as a sacred place where local deities may have been worshipped. It was a moist shady place with huge aged trees sheltering and nurturing vegetation beneath. A significant patch of blue coloured herbs with a height of 45–85 cm were blooming with bright blue coloured flowers in the area of around 4.5–6.5 m² on the slopes of a small hillock. After careful observation in the field, the first author confirmed that the specimen belongs to the genus *Strobilanthes* but confused it with *S. auriculata* or *S. jeyporensis*. After a detailed study of the specimen’s morphology (Image 1) and pollen morphology (Image 2), and comparison with *S. kunthiana*, it was confirmed that the collected species was *S. affinis*. This specimen was again confirmed by professor John Wood, University of Oxford, United Kingdom. After confirmation, literature survey and critical characterization, it was noticed that collected plant had been earlier reported from China (Wood et al. 2003) and the species was also collected by J.D. Hooker and T. Thomson from Meghalaya as *S.acrocephala* more than 10 decades ago. Therefore, it was concluded that the present specimen that was collected is identified as *S. affinis* and reported here for

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the first time from Manipur as an extended distribution.

During our field survey of the flora of Manipur (24.916 N & 93.961 E, 803m), some unidentified species of the genus *Strobilanthes* were collected. The specimens were packed in cloth bags and brought to the laboratory for analysis. The relevant floras were consulted for confirming its identity. A thorough review of the literature, detailed morphological and consultation of herbarium provided by professor Wood revealed that the specimen was *Strobilanthes affinis* (Griff.) Terao ex J.R.I.Wood J.R.Benn. The mounting of herbarium specimens was accomplished using standard taxonomic procedure (Jain & Rao 1977; Bridson & Forman 1998). The voucher specimens (APRF-MAN-BCD/2018-0004) were deposited in the "APRFH" Herbaria, Ambika Prasad Research Foundation, Odisha, India. The photographs of the diagnostic characteristics were taken by Stereo Microscope, (Leica MC 120 HD).

Taxonomic enumeration

Strobilanthes affinis (Griff.) Terao ex J. R. I. Wood J. R. Benn

Description: Herbs, 50–82.5 cm tall with many branches. Stems angular (quadrangular) with white pubescent, hairs. Leaves ovate-elliptic, 7–5 x 4–3 cm to 9–0 x 6–3 cm. Leaves simple, petiole, 3.5–4.3 cm long. Inflorescences axillary or terminal, 3.0–7.5 cm. Each inflorescence has two flowers that have bloomed, after drying another couple of flowers bloom. Flowers size, 1.9–2.3 cm long, 1.2 x 1.2–2.0 x 1.8 cm wide. Floral bracts imbricate with large-celled trichomes. Corolla violet or bluish-violet. Stamens 4 and glabrous, 5.0–2.2 mm. Fruits capsule, pubescent, four seeded, 4.0–7.0 mm. Seeds brown ovate (Image 1).

Flowering & Fruiting: August–March

Distribution: China and India (Manipur, Meghalaya).

Specimen examined: APRF-MAN-BCD-2018-004, India, Manipur, Imphal, 24.916°N & 93.961°E, 803m, Kumar Sanjeet 012 (Holotype APRFH) (Image 4).

Habitat & Ecology: The collected species is *Strobilanthes affinis* found on a small hill top of Manipur State. The collected area lies in the Indo-Burma Biodiversity Hotspot region. It was observed that the habitat is Moist-deciduous vegetation. The most

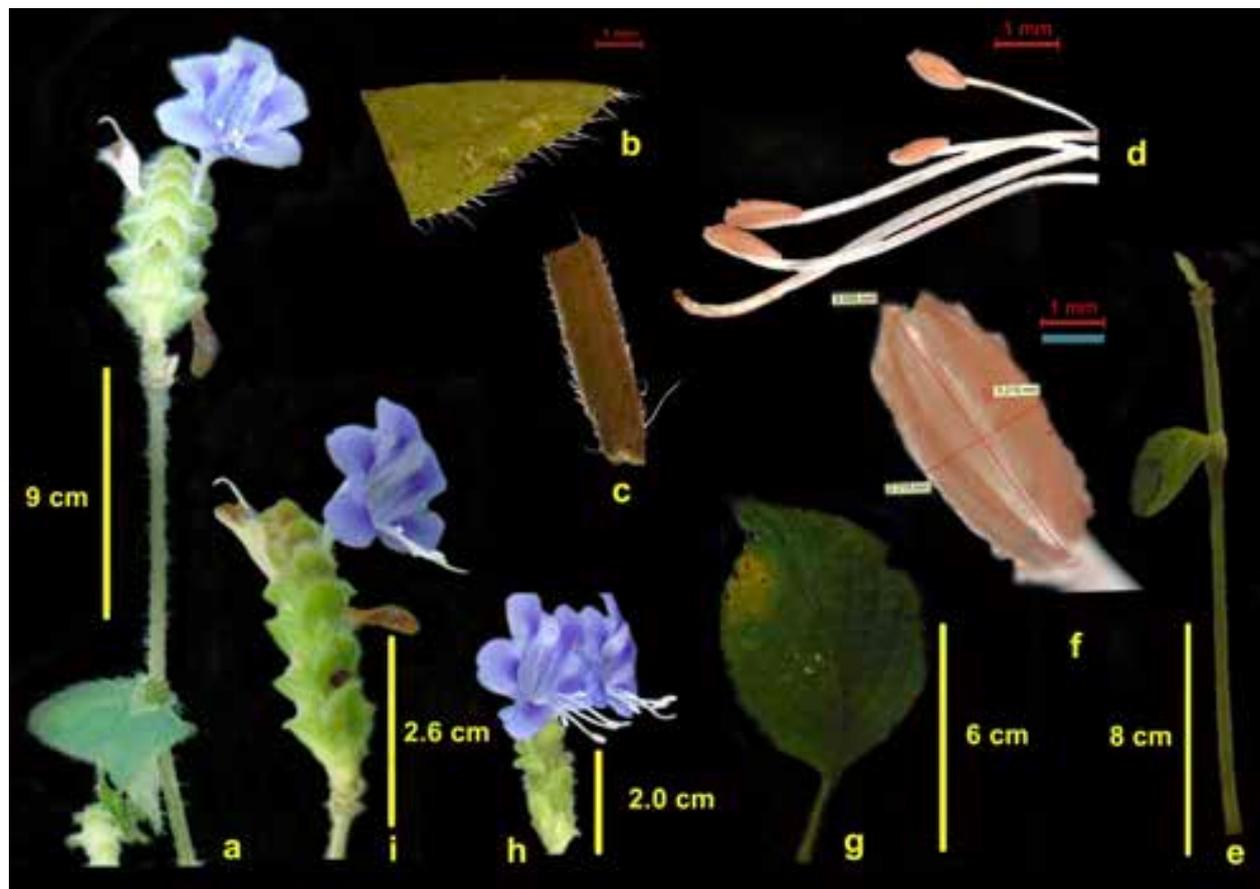


Image 1. Floral parts and morphology of collected *Strobilanthes affinis*. © Sanjeet Kumar



Image 2. Pollen of collected *Strobilanthes affinis*

common associated herbs and climbers are *Passiflora foetida*, *Solanum nigrum*, *Sida cordifolia*, *Cissampelos pareira*, and *Phyllanthus reticulatus* with some tree species like *Bombax ceiba* and *Pongamia pinnata*. It was found in 6.09 x 9.14 – 4.57 x 7.01 m area of a hill top of said region.

Remarks: The most unique characters of collected specimen are four exerted stamens (Image 3). Wood et al. (2003) described it from China and recorded that the flowering period is only from September to November.

In the present study, however, it was observed that it flowers from August to March. It was also observed that *Sida acuta* is the main associate flora of the collected species. As it is found only one region of Manipur, it should be assessed for threat status using categories of IUCN, and take necessary conservation action.

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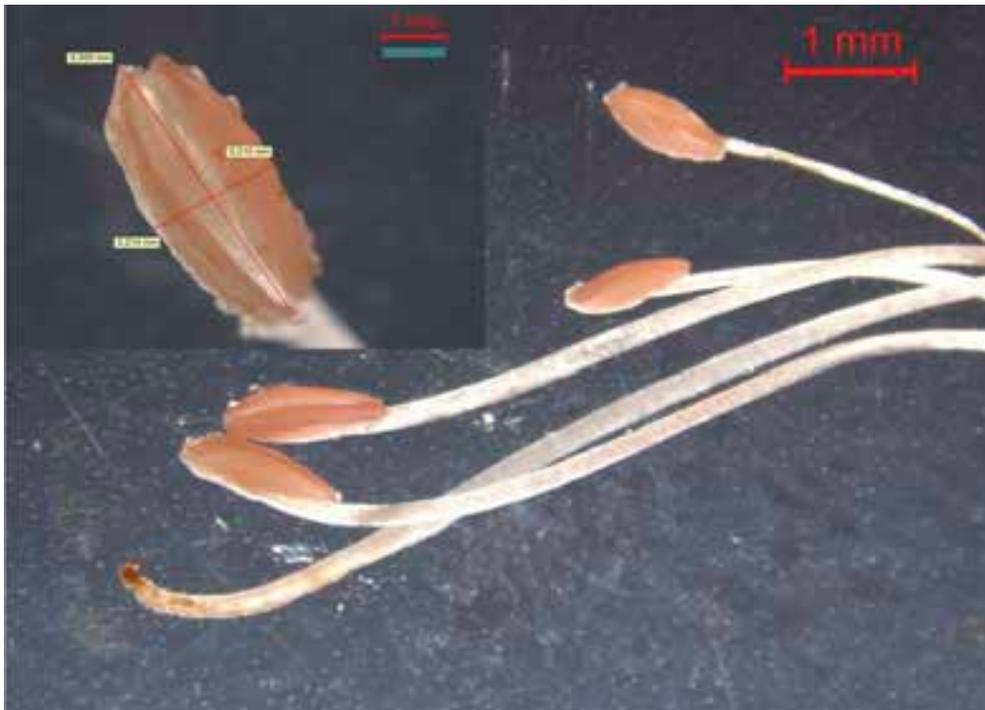


Image 3. Four exerted stamens *Strobilanthes affinis* (Griff.) Terao ex J.R.I. Wood J.R. Benn. © Sanjeet Kumar

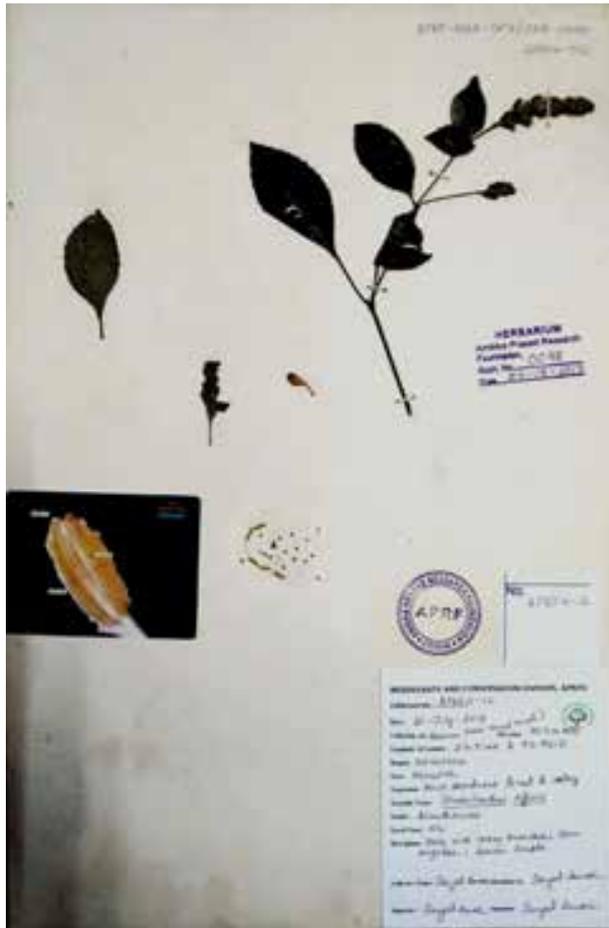


Image 4. Herbarium image of *Strobilanthes affinis*.

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A new species of the genus *Opius* Wesmael, 1835 (Hymenoptera: Braconidae: Opiinae) from Kashmir Himalaya, India

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Abbreviations: AOL—Anterior ocellar line | POL—Posterior ocellar line | OOL—Ocello ocular line | OOD—Ocellus diameter | F—Flagellomere | T₁—First metasomal tergite | MDZUK—Museum Department of Zoology, University of Kashmir.

Opiinae is one of the largest subfamily of Braconidae (Hymenoptera: Ichneumonoidea) with 39 genera and 2,061 described species worldwide. Species of the subfamily Opiinae are solitary koinobiont endoparasitoids of larvae of cyclorhaphous Diptera, but oviposition may take place in the egg of the hosts (egg-larval parasitoids). They play significant role in the control of dipterous pests such as fruit flies (Tephritidae) and leaf-miner flies (Agromyzidae) (Wharton 1984, 1997; Schuster & Wharton 1993). The parasitoid larvae complete their development within the host larvae, pupate and emerge as adults (Li et al. 2013). *Opius* Wesmael is the largest genus of subfamily Opiinae and also one of the largest in the family Braconidae, with 33 subgenera and 1,202 described species (Yu et al. 2019). Some species of *Opius* have great potential in biological control of Agromyzid leaf miners. *O. tirolensis* is recorded as a biological control agent against *Phytomyza*

flavicornis, *O. (G) caucasi* against *Chromatomyia horticola*, and *O. (A) nowakowskii* against *Phytomyza thysseini* (Yu et al. 2012). The genus is represented by 51 species spread over 11 subgenera (including four species of subgenera *Utetes* from India, i.e., *Opius (Utetes) mudigerensis*; *Opius (Utetes) poonchicola*; *Opius (Utetes) parempiformis*; *Opius (Utetes) minicorpus* (Fischer, 1966, 1980, 1987, 1996, 2005, 2012; Samiuddin et al. 2009). In the present work a new species *Opius (Utetes) hazratbalensis* sp. nov. from Kashmir, Himalaya has been described and illustrated.

The specimens were collected by using sweeping net. Slides and card mount specimens were examined under the binocular microscope. The ocular micrometer was used to measure (linear side of 100 divisions) fitted in one of the two eye pieces of the binocular. The ocular micrometer was calibrated with the help of stage micrometer under 8x,10x magnifications. Photographs were taken at different magnifications (4–16x) by the camera attached to Stereozoom binocular (SMZ1500) and Nikon SMZ 25. The terminology and the venation were followed by Achterberg (1993) and for surface sculpture Eady (1968) has been followed.

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Opilus (Utetes) hazratbalensis* sp. nov.*(Image 1 A–G)**

urn:lsid:zoobank.org:act:DBDA5D24-0C95-47DD-B8DB-DAF7DAAD65AE

Material examined: Holotype: MDZUK ZoKU Art/06180, 15.v.2019, female, University campus, Hazratbal, Jammu & Kashmir, India; coll. Zaheer Ahmed. In near future it will be deposited in the Department of Zoology, Aligarh Muslim University, Aligarh.

Paratype: 1 female, same data as holotype.

Female: Body length 5.95mm; forewing 6.07mm.

Head: The width of head in dorsal view 1.7x its length and 1.3x its height; occipital carina complete; length of eye in lateral view 1.3x its width and 2.7x its temple; temple smooth, OOL:POL:AOL:OOD=3:3:14:3; vertex antero-medially smooth, remaining rugose and sparsely setose, width of vertex 2.7x its length; frons smooth, shiny and densely setose near antennal sockets, width of frons 4.3x its length; face punctate reticulate with median longitudinal groove and densely setose, width of face 1.6x its length; tentorial pits deep and wide; intertentorial line 2.6x tentorio-ocular line; malar suture present, length of malar space 1.7x basal width of

mandible; clypeus concave, rugose and densely setose, length of clypeus 2x its width; antennae 47-segmented, length of scape 1.3x its width, length of pedicel 1.16x its width, basal flagellar segments relatively long, first segment (F_1) 3.2x as long as wide, length of F_2 5.3x: F_3 - F_4 5x: F_5 - F_6 4.3x: F_7 - F_{12} 4x: F_{13} - F_{22} 3.3x: F_{23} - F_{26} 3x: F_{27} - F_{34} 4.5x: F_{35} - F_{44} 4x: and F_{45} 4.5x their widths respectively.

Mesosoma: Length of mesosoma 1.8x its width and 2.2x, 2.7x as high as head (95:30) and 1.15x wide as head (52:60), mesonotum 2.3x as long medially, as wide between tegulae; dorsal surface of scutellum punctate, laterally striate; median lobe of mesoscutum reticulate, lateral lobe smooth, polished; notauli deep, crenulate antero-medially, remaining rugose; mesopleuron antero-medially granulate remaining smooth, metapleuron reticulate rugose; mid pit present; precoxal sulcus absent; scutellar sulcus deep and broad with two median carinae; metanotum concave, smooth; propodeum coriaceous entirely without five sided areola.

Wings: Length of forewing 2.7x its width; pterostigma long cylindrical, length of pterostigma 10.2x its width (72:7), length of vein 1-R1, 1.1x length of pterostigma; vein r arising 1/3rd of pterostigma; m-cu and cu-a



Image 1. Female *Opilus (Utetes) hazratbalensis* sp. nov. (magnifications 4–16x): A—adult female | B—dorsal head | C—frontal head | D—mesosoma dorsal | E—metasoma lateral | F—hindwing | G—forewing. © Zaheer Ahmed

<i>Opius (Utetes) hazratbalensis</i> sp. nov.	<i>Opius (Utetes) mudigerensis</i> , Fischer
1. Body length 5.95mm	1. Body length 4mm
2. Basal flagellar segments relatively long, first segment (F ₁) 3.2x as long as wide	2. Basal flagellar segments relatively short, first segment (F ₁) 2x as long as wide
3. Mesonotum 2.3x as long medially as wide between tegulae	3. Mesonotum 1.2x wider between tegulae than longitudinally
4. First tergite 1.75x longer than wide at hind (posteriorly)	4. First tergite 1.25x longer than wide at hind (posteriorly)
5. Propodeum without five-sided areola	5. Propodeum with five-sided areola
6. Hind tibia 3.4x as long as ovipositor sheath	6. Hind tibia as long as ovipositor sheath

Key to the Indian species of the subgenus *Utetes*

- 1 Antennae 24–28 segmented; T₁ 1.2x–1.5x as long as wide; ovipositor sheaths as long as T1 in lateral view (12:12); body length 1.9–4 mm 2
- 1 Antennae 47 segmented; T₁ 1.7x as long as wide; ovipositor sheaths short as compared to T1 in lateral view (8:12); body length 5.95mm *Opius (Utetes) hazratbalensis* sp. nov. female
- 2 F₁ 2x as long as wide *Opius (Utetes) mudigerensis* Fischer, 1980, female
- 2 F₁ 4–6x as long as wide 3
- 3 F₁ 6x as long as wide; clypeus 3x as wide as high; sternaulus with some notches.....*Opius (Utetes) parempiformis* Fischer, 2012, female
- 3 F₁ 4–5x as long as wide; clypeus 1.5x–2x as wide as high; sternaulus densely crenulated 4
- 4 Face as wide as high; propodeum with a broad pentagonal area *Opius (Utetes) minicorpus* Fischer, 2012, female
- 4 Face 1.25x as wide as high; propodeum with basal carina, pentagonal area and costulae, irregularly sculptured *Opius (Utetes) poonchicola* Fischer, 2012, female

postfurcal; SR1+3-SR slightly curved; r: 2-SR:S-R1+3-SR = 11:20:115; vein 3-SR 1.7x as long as 2-SR; 1-CU1: 2-CU1: 3-CU1 = 5:27:3; length of hind wing 4.2x its width; 1M: 1-r-m: 2-SC+R= 26:4:5

Legs: Hind coxa rugose, 1.5x as long as wide; length of hind femur, tibia and basitarsus 5.5x, 9.3x and 7.5x their width respectively; length of hind tibial spurs 0.06x hind basitarsus.

Metasoma: Length of metasoma 2.4x its width and 4.4x its height; first metasomal tergite 1.75x longer than wide at hind (posteriorly); length of first metasomal tergite 2.6x its apical width, apical width 1.5x its basal width, its surface longitudinally striate throughout the length; dorsope present; spiracles present medially; ovipositor short and pointed, ovipositor sheaths short as compared to T1 in lateral view (8:12); hind tibia 3.4x as long as ovipositor sheaths.

Colour: Head, face, mesosoma brownish to blackish except tegula brown, wings membranous with brown venation except pterostigma dark brown, antennae uniformly dark brown, legs brownish-yellow, mandibles dark brown with black teeth, eyes blackish, ocelli transparent, ocellar area black, maxillary and labial palpi yellowish, metasoma blackish, ovipositor reddish, ovipositor sheaths dark brown.

Male: Unknown

Host: Unknown

Etymology: The new species name refers to its type locality.

Discussion: *Opius (Utetes) hazratbalensis* sp. nov. closely resembles with Indian species *Opius (Utetes) mudigerensis* Fischer, 1980. However, it differs from *mudigerensis* in having body length 5.95mm; basal flagellar segments relatively long, first segment (F₁) 3.2x as long as wide; mesonotum 2.3x as long medially as wide between tegulae; first tergite 1.75x longer than wide at hind (posteriorly); propodeum without five sided areola and ovipositor sheath short in lateral view, hind tibia 3.4x as long as ovipositor sheath.

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Larvae of the blow fly *Caiusa testacea* (Diptera: Calliphoridae) as egg predators of *Polypedates cruciger* Blyth, 1852 (Amphibia: Anura: Rhacophoridae)

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Habitat destruction and alteration have been identified as the most detrimental causes of amphibian decline (Kiesecker 2003). The effects of climate change and amphibian diseases, however, are emerging topics, and have taken increased attention in conservation approaches regarding the amphibian fauna (Hayes et al. 2010; Li et al. 2013). Predatory pressure during different life stages of amphibians is another factor that significantly affects populations (Chivers et al. 2001; Blaustein et al. 2012). Diverse invertebrate and vertebrate fauna prey on eggs and tadpoles of aquatic and terrestrial nesting anurans (De Silva 2001a,b; Lingnau & Di-Bernardo 2006). According to Downie (1990), terrestrial foam nests of Rhacophoridae have evolved to protect eggs and embryos from aquatic predators. Some vertebrates (e.g., monkeys and snakes) and invertebrates (e.g., beetles, ephydrid flies, phorid flies, spiders, ants, and blow flies), however, have been identified as egg predators of anuran foam nests (Vonesh 2000; Rödel et al. 2002; Menin & Giaretta 2003; Lingnau & Di-Bernardo 2006; Banerjee et al. 2018). Blow flies of the genus *Caiusa* (Diptera: Calliphoridae) are one of the major predators

of terrestrial Rhacophoridae eggs (Rognes 2015). These flies are one of the major reasons for embryo mortality of some rhacophorid genera, including *Chiromantis*, *Feihyla*, *Polypedates*, and *Rhacophorus* (Lin & Lue 2000). So far, seven known species of *Caiusa* (*C. borneoensis* Rognes, 2015, *C. coomani* Séguy, 1948, *C. indica* Surcouf, 1920, *C. karrakerae* Rognes, 2015, *C. kurahashii* Rognes, 2015, *C. violacea* Séguy, 1925, and *C. pooae* Rognes, 2015) have been identified as foam nest predators and predators of jelly-like egg masses of anurans (Lin & Lue 2000; Rognes 2015; Banerjee et al. 2018). The emerging larvae of these fly species consume eggs and developing embryos in egg masses. There are knowledge gaps in our understanding of the fly-anuran interactions and the wider impact of these flies on anuran population dynamics.

Sri Lanka is a tropical country with more than 120 species of anurans, nearly 104 of which are endemic to the country (De Silva & Wijayathilaka 2019). Approximately 83 (69%) of the reported species belong to the family Rhacophoridae, including arboreal foam nesting *Polypedates* and *Taruga* species (Meegaskumbura et al.

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2010). The majority (more than 75%) of anuran species in the country are categorized as threatened, mainly due to anthropogenic activities (Manamendra-Arachchi & Meegaskumbura 2012; De Silva & Wijayathilaka 2019). Current conservation approaches are mainly aimed at minimizing habitat destruction and other adverse human activities. Only a few studies, however, have reported the effect of amphibian diseases on the population structure of anurans in Sri Lanka (De Silva 1999; Rajakaruna et al. 2007; Jayawardena et al. 2010; De Silva 2011), and hardly any studies have investigated predatory pressure on different life stages of amphibians in the country. Morgan-Davies (1958) reported *Caiusa indica* as predatory in foam nests of *Polypedates cruciger* Blyth, 1852 (Anura: Rhacophoridae) in Sri Lanka. According to De Silva & De Silva (2000), a species of Calliphoridae fly acts as an egg predator of *P. cruciger* frogs, however, these authors did not provide a species-level identification for the flies. Therefore, there are some literature gaps in information about predatory flies and their pressure on the developmental stages of anurans in Sri Lanka. Thus, detailed investigations including systematic and quantified studies to assess the damage caused by the egg predators to anuran eggs are important in relation to conservation actions. In this study, we identified natural dipteran predators of foam nests of *P. cruciger*, an endemic Rhacophoridae species in Sri Lanka. Further, we quantified the egg mortality of *P. cruciger* due to the infestation of the predatory dipteran fly.

The study was conducted from May 2019 to August 2019, at two localities [Gampola (7.150°N, 80.555°E) and Peradeniya (7.259°N, 80.597°E)] in the Kandy District of Sri Lanka. Spawns were searched for in microhabitats with *P. cruciger* (i.e., man-made ponds, cement water tanks, domestic wells, tree-holes, and organically managed agricultural lands). When a fresh spawn was located, it was observed and video recorded for about 10–15 minutes to report spawn visitors. The location of the foam nest and the height from the ground level to the nest were recorded. The spawns were examined daily at both selected localities until the embryos developed into tadpoles. A plastic container filled with 1,000ml of dechlorinated water was kept below each egg mass to collect emerging tadpoles. Observations were made at 24-hour intervals and the developed tadpoles were released to the respective water sources after recording the number. A similar procedure was followed for both infected and non-infected spawns. The presence of maggots, color changes, and the shape of the foam nests were used to distinguish infected nests

from uninfected ones. Three severely infected spawns were carefully removed from the attached substrates and brought to the Insectary, Department of Zoology, University of Peradeniya for further investigations. At the laboratory, the foam nests were placed in dechlorinated water in a tray and transferred to fine-mesh mosquito rearing cages (50 × 50 × 50 cm) for maintenance of the fly colonies (at 25°C temperature, 75% relative humidity, and 12 D: 12 L photoperiodicity). Emerged flies were euthanized at -20°C in a freezer and pinned for identification. Morphological identification was done using the standard taxonomic key in Rognes (2015).

To confirm the identity of the dipteran species, DNA barcoding was also performed. DNA was extracted from some of the collected flies following Livak (1984). The mitochondrial Cytochrome Oxidase I (COI) gene was amplified using the previously described primers C1–J–1718F (5'–GGA GGA TTT GGA AAT TGA TTA GTT CC–3') and C1–N–2191R (5'–CCC GGT AAA ATT AAA ATA TAA ACT TC–3') (Simon et al. 1994). PCR amplification was done in a thermal cycler (Techne–Flexigene, England) following Nolan et al. (2007). Positive PCR products were sequenced using an automatic DNA sequencer (Applied Biosystems Series 3500, U.S.A.) in the Department of Molecular Biology and Biotechnology, University of Peradeniya. The sequence trace files were manually inspected using MEGA V7 (Kumar et al. 2016) and low-quality sequences were excluded from the analysis. The DNA sequences were annotated using the GenBank database (<https://www.ncbi.nlm.nih.gov/>) and BLASTn tool. The newly generated sequences were deposited in GenBank under the accession numbers MN786865 and MN786866.

The dissection and examination of male genitalia were done following Rognes (2009). The tip of the abdomen (from tergite 4) was removed and transferred to a 10% potassium hydroxide solution, then heated in a water bath for about 20 minutes. The abdomen was then transferred to distilled water and rinsed with 95% ethanol for 10 minutes to fix the integument. The male genital organs were separated using fine forceps, for preparation of microscopic slides. The separated male genitalia were mounted using Canada Balsam, and photographs of the prepared slides were taken using an Olympus BX53 Digital Upright Microscope (Olympus Corporation, Florida, USA).

Morphological identification confirmed that the emerged flies belonged to *Caiusa testacea* Senior-White, 1923 of the family Calliphoridae. According to Rognes (2015), the following morphological features were



Image 1. Infected and non-infected foam nests of *Polypedates cruciger* and different life stages of *Calusa testacea* flies: a—adult *Polypedates cruciger* | b—uninfected fresh foam nest attached to a *Polyscias scutellaria* leaf | c—tadpoles from an uninfected foam nest | d—putrefying foam nest due to *C. testacea* infection | e—*C. testacea* fly on a fresh foam nest | f—*C. testacea* 3rd instar larva | g—*C. testacea* pupae | h—lateral aspect of adult *C. testacea* fly. © a,b—Anslem de Silva; c-h— W.G.D. Chathuranga.

identified for them. Cerci short, backwardly bent, and with a pronounced distal separation between the apices in dorsal view. Base of cerci wide proximal to separation (Image 2A). In lateral view, surstylus rather broad and short, very gently curved below. Thoracic dorsum yellow and tergites 4 and 5 of abdomen with slight darkening and lack of metallic bluish sheen (Image 2D). A BLAST search of the GenBank database showed a 96.92% identity to available *Calusa testacea* sequences together with a 100% query cover.

A total of 24 spawns of *P. cruciger* were studied (Image 1a–1d). Observations were carried out on 10 spawns in Gampola (including the three collected spawns) and 14 spawns from the Peradeniya study site. These spawns were located at a height of 0.1–3.0 m above the ground. Plant species such as *Polyscias scutellaria* (Araliaceae), *Nelumbo nucifera* (Nelumbonaceae), *Gliricidia sepium* (Fabaceae), *Echinodorus palifolius* (Alismataceae), *Persea americana* (Lauraceae), and artificial substrates including cement walls, metal wire mesh, ceiling sheets, metal or plastic pipes just above a water source, were the most common spawning sites of *P. cruciger*. Of the examined spawns, 16 (66.7%) were not infected while eight (33.3%) were infected with fly larvae (Image 1f). All the infected spawns were reported from the Gampola study location, representing 80% of the total.

During this study, we observed oviposition of *C. testacea* flies only three times (Image 1e) on fresh foam nests of *P. cruciger*, and the larvae of *C. testacea* emerged from two-day-old infected spawns. An average of 354 ± 67 embryos developed into tadpoles (Image 1c)

from healthy spawns ($n=15$), except one that produced an exceptionally high number of tadpoles (approximately 800). When compared with the healthy spawns, none of the embryos of the infected spawns ($n=8$) developed into tadpoles (Image 1d). According to our observation of eight infected spawns, approximately 400 embryos were destroyed with a single nest infestation. An average of 52 ± 9 *C. testacea* larvae pupariated (Image 1g) and 17 ± 8 emerged as adults from the three collected spawns (Image 1h). Accordingly, an average of 33% (17/52) of the larvae were able to complete their life cycle from a single spawn. The 1st to 3rd instar larval stages of the fly lasted around 6–7 days, while the puparial period lasted 8–11 days. The life cycle of *C. testacea* was completed within 18 to 20 days. Emerged adult flies were freeze-killed and pinned for identification. Larval instars, puparia, and a few adults of *C. testacea* were also preserved in 70% ethanol as voucher specimens and deposited in the Zoonotic and Disease Ecology Laboratory of the Department of Zoology, University of Peradeniya, Sri Lanka. Different morphological body aspects of *C. testacea*, including taxonomic features, are shown in Images 1h, 2A–2D.

Our study highlights the threat caused by *C. testacea* flies to the foam nests of *Polypedates cruciger* frogs and provides an indication of the major impact of these flies on the population dynamics of *P. cruciger*. Even though studies have reported the impact of predatory pressure causing the population decline of amphibians (Lin & Lue 2000; Kiesecker 2003), it has not been listed as a priority factor in conservation approaches in Sri

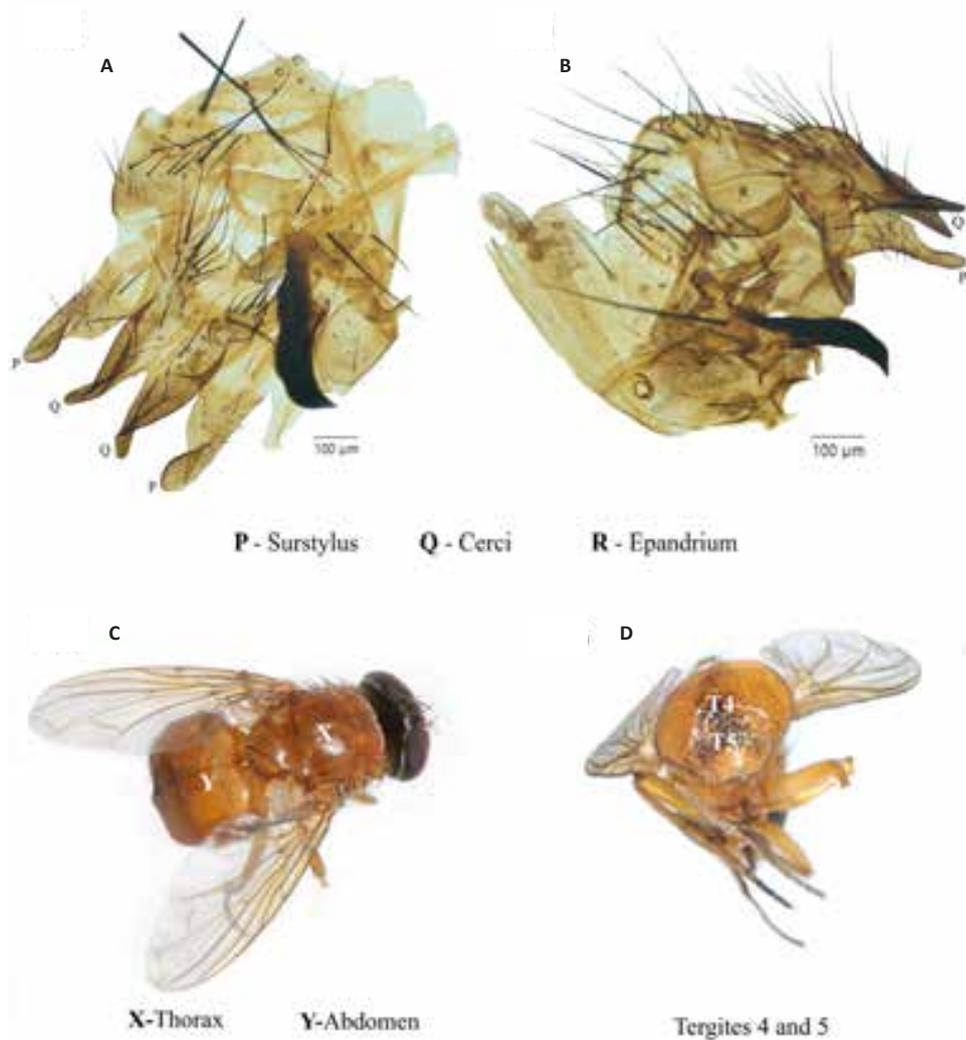


Image 2. Morphological features of *Caiusa testacea* flies: A—posterior aspect of the genitalia | B—lateral aspect of the genitalia | C—yellowish thorax and abdomen | D—yellowish T4 and T5 with slightly darkened patch. © a,b—W.G.D. Chathuranga; c-d—Kumudu Wijesooriya.

Lanka. In this study, we provide data on the natural predatory pressure of the calliphorid fly *Caiusa testacea* on the population structure of the rhacophorid tree frog *Polypedates cruciger*. Further, our results provide evidence of natural threats of Rhacophoridae anurans in Sri Lanka.

The presence of these flies had been reported from Sri Lanka, India, and Nepal by Rognes (2015), however, these flies had not been identified as egg predators of Sri Lankan Rhacophoridae species by any of the earlier studies. Our results reveal that larvae of *C. testacea* flies destructively consume eggs and embryos of *P. cruciger*. In an earlier study, *Caiusa indica* was identified as an egg predator of *P. cruciger* in Sri Lanka (Morgan-Davies 1958); however, previous studies had not identified *C. testacea* as a predator of foam nests of Rhacophoridae, and this is the first study that reports on the feeding

behavior and the life history of *C. testacea*.

Rognes (2015) estimated that the time from the infestation of spawns by *Caiusa* flies to the completion of metamorphosis is nearly a week. In contrast, we observed a relatively longer developmental period, where *C. testacea* flies complete metamorphosis within three weeks. Lin et al. (2000) and Lin & Lue (2000) described the oviposition behavior of *Caiusa violacea* (as *C. coomani*). According to those authors, the flies lay their eggs when the outer surface of the foam nest is soft, within a few hours after the foam nest is formed. Similarly, Banerjee et al. (2018) reported that *Caiusa* flies lay their eggs on foam nests seven hours after the construction of the nest. Our study confirmed the oviposition of *C. testacea* flies on fresh foam nests of *P. cruciger* (Image 1e), however, we were not able to provide more specific information about the timeframe

during which the flies are attracted to the nests. Our observations showed that larvae appeared within 2 to 3 days after oviposition and that the life cycle was completed (to metamorphosis) within 18 days.

Rognes (2015) reported that most of the dipteran predators of foam nests are able to respond to chemical cues released from the fresh foam nests built by the frogs. Thus, the gravid females of *C. testacea* flies may respond to chemical cues of freshly formed foam nests or chemical signals produced by *P. cruciger* frogs during spawning. Our data could not, however, confirm this hypothesis. There are interesting hypotheses explaining the selection of foam nests by dipteran flies as oviposition sites. For example, Banerjee et al. (2018) hypothesized that the frog eggs represent easier prey for *Caiusa* larvae compared to mobile tadpoles, which may allow these flies to overcome environmental constraints and resource limitations.

The distribution of *P. cruciger* extends 1,500m in the wet zone of central and southwestern parts of Sri Lanka (De Silva & De Silva 2000). *Caiusa testacea* has also been reported from similar locations in the central part of Sri Lanka, including Maskeliya, Suduganga, Kandy, and Niroddumunai (Rognes 2015), where *P. cruciger* is also reported. This habitat overlap of the predatory flies and *P. cruciger* may have driven the evolution of the predatory behavior of this fly species on the foam nests of *P. cruciger*. At the same time, this habitat overlap may negatively affect *P. cruciger* as it gives more opportunities for *C. testacea* flies to attack their nests. According to IUCN Red list 2012 categories, *P. cruciger* is listed as a Least Concern (LC) anuran species (Manamendra-Arachchi & Meegaskumbura 2012); however, the continual increase of anthropogenic impacts and changing climatic factors, together with infestations of *C. testacea*, may negatively affect *P. cruciger* populations, causing it to become a 'threatened species'. Furthermore, Sri Lanka harbors four more foam nesting anuran species in the family Rhacophoridae [(*Polypedates maculatus* Gray 1830, *Taruga eques* Günther, 1858, *Taruga fastigo* (Manamendra-Arachchi & Pethiyagoda, 2001), and *Taruga longinasus* (Ahl, 1927)] (Meegaskumbura et al. 2010). As a result, there are possibilities for all other foam nesting Rhacophoridae anurans to be endangered by nest predation by *Caiusa testacea* flies. As we have seen the habitat overlap of Rhacophoridae species and these flies, there is a high chance of egg predation by *Caiusa* on these tree frogs in Sri Lanka. A proper understanding of the biology, distribution, and population assessments of both *C. testacea* and *P. cruciger*, however, will be

vital in assessing the threats of *C. testacea* flies on the population dynamics of *P. cruciger* in the country.

In summary, we report *C. testacea* as a predator of foam nests of *P. cruciger* frogs of the family Rhacophoridae in Sri Lanka for the first time. More importantly, we recognize the predatory pressure of these flies on spawns of *P. cruciger*, highlighting their needful consideration in conservation approaches concerning these frogs.

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Blank Swift *Caltoris kumara moorei* (Evans, 1926) (Lepidoptera: Hesperiiidae) in Dehradun Valley, Uttarakhand, India: a new record for the western Himalaya

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Caltoris kumara (Moore, 1878) (Hesperiiidae: Hesperinae: Baorini) is known to occur as three races. The nominate race *C. k. kumara* (Moore, 1878) occurs from northern Gujarat to southern India (Western Ghats –(Maharashtra southwards to Kerala, Bangalore, Tamil Nadu), Madhya Pradesh, and Chattisgarh). The second race *C. k. moorei* (Evans, 1926) is found from western Nepal to Myanmar (Dawnas), through Sikkim, Arunachal Pradesh, northeastern India besides West Bengal, Bangladesh, and southern Yunnan (China). While the third race *C. k. lanka* is restricted to Sri Lanka (Wynter-Blyth 1957; Smith 1989, 2006; Zhang et al. 2010; Kehimkar 2016; Gasse 2017). The species is known to prefer forested areas up to 1,950m and is in flight from January–December (Kehimkar 2016), however, the species has never been reported from Uttarakhand or even the western Himalaya (Singh & Sondhi 2016).

On 21 September 2018 the author recorded *C. k. moorei* (Evans, 1926) in the New Forest Campus of Forest Research Institute, Dehradun (30.341°N & 79.997°E; 670m), India. The butterfly was trapped in the window of the Forest Research Institute building. The specimen was collected, pinned, and preserved as an unidentified

specimen, however, it was recently identified by the author based on wing morphology and its genitalia. Distinctive morphological features of this individual (female; forewing length: 21mm; Image 1) being: no spot in upper forewing cell area, prominent large round spots in 1b in the upper forewing inner edge which is visible as a diffuse spot on the underside (Wynter-Blyth 1957). The female genitalia was dissected (Image 2, Figure 1) and the species was identified and compared with that of three other congeners [*C. confusa* (Evans, 1932); *C. philippina* (Herrich-Schäffer, 1869) & *C. tulsii* (de Nicéville, [1884]) in the *Caltoris* group (Devyatkin 2010). In the female genitalia of *C. k. moorei* the postvaginal plate (PPL) is distally concave, with its outer angles acute and lateral lobes of the antevaginal plate projections are only slightly narrower than PPL which is distinct from others (Devyatkin 2010).

The larval food plants of this species are known to be Poaceae (Robinson et al. 2010), *Bambusa* (Swinhoe 1913; Wynter-Blyth 1957), *Bambusa vulgaris* (Kalesh & Prakash 2015), *Bambusa tuldoidea* (Nitin et al. 2018), *Imperata cylindrica* (Wynter-Blyth 1957; Robinson et al. 2010), *Ochlandra scriptoria* (Kalesh

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Image 1. Blank Swift *Caltoris kumara moorei* (Evans, 1926) female collected from New Forest, Dehradun, Uttarakhand (lower western Himalaya) on 21.ix.2018: a—dorsal view | b—ventral view. © Arun Pratap Singh



Image 2. Female genitalia organ of Blank Swift *Caltoris kumara moorei* (Evans, 1926) collected from the New Forest, Dehradun, Uttarakhand (lower western Himalaya) on 21.ix.2018.

& Prakash 2015), *Ochlandra travancorica* (Kalesh & Prakash 2015), and *Oryza sativa* (Davidson & Aitken 1890). Amongst these the genus *Bambusa* is present at a number of places in the New Forest Campus in plantations, botanical garden, bambusetum and also as clumps in hedge-groves, thus it also likely that the species is breeding here but needs further investigation.

This is the first record of this species from the state of Uttarakhand, India and the western Himalaya.

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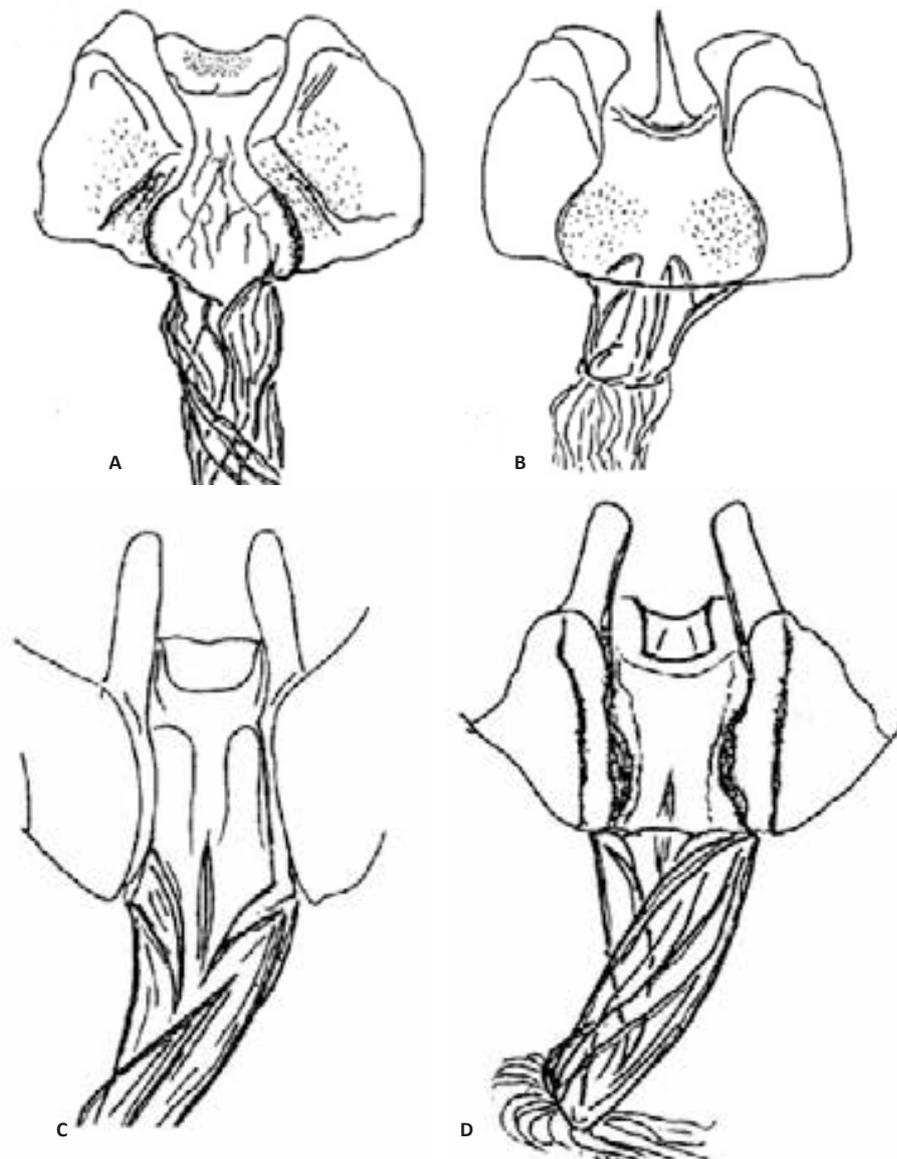


Figure 1. The female genitalia of four species of the genus *Caltoris* Swinhoe, 1893 (adapted from Devyatkin 2010): A—*Caltoris philippina* (Herrich-Schäffer, 1869) | B—*Caltoris tulsii* (de Nicéville, [1884]) | C—*Caltoris confusa* (Evans, 1932) | D—*Caltoris kumara moorei* (Evans, 1926).



First photographic record of the Asiatic Brush-tailed Porcupine *Atherurus macrourus* (Linnaeus, 1758) (Mammalia: Rodentia: Hystricidae) from the Barak Valley region of Assam, India

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An Asiatic Brush-tailed Porcupine *Atherurus macrourus* (Linnaeus, 1758) was photographed on the 25 February 2019 at 23.49h in semi-evergreen forest, close to Dosedewa Village in Karimganj District of Assam (92.405°N & 24.358°E) using a camera trap during biodiversity documentation activities. This is the first photographic record of this species from the Barak Valley region of Assam (Image 1–3). The Asiatic Brush-tailed Porcupine is a species of rodent in the Hystricidae family and is the smallest and rarest of the three species of porcupines found in southern Asia (Molur 2020). In India, this species occurs only in northeastern India and has been reported from Namdapha National Park of Arunachal Pradesh, Garampani Wildlife Sanctuary of Assam, Intanki National Park in Nagaland, Ngengpui Wildlife Sanctuary of Mizoram, and Khasi Hills of Meghalaya (Talukdar et al. 2019); there are also records from the other states (Choudhury 2013). In 2014, the species was observed from the Gedu Territorial Forest Division of western Bhutan (Dhendup & Dorji 2017). Worldwide, it has also been reported from southern China, northern Myanmar, northern Thailand, Laos, Vietnam, Malay peninsula, and adjacent islands (Molur 2020).

The species is distinguished by absence of crest, back covered with rigid spines and tail bearing scales with short spiny bristles in-between and ending in a cluster of alternately expanded and contracted papery hairs 8–10cm long giving a characteristic brush-like appearance which was clearly captured in the photographs (Agarwal 2000). Being a shy, nocturnal, and fossorial animal, it is very difficult to get diagnostic



Image 1. Asiatic Brush-tailed Porcupine *Atherurus macrourus*

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Image 2 & 3. Asiatic Brush-tailed Porcupine *Atherurus macrourus*

images of this species, except through camera trapping. Porcupines are seen as a pest by farmers and are also trapped and consumed as meat. Although this species is classified as Least Concern (LC) in the IUCN Red List it is protected in India under Schedule II – Part I of the Indian Wildlife Protection Act. The presence of this species in the Barak Valley of Assam is significant as it indicates that good mammalian diversity occurs in this region, inspite of rapid degradation of the forests.

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A look over on Red Sanders

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One among the rare and highly valuable timbers of the Indian sub-continent is the Red Sanders. Its natural distribution is confined to a very small region in India. Hyped as 'Pride of Eastern Ghats', its distribution is restricted to Kadapa, Chittoor, Kurnool, and Nellore districts in Andhra Pradesh and a few patches in Vellore, Tiruvannamalai, and Chengalpattu districts in Tamil Nadu.

For a deciduous tree growing up to 15m height, with round symmetrical crown architecture, the tree typically dominates other vegetation in the forest. The tree remains leafless for a short stint of time from February to March. This is followed by new green foliage and bright yellow coloured racemose inflorescence in the subsequent months. The pods mature before the next leaf fall. Red Sanders has a distinctive bark pattern concealing the bright red coloured wood beneath it. This deep red-coloured wood has a high price in the international market because of which this tree is felled rampantly, thus pushing this endemic species to the verge of extinction.

From the 2000s, the Red Sanders has captured the attention of the media due to the illicit felling and trade network, especially the social implications of its illegal trade network. It is believed that the locals and tribal men nearby the forest regions are involved in felling of the trees. Investigations, however, revealed that these men are the mere puppet of the wildlife trade kingpins. These mastermind kingpins engage the locals and tribal men as daily labour for felling the trees. Having little awareness nor any idea about the consequence of felling Red Sanders trees, these labours simply fell the trees to earn their daily wage.

There are dedicated forest staff to patrol and protect the Red Sanders and this has proven to reduce the felling

but there is a dark side. Many times, the fellers get shot down by the forest officials or the forest officials get injured or killed. There was a news story published in a leading English daily; The Hindu on 07 April 2015. It was about the 20 tree fellers from Tamil Nadu gunned

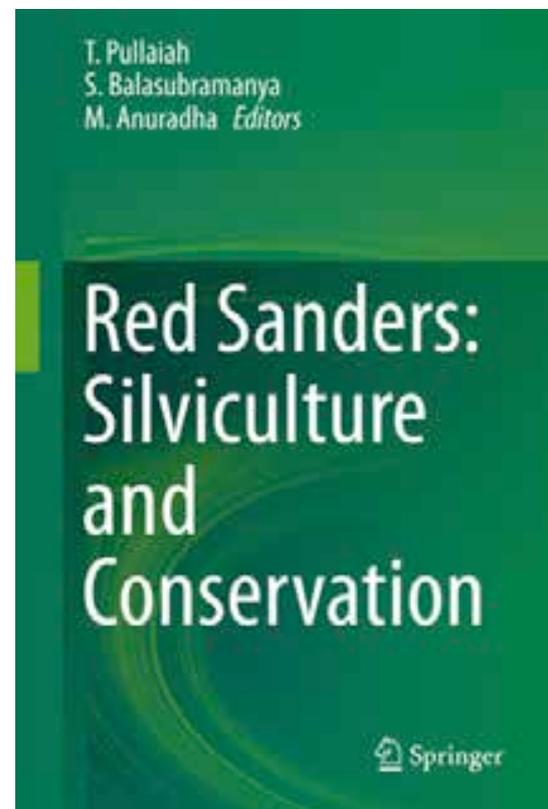
Red Sanders: Silviculture and Conservation

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by the Andhra Pradesh police department. Another story published in *The New Indian Express* (on 24 February 2018) reports that “Between January 2015 and December 2017, 10,558 Tamil Nadu residents were arrested by Andhra Pradesh police, forest department and Red Sanders Anti-Smuggling Task Force (RSASTF)”. This led to hatred, ethnicity controversy, insulting and even name-calling of the officials. Overall, it further added to the regional dispute existing between two states, Andhra Pradesh and Tamil Nadu.

Is it a controversial tree? It is one more question in my mind when I began my quest to understand this tree. As a forestry graduate, I first encountered this species in plantations raised in our college. It did not have the appeal, nor the enchantment because it was grown outside the natural geographical zone, however, it was my first encounter with this species. Before my graduation, I did a six-month project work on the characterization of Red Sanders. It was aimed at understanding the phenotypical difference across different provenances. And personally visited some regions in Tamil Nadu where the tree was occurring naturally. More I learned about the species, the more questions popped up in my mind. Some of the questions remain a mystery to me, even today!

So, when I stumbled upon a monograph on Red Sanders, published by Springer, I was instantly curious to open the book anticipating answers to my questions. The book under review pieced together by T. Pullaiah, S. Balasubramanya, and M. Anuradha is an important addition to the literature on Red Sanders. Three editors have meticulously worked to bring out this 15-chapter book. Apart from the editors, there are other prominent researchers and scientists who have also shared their work on Red Sanders in this book.

Now begins the story of the book, with a formal introduction chapter. It is intended for any beginner, giving exposure to Red Sanders as a tree – botanically. Chapter two is all about the taxonomy, one of the core areas of botanical science, and distribution of Red Sanders. The lead author of the first two chapters, Dr. T. Pullaiah’s experience is reflected here. I am sure, even if you have not seen this tree so far, you might become curious like me to know more. On finishing chapter three, the readers might be able to comprehend the reasons – Why this tree is so valued for? Short, explicit narration about the uses of Red Sanders’ wood and its parts are there in chapter three, along with beautiful photographs. Some typical wood characteristic features are also enlisted in chapter 3. Some properties described

here may not be familiar, still, they do have significance while identification of wood.

In my opinion, the chapters in this book can be grouped into four sections. The first three chapters form section I, aimed at introducing the species. Chapter four to six is more technical content on biochemistry and pharmacology of Red Sanders. This forms section II. Silviculture, the art and science of cultivating trees is the centre-hub in forestry science. Chapters seven to eleven deal with the silvicultural aspects of Red Sanders. This can be section III. The remaining chapters can be grouped as section IV dealing with a socio-economic and ecological aspect such as genetic diversity, threats, conservation and illegal trade networks. Except for chapter 15, titled as ‘Tree Improvement of Red Sanders’ can be placed after chapter 9, so that there is a coherence.

At the beginning of the article, I was narrating about the media highlighting the conflict due to Red Sanders tree felling. In Chapter 14, there is a long list of news stories published in different newspapers in India. This support my narration about the social significance of this species.

There is repetition of information and content which I believe is inevitable in an edited book, especially when it is about a single species. For instance, there is the redundancy of information on the natural distribution of Red Sanders in many chapters. This does affect the readability of the content. However, there is a specific need to mention the repetition of information even within the same chapter. In chapter three, there are two sub-sections with the same heading – ‘Fluorescence Analysis’. This is one example which indicates that the book needs to be brushed up.

If I need to recommend this book, I feel myself at conflict. For this is not a textbook, nor mere summarization of the information about Red Sanders but I feel it is a book written mainly for researchers. This monograph gives an idea about the research work carried out on Red Sanders and also points out the potential areas to work in future. More specifically, it implies understanding how certain species might be unique and need more focus.

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Due to paucity of space, the list of reviewers for 2017–2019 is available online.

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